



## MXenes as sustainable functional nanomaterials for photocatalytic degradation of dye pollutants: Performance, effect of process parameters, stability and re-useability evaluation – A critical review

Stephen Sunday Emmanuel<sup>a</sup>, Ademidun Adeola Adesibikan<sup>b</sup>, Shepherd M. Tichapondwa<sup>a</sup>, Manoj P. Rayaroth<sup>c</sup>, Francisco Bosca<sup>d</sup>, Maria Luisa Marin<sup>d</sup>, Bakhtiar Ali Samejo<sup>e</sup>, Grzegorz Boczkaj<sup>e,f,\*</sup>

<sup>a</sup> Department of Chemical Engineering, Sustainable Environmental and Water Utilisation Processes Division, University of Pretoria, Pretoria 0002, South Africa Private Bag X20, Hatfield 0028, South Africa

<sup>b</sup> Department of Industrial Chemistry, Faculty of Physical Sciences, University of Ilorin, P. M. B. 1515, Ilorin, Nigeria

<sup>c</sup> Department of Life Sciences, School of Science, GITAM (Deemed to be) University, Visakhapatnam, 530045, India

<sup>d</sup> Instituto de Tecnología Química, Universitat Politècnica de València-Consejo Superior de Investigaciones Científicas, Avda. de los Naranjos s/n, Valencia E-46022, Spain

<sup>e</sup> Department of Sanitary Engineering, Faculty of Civil and Environmental Engineering, Gdansk University of Technology, Gdansk, G. Narutowicza St. 11/12, Poland

<sup>f</sup> School of Civil, Environmental, and Architectural Engineering, College of Engineering, Korea University, 145 Anam-ro, Seongbuk-gu, Seoul 02841, Republic of Korea

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### ABSTRACT

MXene, an emerging cutting-edge functional material with unique properties like high electron conductivity, improved surface functionality, efficient charge separation, good optical properties, and lower Schottky barrier for photocatalytic applications, has gained a high interest of the scientific community in the recent decade. This work aimed to critically review the advances in the application of MXENE-based materials (MXBM) for photocatalytic degradation of various hazardous dye contaminants for water and wastewater treatment applications. Specifically, the effect of various technical operating parameters that are of high importance for the evaluation of MXBM's real applicational value was analyzed. Mechanisms of photocatalytic degradation in MXBM were rationally discussed. Usefulness of MXenes in photocatalytic materials include enhancement of photo-generated charge carrier separation, serving as stable support, limitation of photocatalyst size, improving reactant adsorption.  $Ti_3C_2$  ( $Ti_3C_2T_x$ ) is the most popular MXene for this application, while few attempts have been made also for  $Nb_2CT_x$ . Effective nanocomposites were synthesized with  $TiO_2$ , MOFs,  $g-C_3N_4$ , chitosan, cellulose, BiOBr, CdS. Other attempts included hybrid materials containing  $Ag_3PO_4$ , CuO,  $MoS_2$ ,  $NiCo_2S_4$ ,  $Nd-NiCo_2O_4$ , ZnO. Coating the MXenes with Ag and Pd nanoparticles (NPs) was another effective method. Several effective applications were developed for degradation of Congo Red (CR), Crystal violet (CV), Direct Red 31 (DR31), Malachite Green (MG), Methylene Blue (MB), Methyl Orange (MO), Orange G, Rhodamine B (RhB). Based on the performed review, it was found that for many applications, the photocatalytic degradation efficiency exceeds 90 % for 60–80 min treatment. Typically, the optimal dose of MXBM photocatalyst is in the range of 10–100 mg/L, and the optimal pH is in the range 2–7.5. The major reactive species in the photocatalytic reaction catalysed by MXBM were hydroxyl radicals ( $\cdot OH$ ) and superoxide radical anions ( $\cdot O_2^-$ ), which are accountable for the degradation of the dye contaminants via Advanced Oxidation Processes (AOPs). The reusability of the materials is an important advantage of MXBM, as it can be treated mostly employing pure water or with addition of ethanol and can be reused up to 3–8 cycles while still maintaining degradation efficiency of >75 %. In the end, the knowledge gap and problematic issues that can help to improve the MXBM, as well as provide future research directions for investigators, were highlighted. Future studies should focus, beside the degradation effectiveness, on intermediate products (by-products) formation, cost analysis, sustainable scale-up strategies under real-case scenario conditions. Importantly, a long-term toxicity assessments of MXBMs should be studied along with research on particle size-dependent toxicity of different types of photocatalytic MXene-based materials. Due to

\* Corresponding author at: Department of Sanitary Engineering, Faculty of Civil and Environmental Engineering, Gdansk University of Technology, Gdansk, G. Narutowicza St. 11/12, Poland.

E-mail addresses: [grzegorz.boczkaj@pg.edu.pl](mailto:grzegorz.boczkaj@pg.edu.pl), [grzegorz\\_boczkaj@korea.ac.kr](mailto:grzegorz_boczkaj@korea.ac.kr) (G. Boczkaj).

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utilisation of hydrated electrons in specific conditions MXBMs could find applicability also in Advanced Reduction Processes (ARPs). Photocatalyzed reactions could utilize sulphite, dithionate or gaseous hydrogen as reductants and sources of reductive radicals. This application of MXBMs seems to be overlooked by scientific community.

## 1. Introduction

Dyes are among the major contributors to environmental pollution, often discharged in untreated or inadequately treated form from industries such as pharmaceuticals, plastics, leather, food, and textiles into the water bodies, especially [1,2]. Therefore, sustainable solutions are needed to reduce dye concentrations to an acceptable level [3].

Notably, adsorption [4–6], coagulation [7], membrane filtration [8,9], bioremediation, chemical oxidation, and electrochemical methods [10,11] have been used as some of the approaches for the treatment of wastewater containing dyes. However, it is worth mentioning that the majority of these strategies (excluding electrochemical and biodegradation) move contaminants from one medium to another or generate extra sludge, which needs additional treatments and disposal [12]. Also, some of the adsorbents and catalysts employed for the aforementioned methods are linked to diverse drawbacks, such as low adsorption capacity, slow adsorption kinetics, a small number of reactive sites, low physical and chemical stability, and relatively low-moderate reusability [13].

To overcome these issues, a very effective, ecofriendly, and economically feasible process called advanced oxidation processes (AOPs) which is focused on the in-situ generation of highly oxidizing reactive oxygen species (especially,  $\bullet\text{OH}$  radicals) are used to degrade the toxic organic contaminants into less toxic molecules and finally to  $\text{CO}_2$ , water and inorganic ions (sulfate, nitrate, etc.) [14–16]. There are various AOPs, such as photolysis, Fenton process, photocatalysis, cavitation, and electrochemical methods, reported in the removal of dye waste from water matrices. The main advantages of these AOPs are that they don't produce any sludge, can be operated with very less or without toxic chemicals, and have a rapid and high mineralization efficiency. Among them, photocatalysis has advantages such as efficiency in natural reaction conditions, low operational cost, reusability, and eco-friendliness. Semiconductors like  $\text{TiO}_2$ ,  $\text{Fe}_2\text{O}_3$ ,  $\text{ZnO}$ ,  $\text{ZnS}$ , and  $\text{CdS}$  are widely used in photocatalytic applications. [1,17]. However, photocatalytic materials employed also have several limitations. The

recombination reaction of the electron-hole pair and the subsequent processes affect the degradation efficiency in many of the semiconductor-based photocatalysts. As a result, several modifications or photocatalysts of different morphologies have been developed for environmental remediation applications. Two-dimensional (2D) photocatalysts have emerged as important materials for water purification due to their unique physicochemical properties. Their planar structure allows for functionalization with various co-catalysts and surface atoms, providing numerous active sites for catalytic reactions [18,19].

MXenes are 2D materials produced by etching their ternary layered nitride or carbide precursors having a common formula of  $\text{M}_{n+1}\text{X}_n\text{T}_x$ , and are composed of stacked two-dimensional sheets, where M denotes an early transition metal, X is C and/or N, and T represents surface functional groups/terminators such as O, OH, and F, as color-coded in Fig. 1 [20,21].

These materials are very effective in photocatalytic degradation of pollutants owing to their numerous unique properties such as outstanding electron conductivity, exceptional crystal structure, large specific surface area (SSA), good conductivity, good hydrophilicity, high catalytic performance, efficient separation [22–24], excellent biocompatibility [25], good optical properties, and a lower Schottky barrier for photocatalytic applications [26]. Therefore, MXene-based materials are explored for various environmental applications, including wastewater treatment. Many studies have reported the removal of antibiotics and heavy metals using MXenes [27,28]. Meanwhile, some reviews have been published focusing on MXenes for the adsorption and photocatalytic degradation of organic pollutants [29–31]. For example, Ill-ahi's team [29] review study focused on the photocatalytic degradation efficiency of only metal oxide/semiconductor doped MXene for azo dyes, while Mathew et al. [30] work provides an overview of  $\text{Ti}_3\text{C}_2$  MXene fabrication and their performance, as well as mechanisms for dye photocatalytic degradation. Recently, another review study by Seling et al. also concentrated solely on the synthesis of  $\text{Ti}_3\text{C}_2\text{T}_x$  MXene-based hybrid photocatalysts and their efficiency for the degradation of organic dye and other water pollutants like pharmaceuticals. The study also

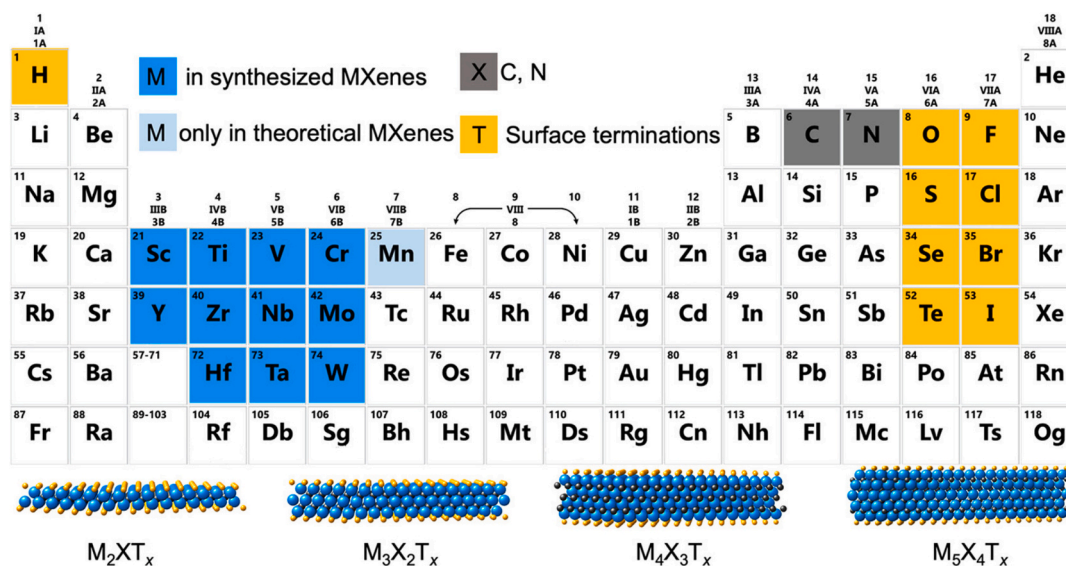


Fig. 1. Periodic table displaying the chemical makeup of MXenes. MXenes' building blocks are color-coded. At the bottom, four schematics of MXenes' quintessential frameworks are shown [20]. Reused with approval from the American Chemical Society (Copyright 2021).

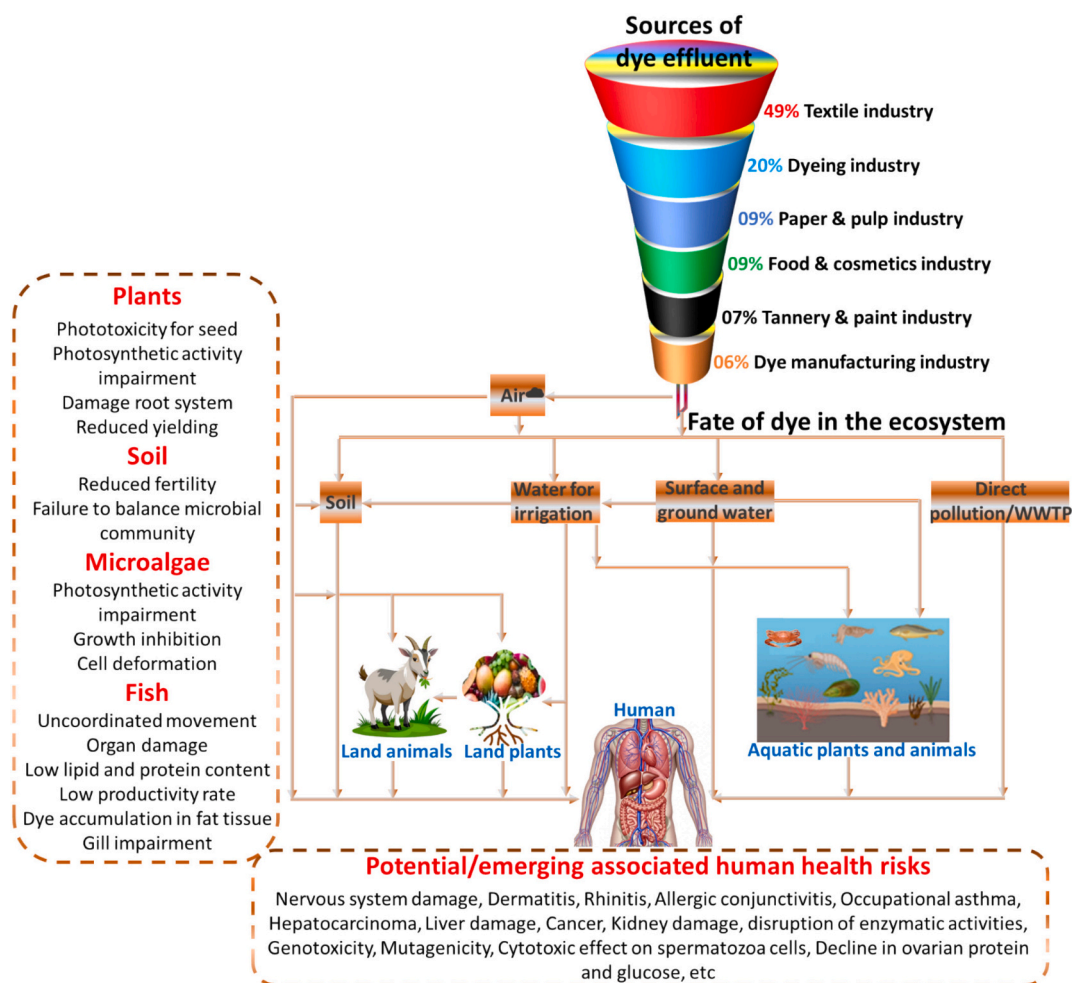


Fig. 2. Fate of dye pollutants within the ecosystem and associated ecotoxicities/health risks.

underscores computational methods for evaluating and improving the performance of  $Ti_3C_2T_x$  MXene-based hybrid photocatalysts [31].

However, for a thorough grasp of the MXenes photocatalytic degradation process for dye pollutants, these existing review papers lack in-depth elucidative analysis of the effect of various technical operating parameters on photocatalytic degradation performance, which is crucial for process and cost effectiveness optimization in any of the sustainable remediation approaches. Moreover, these reviews lack empirical analysis on the identification of reactive species in various MXene-based materials and do not critically evaluate the formation of dominant radical species facilitating dye degradation. A thorough investigation of the effect of co-existing ions in MXene materials has not been reviewed. The reusability studies were also not covered in the previous reviews.

Therefore, this review covers all the aspects of the studies related to dye remediation using MXene-based materials, including their economic benefits and real-world/pilot-scale application potential. This review furthermore discussed various challenges in the full implementation of MXenes-based materials for the photocatalytic degradation of dye pollutants, while identifying emerging future research directions. Therefore, the review will be beneficial for industries, engineers, investors, and various stakeholders to pragmatically establish MXenes' scalability, sustainability, fiscal feasibility, and industrial applicability, particularly in wastewater treatment applications.

## 2. Sources, fate, toxicities, and environmental impact of dyes

Dyes are colouring substances that are employed to impart color to or

polish substrates like clothes, polyethylene bags/plastics, cosmetics, papers, paints, drugs, food, etc. [32–34]. Thus, they have found applications in various industries that are very important to humanity. These industries serve as the major sources of dye effluent within the environment, with the textile industry leading the chart, followed by the dyeing industry, as presented in Fig. 2 [35,36]. Specifically, approximately 280,000 tons of dyes are wasted every year and released into the environment without proper treatment [3,37]. Another direct potential source of dyes in water environments could be aquaculture, where dyes like methylene blue, crystal violet, malachite green, etc., are used unlawfully to treat or manage parasitic, bacterial, or fungal infections in fish [38]. Household wastewater is another source of dye pollutants in the water because people throw away old or unused dye-based products, hair dyeing, use cosmetics or household chemicals that contain dyes, which are released into the sewer and then end up in a sewage treatment plant or are released directly into the environment [38,39]. These dye effluents mainly consist of aromatic and heterocyclic dyes with complex and stable structures and are often persistent in the environment owing to their thermal and photo-stable properties. Thus, when these effluent is emitted from various sources presented in Fig. 2, they migrate through different places of the ecosystem [38,40]. For example, they can reach surface water like rivers and lakes, which normally serve as a source of drinking water and irrigation water in drought zones or during the dry season [40,41]. Once dyes get to these surface water systems, they remain dissolved or adsorb onto suspended solids or degrade partially to form toxic aromatic amines and thus affect water quality, including color and taste. Furthermore, when dye-contaminated water is used for

irrigation, dye pollutants enter the soil, plants, and get to animals within the food chain [40]. Moreover, dyes can enter groundwater from surface water, through leaching from contaminated soils, or infiltration of dye wastewater from unlined disposal sites or landfills where dye-containing wastes are stored [38,42]. This makes dye-contaminated groundwater used for drinking, irrigation, or livestock a potential route of exposure to humans. Also, soil can be polluted by dyes through sludge application and airborne [36,38]. Specifically, even though dyes are not typically volatile, they can enter the air as fine/powder dust, aerosol droplets, particulates, and toxic vapors during manufacturing, spray applications, handling, textile printing, or incineration of dye-containing waste. The airborne dye pollutants then later settle onto land or aquatic systems through wet or dry deposition to continue their migration within the ecosystem. Notably, numerous toxicological and environmental impacts have been associated with the presence of dye pollutants in the ecosystem [36,38,42]. For instance, they can negatively affect soil chemistry by inhibiting soil microbial activity, altering nutrient cycling, and reducing soil fertility, which later affects plants and animals, including human beings [36,43]. In aquatic systems, the presence of dye pollutants can hinder the penetration of sunlight, which affects photosynthesis and therefore affects aquatic flora and fauna [36,44,45]. The presence of dyes in the water increases the chemical oxygen demand, biological demand, and changes the dissolved oxygen level of wastewater. In addition, the intake of dyes by plants can ultimately affect their growth as well as affect human health [36,37]. The dye contamination in water affects the growth and reproduction, development, and behavior of aquatic organisms. Ingestion is the major exposure pathway for humans. The biotransformation of dyes into reactive intermediates or toxic metabolites like aromatic amines was reported in the human body [3]. In particular, when humans are exposed to these reactive intermediates or toxic metabolites by eating plants and meat that have been impacted by dye pollutants within the food chain, or by occupational exposure, or through drinking, bathing, or swimming, it can cause many health-related issues. In particular, human exposure to dye pollutants has been linked to cancer, DNA damage, oxidative stress, nervous system damage, skin problems, asthma, liver damage, disruption of enzymatic activities, genotoxicity, and kidney damage, among others, as highlighted in Fig. 2 [3,36,46,47]. Reproductive issues such as cytotoxic effect on spermatozoa cells, decline in ovarian protein, and glucose have also been linked to dye pollutant exposure [37,48].

Understanding these fates is critical for mitigating the ecological and health concerns linked with dye pollutants, and this may require a combination of source reduction measures as well as the use of advanced remediation techniques. The photocatalytic degradation of dyes with MXenes-based materials, with their advantages, is underscored in the following sections.

### 3. Review methodology

A systematic literature search was performed on various reputable scientific databases such as Scopus, Google Scholar, Science Direct, and PubMed (up to January 2025) related to the theme of the proposed review, following the PRISMA protocol (Preferred Reporting Items for Systematic Reviews and Meta-Analyses) used by Bayode et al. [49] with slight modifications. MXene/metal carbonitrides / $Ti_3C_2T_x/Ti_3C_2$ , photocatalytic degradation, dye, electron trapping, effect of pH, effect of dose, effect of initial concentration, effect of ionic species, effect of time, effect of radical scavengers, regenerability, and recyclability/reusability were a few of the keywords. These were organized in different combinations using the Boolean “AND” and “OR” to refine the search results.

Numerous articles were collected and subsequently sorted and filtered based on various predetermined inclusion and exclusion criteria. Only peer-reviewed articles published in English were selected. The results of opinions, corrigenda, retracted articles, abstracts, encyclopedias, and seminars were not considered. In addition, the core analysis did not take into account experimental articles that did not examine

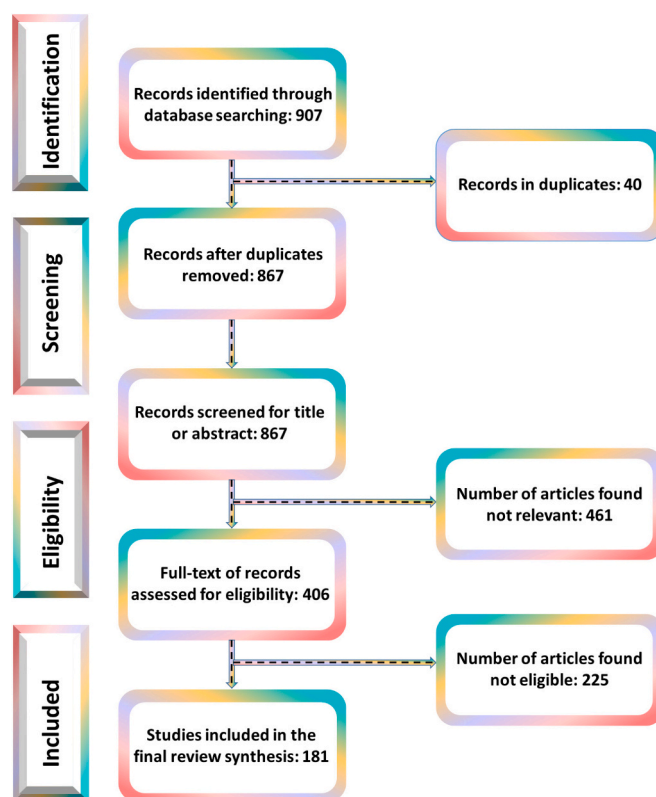


Fig. 3. Review methodology process.

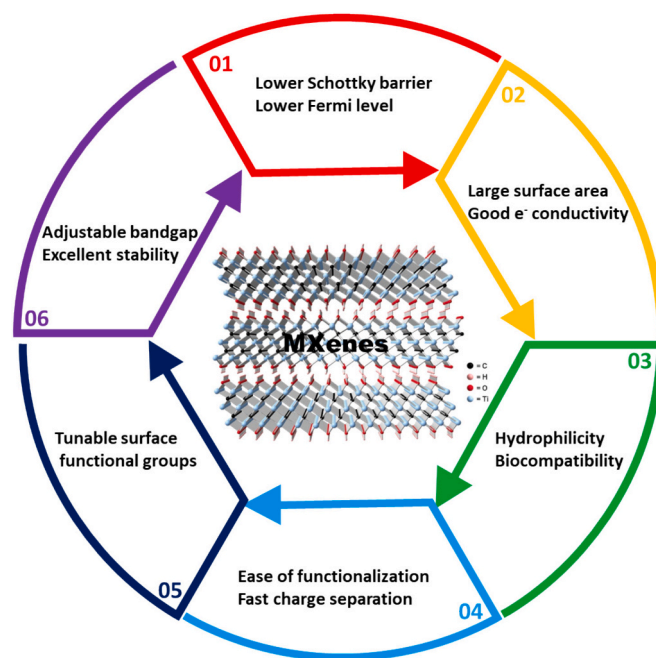


Fig. 4. Major unique properties of MXenes.

MXenes for the photocatalytic degradation of dye. Then, a collaborative article evaluation and discussion were done by the authors. After that, articles that satisfied the above criteria were deemed appropriate for the review work, and a total of 181 papers were finally used for the preparation of this review paper. A schematic overview of the review methodology is presented in Fig. 3.

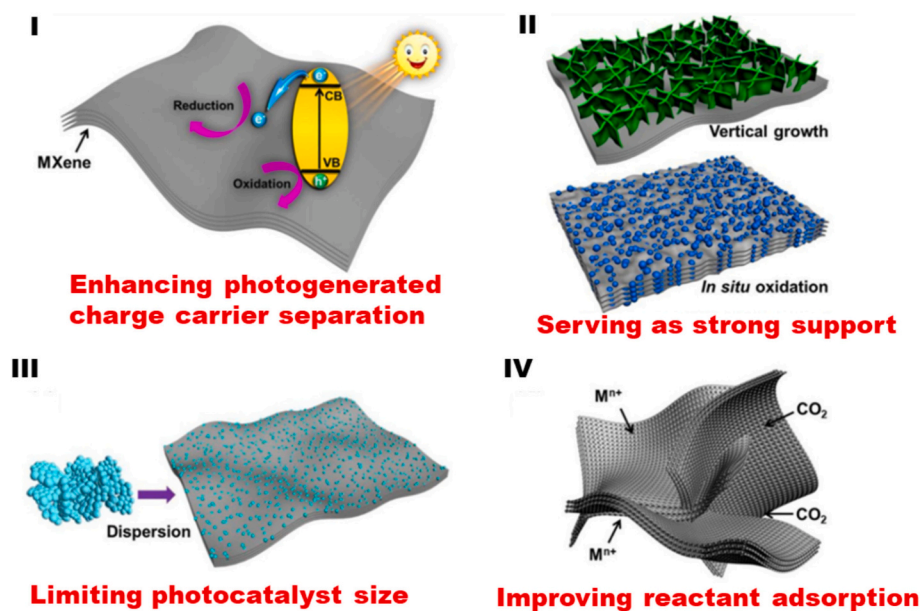


Fig. 5. Major advantageous functions of MXene in photocatalytic degradation [50]. Reprinted with approval from Elsevier (License code: 5612020177935).

#### 4. Advantages of MXene in chemical reactions

MXene, owing to its lower Fermi level, is often named as a very good material for photocatalysis [50]. However, considering MXenes' uniqueness (Fig. 4), such as the possibility to modify the surface functional groups and large SSA, the true functions of MXene in boosting the photocatalytic activity are far beyond the photo-generated electron ( $e^-$ ) acceptor [50]. Moreover, in general chemical reactions, high and tunable surface area and excellent chemical stability allow MXenes optimization for the degradation of specific chemical pollutants and ensure their reliability in diverse chemical environments while facilitating stable performance. The exclusive advantages of MXene are shown in Fig. 5 and briefly highlighted below.

- I. **Enhancing photo-generated charge carrier separation:** The recombination of  $e^-$  and holes ( $h^+$ ) in a photocatalyst can somewhat reduce the separation efficiency of charge carriers. As a result, one of the best methods for enhancing photocatalytic activity is to provide a conjunction of the photocatalyst with a co-catalyst for separating the photo-generated charge carriers [50,51]. Furthermore, MXenes show metallic conductivity characteristics, and compared to most n-type nanosemiconductors, their relatively low Fermi energy level is more positive, suggesting that they may be used as a co-catalyst for exchange and accumulation of photo-generated  $e^-$  [50,52]. More specifically, by utilizing MXenes as co-catalysts, a Schottky junction could be created to speed up the photo-generated charge carrier separation [50,53]. Likewise, due to their lower Fermi level during the photocatalytic reaction,  $e^-$  on the CB might move to  $Ti_3C_2$ . Additionally,  $H^+$  may be effectively reduced to create  $H_2$  due to  $Ti_3C_2$ 's strong electrical conductivity and mild  $H$  adsorption-free energy ( $\Delta G_H$ ). These electrons transported to MXene can also be used in other photocatalytic processes to convert  $CO_2$  into hydrocarbons (including methane) or formaldehyde, reduce sorbed nitrogen to  $NH_3$ , and form superoxide radical anions ( $\bullet O_2^-$ ) and hydroxyl radicals ( $\bullet OH$ ) for the degradation of organic pollutants [54,55].
- II. **Serving as strong support:** Given MXene's exceptional physico-chemical durability, it can be employed as a good support for other nanophotocatalysts [51,56] such as  $TiO_2$  [57,58],  $ZnO$  [59,60], and  $MoS_2$  [61] to prevent agglomeration and enhance its

performance [50,62]. It is also important to note that MXene can act not only as a strong support [63,64] but also as a starting material for the in-situ development of associated metal oxide photocatalysts [65,66].

- III. **Limiting photocatalyst size:** The dimensions of the MXene-based materials (MXBM) photocatalyst are important for photocatalytic operation; nevertheless, owing to their surface energy, they tend to aggregate into larger sizes, which reduces their contact area with pollutants [50]. Due to the MXBM's numerous surface functional groups, which are often formed using the wet chemical etching process, they may afford a better growth platform for the regulation of the morphological dimensions of the photocatalyst [67,68]. More specifically, the metal cations could be adsorbed on the negatively charged surface, subsequently enabling the uniform growth of photocatalysts [69]. As a result, the MXBM photocatalyst could reach a high SSA to afford a large number of surface-active sites to facilitate photocatalytic activity [50,51].
- IV. **Improving reactant adsorption:** Another crucial stage in the modification of the MXBM is the enhanced reactant adsorption. Owing to the abundant surface functional groups on MXene, it can also enhance the photocatalyst's capacity to adsorb pollutants [70,71]. For instance, due to the negatively charged surface, MXBM can effectively sorb the positively charged molecules [72]. Besides, the hybrid photocatalyst's porous structure is advantageous for adsorption [50]. Specifically, numerous hydrophilic terminal groups ( $-OH$ ,  $-O$ , and  $-F$ ) on the  $Ti_3C_2$  surface can improve its interaction with other nanosemiconductors as well as with  $H_2O$  molecules or target pollutants, both of which are necessary for the migration of  $e^-$  inside the hybrids and the chemical reaction in aqueous solution [73–75]. These unique properties and advantages play an important role in MXene degradation of dye pollutants.

#### 5. Overview of the photocatalytic degradation mechanism for MXBM

When photocatalysts are exposed to light, photon energy higher than the bandgap energy is absorbed from the light source to generate  $e^-$ , and the photo-generated  $e^-$  are captured and promoted to the upper band called the conduction band (CB), creating  $h^+$  in the lower band called the valence band (VB) [76]. The active functional group present on the

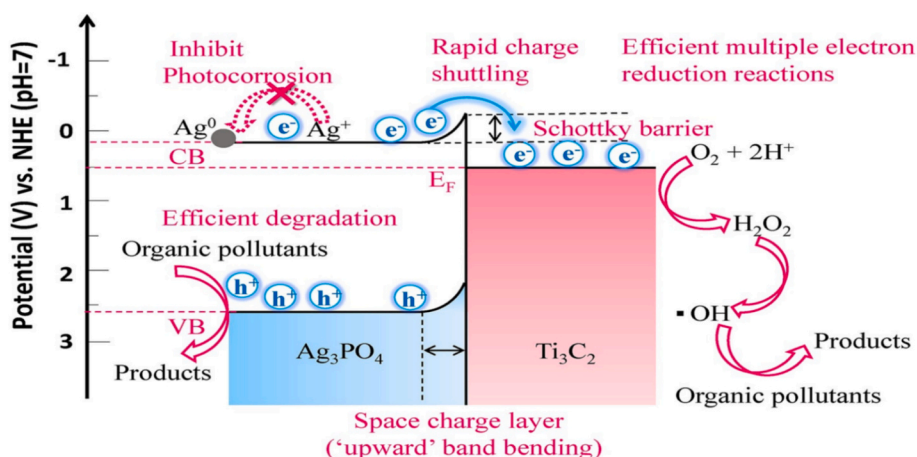


Fig. 6. Plausible MXBM photocatalytic degradation mechanism of dye pollutants using  $\text{Ag}_3\text{PO}_4/\text{Ti}_3\text{C}_2$  [80]. Reused with approval from Elsevier (Copyright 2018).

surface of MXene can help to separate the electron-hole pair. Theoretically, hydroxyl radicals ( $\cdot\text{OH}$ ),  $\text{h}^+$  and superoxide radicals ( $\cdot\text{O}_2^-$ ) are the three reactive species (RS) that account for the effective photocatalytic reactions [76,77]. In most of the photocatalysts containing MXenes, a transfer of charge from the photocatalyst to MXenes boosts the  $\text{e}^-/\text{h}^+$  pair separation and decreases its recombination, increasing the photocatalytic activity. Thus, photo-generated electrons are effectively utilized in the MXenes due to their outstanding electron-capture capabilities and efficient electron-hole separation. It has been suggested that MXene-based photocatalysts participate in photocatalytic activity by reacting with  $\text{H}_2\text{O}$  and  $\text{O}_2$  to produce  $\cdot\text{OH}$ ,  $\cdot\text{O}_2^-$ , and  $\text{h}^+$  [76,78], which are effective in several degradation reactions with organic pollutants, including attack on the aromatic ring, azo bond including the  $\text{N}=\text{N}$  bond, and  $\text{C}=\text{N}$  bond of pollutant molecules to degrade them into harmless molecules like  $\text{CO}_2$  and  $\text{H}_2\text{O}$  [79] as shown in Fig. 6. The Schottky barrier prevents the photo-charged electrons from returning from MXene to semiconductors. In the MXene, the photocatalytic degradation of water pollutants like dyes was enhanced by the Schottky connection as well as the excellent Fermi energy level of MXBM compared to the bare semiconductor photocatalytic materials, and this can be seen in their significantly improved degradation effectiveness discussed in the next section [76].

## 6. Performance evaluation of MXene-based material for dye photodegradation

This section discusses the degradation efficiency of various MXenes for the degradation of different dye pollutants. The treatment time and energy input are the most vital parameters that influence the process and thus control the degradation efficiency. These parameters were included in the comparison presented in Table 1. For example, the photocatalytic efficiency of the  $\text{Ti}_3\text{C}_2/\text{TiO}_2$  nanocomposites was explored for the degradation of rhodamine blue under visible light illumination. In comparison to pure  $\text{TiO}_2$  and commercial  $\text{TiO}_2$ , the composite of  $\text{TiO}_2$  microspheres and multilayer  $\text{Ti}_3\text{C}_2$  has comparatively high SSA, substantial photocurrent, and considerably less impedance. This material has offered a 94 % degradation in 60 min, which outperformed pure  $\text{TiO}_2$  by a factor of two. The  $\text{Ti}_3\text{C}_2$  addition successfully inhibited electron-hole pair recombination. The photo-generated  $\text{e}^-$  in the MXBM moved from  $\text{TiO}_2$  to  $\text{Ti}_3\text{C}_2$ . It was successfully blocked by the Schottky barrier at the  $\text{TiO}_2\text{-Ti}_3\text{C}_2$  [81]. A high rhodamine B (RhB) photocatalytic degradation performance (97 %) was also observed for  $\text{TiO}_2@/\text{Ti}_3\text{C}_2$  in 40 min. Comparing the effectiveness of the  $\text{TiO}_2@/\text{Ti}_3\text{C}_2$  heterojunction to that of pristine and commercial  $\text{TiO}_2$ , with respect to RhB degradation, a significant improvement was confirmed for the hybrid material [82]. In another study [83],  $\text{Ti}_3\text{C}_2$  MXenes were synthesized by

employing several HF-based etching approaches and were then treated by alkalization to substitute O-containing terminal functional groups for  $-\text{F}$  to give multilayer  $\text{Ti}_3\text{C}_2(\text{OH})_2$ . The performance of the multilayer  $\text{Ti}_3\text{C}_2(\text{OH})_2$  was assessed for photocatalytic degradation of methylene blue (MB), and about 81.2 % was degraded in 2 h under a 300-W Xe lamp [83]. In another work, Othman's group [84] investigated the degradation of RhB and MB utilizing  $\text{Ti}_3\text{C}_2\text{T}_x$  MXene coated with Ag and Pd NPs. The authors studied the catalytic degradation properties of both pristine and fabricated MXBM under sunlight and UV irradiation. The AgNPs/ $\text{TiO}_2/\text{Ti}_3\text{C}_2\text{T}_x$  demonstrated the highest degradation efficiency. The surface plasmon resonance effect for Ag played an important role in enhancing the absorption in the visible range of light to boost light harvesting and improve the charge separation step [84]. Improved efficiencies of such hybrids were also reported in other similar studies [85]. Another corresponding study reported the degradation of malachite green (MG) dye using  $\text{NiCo}_2\text{O}_4$ , doped  $\text{Nd-NiCo}_2\text{O}_4$ , and MXBM  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  [86]. It was found that the  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  nanocomposite showed the highest degrading activity of 98 % in 180 min compared to doped (61 %) and pure (52 %) forms. The synergy between Nd-doping and composite creation, which produces improved visible light absorption and effective  $\text{e}^-/\text{h}^+$  decoupling, accounts for the  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  increased photocatalytic performance [86]. In another study, a  $\text{Ti}_3\text{C}_2\text{T}_x$  nanosheet decorated with a metal organic framework (MOF) containing zinc for photocatalytic degradation of MB and DR31 was investigated. As presented in Table 1, for both the MB and Direct Red 31 (DR31) dyes, the MXene/MOF photocatalyst shows a degradation performance of 62 % and 35 %, respectively. Notably, the MXene/MOF was reported to exhibit better photocatalytic performance than the bare MOF and MXene. Furthermore, increased effectiveness takes place in the presence of  $\text{H}_2\text{O}_2$  as it allows for maximizing the radical generation. Specifically, employing MXene/MOF, 74 % of MB and 51.6 % of DR31 dyes were photodegraded in the presence of  $\text{H}_2\text{O}_2$  after 80 min. The authors stated that the excellent porosity of MXene/MOF nanocatalyst with an adjustable bandgap energy and high SSA facilitated fast charge transfer and prohibited the photo-generated  $\text{e}^-$  and  $\text{h}^+$  recombination [87]. In another photocatalytic study,  $\text{Ti}_3\text{C}_2/\text{BiFeO}_3$  was employed to degrade Congo Red (CR) under visible light irradiation [88]. In just 42 min, 100 % CR was degraded. The authors pointed out that two variables contribute to CR's rapid and efficient degradation. Firstly, a greater SSA ( $147 \text{ m}^2/\text{g}$ ) and secondly, a decreased rate of  $\text{e}^-/\text{h}^+$  recombination. Additionally, the control experiment performed under dark conditions only offered 21 % of the degradation, suggesting that  $\text{Ti}_3\text{C}_2/\text{BiFeO}_3$  functions better as a photocatalyst, while exhibiting affinity (adsorption) to the pollutant. In parallel, an excellent photocatalytic degradation efficiency of 2D  $\text{Ti}_3\text{C}_2\text{T}_x$  was reported by Bury [89]. Their work clarifies, for the first time, how the photocatalytic

**Table 1**  
Summary of MXenes-based material performance for photocatalytic degradation of dyes.

MXenes-based photocatalyst (MXBM)	Dye	Light input source	Dose (mg/L)	% Degradation	Time (Min.)	Ref.
ZIF-8@Ti <sub>3</sub> C <sub>2</sub> MXene	Methylene blue	Xe lamp 160 W	400	95	120	[90]
Ag/Ag <sub>3</sub> PO <sub>4</sub> /MXene	Crystal violet	Xe lamp 300 W		84	90	[91]
ZnO/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	Xe lamp	500	97	45	[92]
Sm <sub>2</sub> O <sub>3</sub> /MXene/g-C <sub>3</sub> N <sub>4</sub>	Orange G	Tungsten halogen lamp 500 W		~100	300	[93]
MXene@Cu/Tb/MgFe <sub>2</sub> O <sub>4</sub>	Crystal violet	Tungsten bulb	429	95	120	[94]
MXene@carbon quantum dots	Malachite green	Halogen lamp 35 W		96	25	[95]
NiMoO <sub>4</sub> /MXene	Methylene blue	Visible light	200	97	8	[96]
MXene@V-CuO	Malachite green	Bulb 200 W	429	88	100	[97]
g-C <sub>3</sub> N <sub>4</sub> /MXene	Methylene blue	Halogen lamp 500 W		~99	150	[98]
MXene@g-C <sub>3</sub> N <sub>4</sub> /TiO <sub>2</sub>	Rhodamine B	Xe lamp 300 W		99	120	[99]
MXene@Cu/Tb/MgFe <sub>2</sub> O <sub>4</sub>	Congo red	Xe lamp	429	97	40	[94]
MXene/CuO/B-g-C <sub>3</sub> N <sub>4</sub>	Malachite green	Sunlight		98	120	[100]
MnWO <sub>4</sub> /MXene/carbon xerogel	Crystal violet	Xe lamp 500 W		95	60	[101]
MXene/ZnO-tetrapod	Rhodamine B	Sunlight		78	120	[102]
MXene/ZnO-tetrapod	Methyl orange	Sunlight		77	120	[102]
ZnO/NiWO <sub>4</sub> /V <sub>2</sub> C MXene	Rhodamine B	Xe lamp		92	120	[103]
ZnO/NiWO <sub>4</sub> /V <sub>2</sub> C MXene	Methylene blue	Xe lamp		93	120	[103]
Ti <sub>3</sub> C <sub>2</sub> /TiO <sub>2</sub>	Rhodamine B	300 W Xe lamp	20	94	60	[81]
ML-Ti <sub>3</sub> C <sub>2</sub> (OH) <sub>2</sub>	Methylene blue	300-W Xe lamp	6	83	120	[83]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	300 W Xe lamp	30	97	40	[82]
MXene/g-C <sub>3</sub> N <sub>4</sub>	Methylene blue	500 W halogen lamp	500	~70	180	[104]
Bi <sub>1-x</sub> La <sub>x</sub> Fe <sub>1-y</sub> Mn <sub>y</sub> O <sub>3</sub> /Ti <sub>3</sub> C <sub>2</sub>	Congo red	300 W xenon lamp and 5 W light-emitting diode	100	~100	30	[105]
Fe <sub>2</sub> O <sub>3</sub> /ZnFe <sub>2</sub> O <sub>4</sub> @Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	300 W Xe lamp	20	97	150	[106]
AgNPs/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	400 W Hg lamp		99	30	[84]
Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Bromphenol blue	Solar	35	~95	40	[107]
AgNPs/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	400 W Hg lamp		99	40	[84]
TiO <sub>2</sub> /MXene	Methylene blue	500 W mercury lamp	10	99	60	[85]
TiO <sub>2</sub> /MXene	Methyl orange	500 W Mercury lamp	50	91	75	[108]
Ag/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Safranin	Visible	5	99	15	[109]
MXene/Zn-MOF	Direct red 31	250 W mercury vapor lamp + H <sub>2</sub> O <sub>2</sub>	30	52	80	[87]
AgNPs/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	Solar		96	120	[84]
TMAOH-MXene	Bromocresol green	128 W UV	10	100	180	[89]
Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Rhodamine B	Solar	35	~95	40	[107]
MXene/Zn-MOF	Methylene blue	250 W mercury vapor lamp + H <sub>2</sub> O <sub>2</sub>	30	70	80	[87]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub>	Methyl orange	UV	10	97	50	[110]
AgNPs/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	Solar light		96	80	[84]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	500 W Xenon lamp		92	50	[111]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /NH <sub>2</sub> -MIL-88B(Fe)	Methylene blue	300 W Xenon lamp	10	83	120	[112]
Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Methyl orange	Solar	35	99	40	[107]
Ti <sub>3</sub> C <sub>2</sub> /BiFeO <sub>3</sub>	Congo red	300 W xenon lamp	100	100	42	[88]
TMAOH-MXene	Methylene blue	128 W UV	10	100	180	[89]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /NH <sub>2</sub> -MIL-88B(Fe)	Methyl orange	300 W Xenon lamp	10	44	120	[112]
ZnO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	40-W UV lamp	50	86	60	[113]
ZnO@Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	300 W Xe arc lamp	20	95	180	[114]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub>	Methyl orange	175 W mercury lamp		98	30	[115]
TiO <sub>2</sub> @Ti <sub>3</sub> C <sub>2</sub> /g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	300-W Xe lamp	100	100	60	[116]
Ti <sub>3</sub> C <sub>2</sub> /TiO <sub>2</sub> /CuO	Methyl orange	175W mercury lamp	100	99	80	[117]
Bi <sub>2</sub> WO <sub>6</sub> /Nb <sub>2</sub> CT <sub>x</sub>	Rhodamine B	500 W Xenon lamp	50	100	90	[118]
Ti <sub>3</sub> C <sub>2</sub> /MoS <sub>2</sub>	Methyl orange	400W metal halide lamp	50	97	30	[119]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	300 W visible		98	60	[120]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /NH <sub>2</sub> -MIL-88B(Fe)	Rhodamine B	300 W Xenon lamp	10	97	30	[112]
ZnO@Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Congo red	300 W Xenon arc lamp		89	270	[121]
Bi <sub>2</sub> WO <sub>6</sub> /Nb <sub>2</sub> CT <sub>x</sub>	Methylene blue	500 W Xenon lamp	50	93	90	[118]
NiMnO <sub>3</sub> / NiMn <sub>2</sub> O <sub>4</sub> - Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	LED lamp		100	180	[122]
Ti <sub>3</sub> C <sub>2</sub> /ZnO	Rhodamine B	UV	100	98	18	[59]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Methylene blue	300 W UV lamp		100	55	[78]
Ti <sub>2</sub> C/ZnCo <sub>2</sub> O <sub>4</sub>	Congo red	Visible	50	78	270	[123]
Ti <sub>3</sub> C <sub>2</sub> /g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	270W xenon lamp	20	98	60	[124]
Bi <sub>0.90</sub> Gd <sub>0.10</sub> Fe <sub>0.80</sub> Sn <sub>0.20</sub> O <sub>3</sub> /MXene	Congo red	300 W Xenon lamp	100	100	120	[125]
ZnO@Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	300 W Xenon arc lamp		83	270	[121]
NiMnO <sub>3</sub> / NiMn <sub>2</sub> O <sub>4</sub> - Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	LED lamp		100	20	[122]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Methyl orange	300 W UV lamp		99	360	[78]
NiMnO <sub>3</sub> / NiMn <sub>2</sub> O <sub>4</sub> - Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	LED lamp		100	180	[122]
Mn <sub>2</sub> O <sub>3</sub> -Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	250 W LED lamp	20	93	35	[126]
ZnO@Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	300 W Xenon arc lamp		92	270	[121]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Rhodamine B	300 W UV lamp		99	20	[78]

(continued on next page)

Table 1 (continued)

MXenes-based photocatalyst (MXBM)	Dye	Light input source	Dose (mg/L)	% Degradation	Time (Min.)	Ref.
MXene/ZnS/chitosan-cellulose	Acid blue 80	1000 W deuterium lamp	1	>70	–	[127]
Nb <sub>2</sub> C/CdS	Rhodamine B	300 W Xe lamp	25	99	20	[128]
Nd-NiCo <sub>2</sub> O <sub>4</sub> /MXene	Malachite green	solar	40	98	180	[86]
ZnO/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	300 W xenon lamp	30	100	50	[129]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Orange G	250 W Xe lamp	100	89	60	[130]
ZnO/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	300 W xenon lamp	30	100	70	[129]
Mn <sub>2</sub> O <sub>3</sub> -Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	250 W LED lamp	20	100	25	[126]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	250 W Xe lamp	100	~99	40	[130]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Methylene blue	250 W Xe lamp	100	99	60	[130]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Methyl orange	250 W Xe lamp	100	60	60	[130]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Malachite green	250 W Xe lamp	100	85	60	[130]
CuO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	Sunlight	50	100	80	[131]
NiCo <sub>2</sub> S <sub>4</sub> /MXene	Rhodamine B	250 W Xenon lamp	5	~100	120	[132]
ZnO/V <sub>2</sub> C	Methylene blue	500 W Xe lamp	50	100	120	[60]
BiOBr/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	300 W Xe lamp	20	100	30	[133]
Ag <sub>3</sub> PO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Methyl orange	300 W Xe lamp	50	98	120	[134]
CuFe <sub>2</sub> O <sub>4</sub> /MXene	Methylene blue	Visible + H <sub>2</sub> O <sub>2</sub>	50	94	40	[135]
Ag <sub>3</sub> PO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Malachite green	300 W Xe lamp	50	99	120	[134]
g-C <sub>3</sub> N <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	500 W metal halide lamp	40	100	25	[136]
NiFe <sub>2</sub> O <sub>4</sub> /MXene	Methylene blue	UV visible		74	70	[137]
InSe/TiO <sub>2</sub> /MXene	Methyl orange	300 W Xe lamp	60	92	60	[138]
ZnO/Nb <sub>2</sub> C	Methylene blue	500 W Xenon lamp	50	63	120	[60]
Ti <sub>3</sub> C <sub>2</sub> /NiTiO <sub>3</sub>	Methylene blue	Sunlight	5	97	120	[139]
Ag@ZnO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	300 W Xenon lamp	50	91	50	[69]
Ag <sub>3</sub> PO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Rhodamine B	300 W Xenon lamp	50	94	120	[134]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	500 W mercury lamp	50	98	30	[140]
BPQDs/Ti <sub>3</sub> C <sub>2</sub> @TiO <sub>2</sub>	Methyl orange	400 W metal halide lamp	50	93	60	[141]
I-Ti <sub>3</sub> C <sub>2</sub> @graphitic carbon-supported TiO <sub>2</sub>	Methylene blue	500 W Hg lamps	50	99	60	[142]
CdS@Ti <sub>3</sub> C <sub>2</sub> @TiO <sub>2</sub>	Rhodamine B	Visible	50	100	150	[57]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Bi <sub>4</sub> Ti <sub>3</sub> O <sub>12</sub>	Methyl orange	500 W Xe/Hg arc lamp	35	100	60	[143]
Ag <sub>3</sub> PO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Methylene blue	300 W Xenon lamp	50	91	120	[134]
BiOI/TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	500 W Xe lamp	50	100	80	[144]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	500 W mercury lamp	50	80	30	[140]
CdS@Ti <sub>3</sub> C <sub>2</sub> @TiO <sub>2</sub>	Methylene blue	Visible	50	100	150	[57]
TiO <sub>2</sub> @Ti <sub>3</sub> C <sub>2</sub> /g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	300 W Xe lamp	100	100	60	[116]
BiVO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub>	Methyl orange	300 W Xenon lamp	100	99	130	[145]
WO <sub>3</sub> /MXene	Methylene blue	Sunlight	5	99	80	[146]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Bi <sub>4</sub> Ti <sub>3</sub> O <sub>12</sub>	Rhodamine B	500 W Xe/Hg arc lamp	35	100	50	[143]
Ti <sub>3</sub> C <sub>2</sub> /NiTiO <sub>3</sub>	Malachite green	Sunlight	5	98	120	[139]
Ti <sub>3</sub> C <sub>2</sub> /TiO <sub>2</sub> /g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	300 W Xenon lamp	60	98	50	[147]
Ti <sub>3</sub> C <sub>2</sub> /NiTiO <sub>3</sub>	Methyl orange	Sunlight	5	87	120	[139]
BiVO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub>	Congo red	300 W Xenon lamp	100	100	60	[145]
α-Fe <sub>2</sub> O <sub>3</sub> @SnO <sub>2</sub> /MXene	Rhodamine B	300 W Xe lamp		72	140	[148]
ZnO/Ti <sub>2</sub> C	Methylene blue	500 W Xe lamp	50	99	120	[149]
MXene/PbCrO <sub>4</sub>	Methylene blue	Sunlight	20	93	120	[150]
Ti <sub>3</sub> C <sub>2</sub> (OH) <sub>2</sub>	Methylene blue	150 W Xe Arc Lamp	20	97	80	[151]

characteristics of model 2D Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> are affected by oxidation stabilization by enhancing efficiency. Two well-established methods are used to fabricate the delaminated MXene: minimum intensive layer delamination with hydrochloric acid/lithium fluoride (MILD-MXene) and hydrofluoric acid/tetramethylammonium hydroxide (TMAOH-MXene). L-ascorbic acid is then used to stabilize the product. In the 180 min under 128 W UV, 100 % degradation of methylene blue and bromocresol green was achieved. In addition, the process of industrial textile dye effluents treatment at 100 times higher concentration compared to the model dye solution must also ensure degradation effectiveness to attain industrial viability. These studies revealed that MILD-MXene exhibits superior efficiency over TMAOH-MXene because of its narrower optical band gap. To completely degrade the dye, the MILD-MXene only needed a few seconds of exposure to UV light. In another report, at around one hour, a 90 % MB degradation by CuO/Ti<sub>3</sub>C<sub>2</sub> was achieved under sunlight irradiation. The enhanced rate of MB degradation by the MXBM can be attributed to the efficient intercalation of NPs between the Ti<sub>3</sub>C<sub>2</sub> sheets, which enhances the photocatalyst's capacity to capture electrons.

Also, Liu et al. reported that a ZnO@Ti<sub>3</sub>C<sub>2</sub> composite increases MB's photocatalytic degradation capacity and yields 94.84 % degradation in 180 min. The authors highlighted that the ZnO@Ti<sub>3</sub>C<sub>2</sub> MXBM's photocatalytic activity is significantly enhanced by the Schottky junction, which inhibits photo-generated e<sup>-</sup> from returning to ZnO and encourages their separation from h<sup>+</sup> [114]. In another research, using a

hydrothermal process, a unique 2D Ti<sub>3</sub>C<sub>2</sub>/MoS<sub>2</sub> nanosheets were effectively fabricated by Yao and co-workers [119]. It was found that the Ti<sub>3</sub>C<sub>2</sub>/MoS<sub>2</sub> has shown methyl orange (MO) degradation of 97.4 % in 30 min under irradiation by a 400 W metal halide lamp. It is attributed to its increased SSA and optical absorption capacity. Furthermore, Ti<sub>3</sub>C<sub>2</sub> in conjunction with sheet-like MoS<sub>2</sub> NPs lowers the impedance, which brings an improved stimulated semiconductor e<sup>-</sup> transfer, inhibition of e<sup>-</sup>/h<sup>+</sup> recombination, and then enhances photocatalytic degradation performance.

A summary of the research on the photocatalytic properties of MXenes for dye degradation is provided in Table 1. In most cases, the dye degradation efficiencies reached close to 100 %. This finding is noteworthy because it shows the possibility of using MXBM for pollutant elimination. However, it is important to highlight that most studies involving MXBM did not report the formation of intermediate products, the reaction mechanism, or the extent of mineralization. The fact of target pollutant removal confirmed by the analytical method does not ensure effective degradation of the formed by-products. Moreover, although the •OH generation is suggested in most of the research articles to justify the photocatalytic degradation of dyes, the intermediate product formation and the mechanism were not clearly demonstrated. In addition, the dependency of degradation efficiency on operating conditions is critically discussed in the following sections to properly elucidate the potential of MXenes for dye photocatalytic degradation in

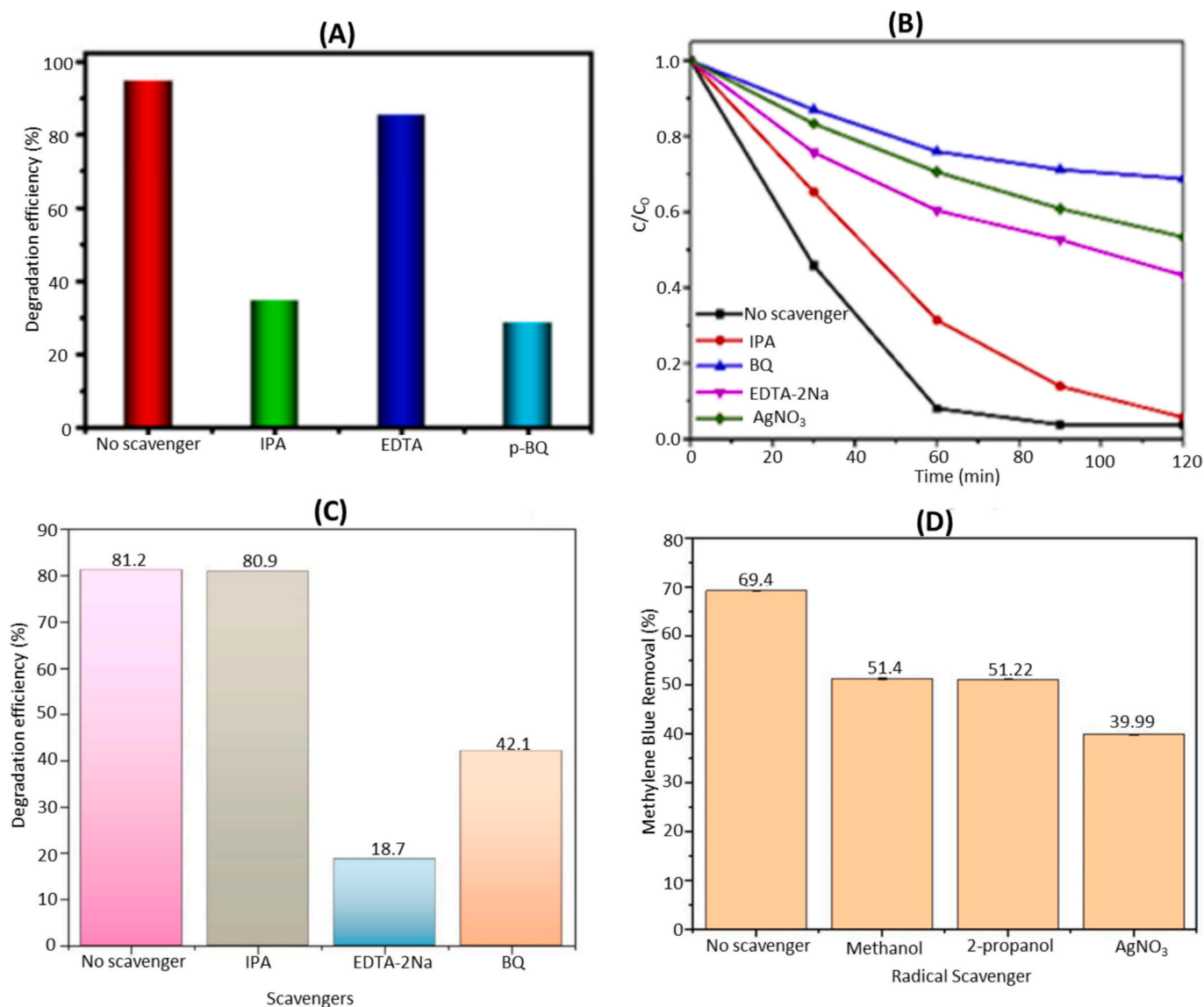


Fig. 7. Effect of radical scavenging on photocatalytic degradation of (a) RhB using  $\text{TiO}_2@Ti_3C_2$  [82]. Reused with approval from Elsevier (Copyright 2020), (b) RhB using  $\text{Fe}_2\text{O}_3/\text{ZnFe}_2\text{O}_4@Ti_3C_2$  [106]. Reused with approval from Elsevier (Copyright 2019), (c) MB using  $\text{ML-Ti}_3\text{C}_2(\text{OH})_2$  [83]. Reused with approval from Elsevier (Copyright 2022), (d) MB using  $\text{MXene/g-C}_3\text{N}_4$  [104]. Reused with open access approval from MDPI (Copyright 2022).

industrial conditions.

## 7. Identification of dominant radical species

As given in the previous section, the major reactive species that could be evolved during dye photocatalytic degradation using MXBM as photocatalyst are  $\cdot\text{OH}$ ,  $\cdot\text{O}_2^-$ , and  $\text{h}^+$ . In this context, scavenger tests could help to identify the reactive species [152–154]. The scavenger interferes with the pollutants' reactions with ROS, and some scavengers affect the catalytic activities [152]. Therefore, various studies have investigated the impact of radical scavengers on the photocatalytic degradation of dye pollutants to identify the contribution of ROS in MXBM. The major scavengers used in photocatalysis are methanol (MeOH) and isopropyl alcohol (IPA) (scavengers of  $\cdot\text{OH}$ ), p-benzoquinone (p-BQ) (as a scavenger of  $\cdot\text{O}_2^-$ , triethanolamine (EDTA), (scavenger of  $\text{h}^+$ ), and Furfuryl alcohol (FFA, scavenger for  $^1\text{O}_2$ ). However, in general, scavengers are not specific for any reactive intermediate, as several RSs can react with the same scavenger with a bit different rate constant. Clarifying the relationship between quencher-induced inhibition and the contribution of reactive species is significant. The

effectiveness of the quencher method is indisputable, and rather than simply disproving it, more research into its adequate use is needed [155]. Therefore, it is needed to take into account all their bimolecular rate constant values with all the reactive intermediates.

Numerous studies were carried out to identify the primary RSs that are responsible for the photocatalytic process and provide data about the RhB photodegradation mechanism in  $\text{TiO}_2@Ti_3C_2$ . Three different types of trapping agents were utilized in photocatalytic degradation: p-BQ, EDTA, and IPA. The addition of EDTA does not appear to have any discernible effect on RhB degradation using  $\text{TiO}_2@Ti_3C_2$  photocatalyst, as seen in Fig. 7a. Thereby, it was suggested that  $\text{h}^+$  is not the primary RS responsible for RhB degradation processes. Interestingly, it was observed that  $\cdot\text{OH}$  and  $\cdot\text{O}_2^-$  could have a weighty impact on photocatalytic processes because RhB degradation efficiencies sharply decline to around 50 % and 30 % by the introduction of IPA and p-BQ quenchers, respectively. Thus,  $\cdot\text{OH}$  and  $\cdot\text{O}_2^-$  radicals were involved in the photodegradation of RhB over  $\text{TiO}_2@Ti_3C_2$  nanocomposite [82]. However, this study did not consider that it is not possible to discard the high influence of  $\text{h}^+$  in RhB degradation processes. Because, although the three quenchers are at the same concentrations, IPA and p-BQ are at

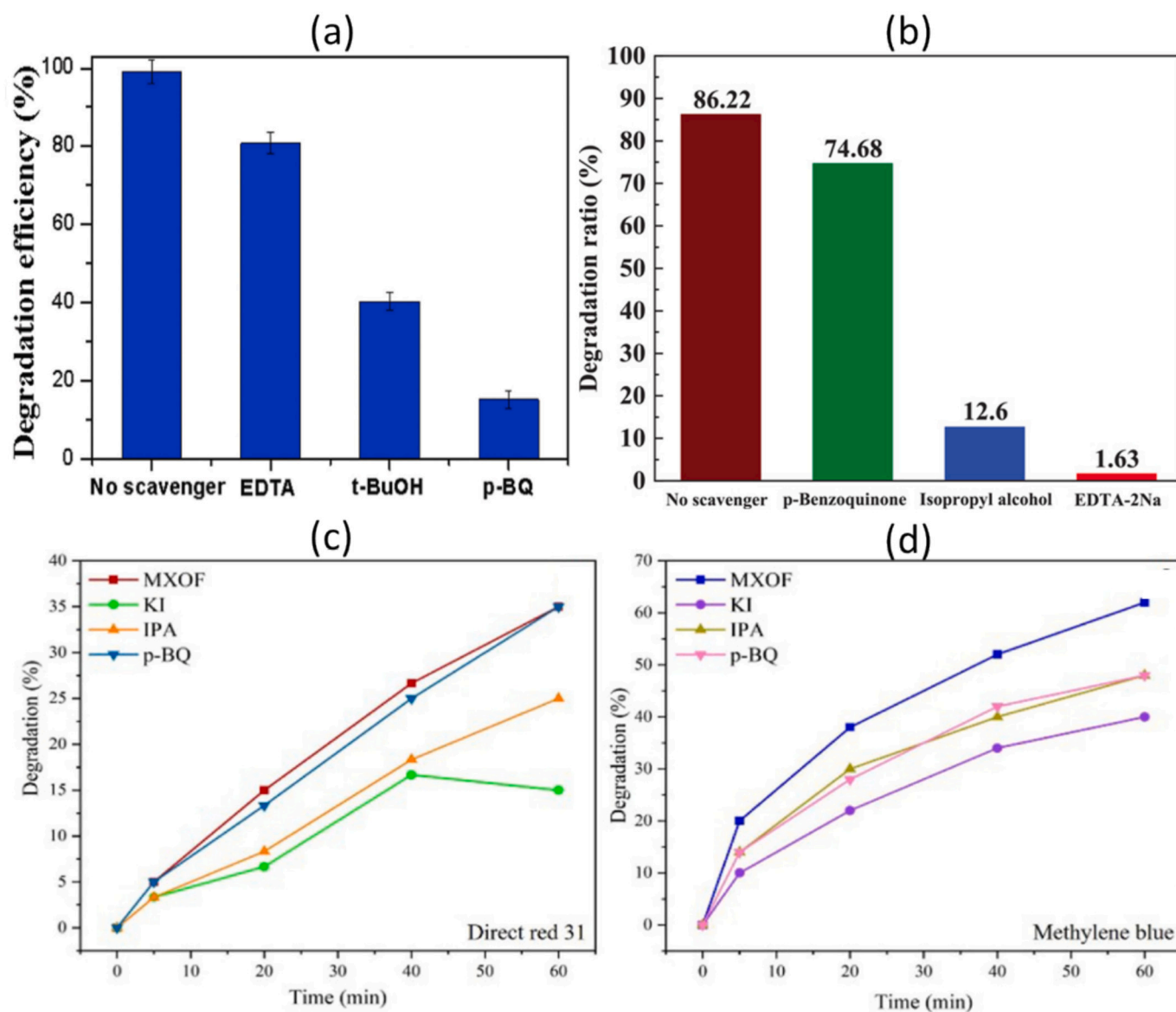


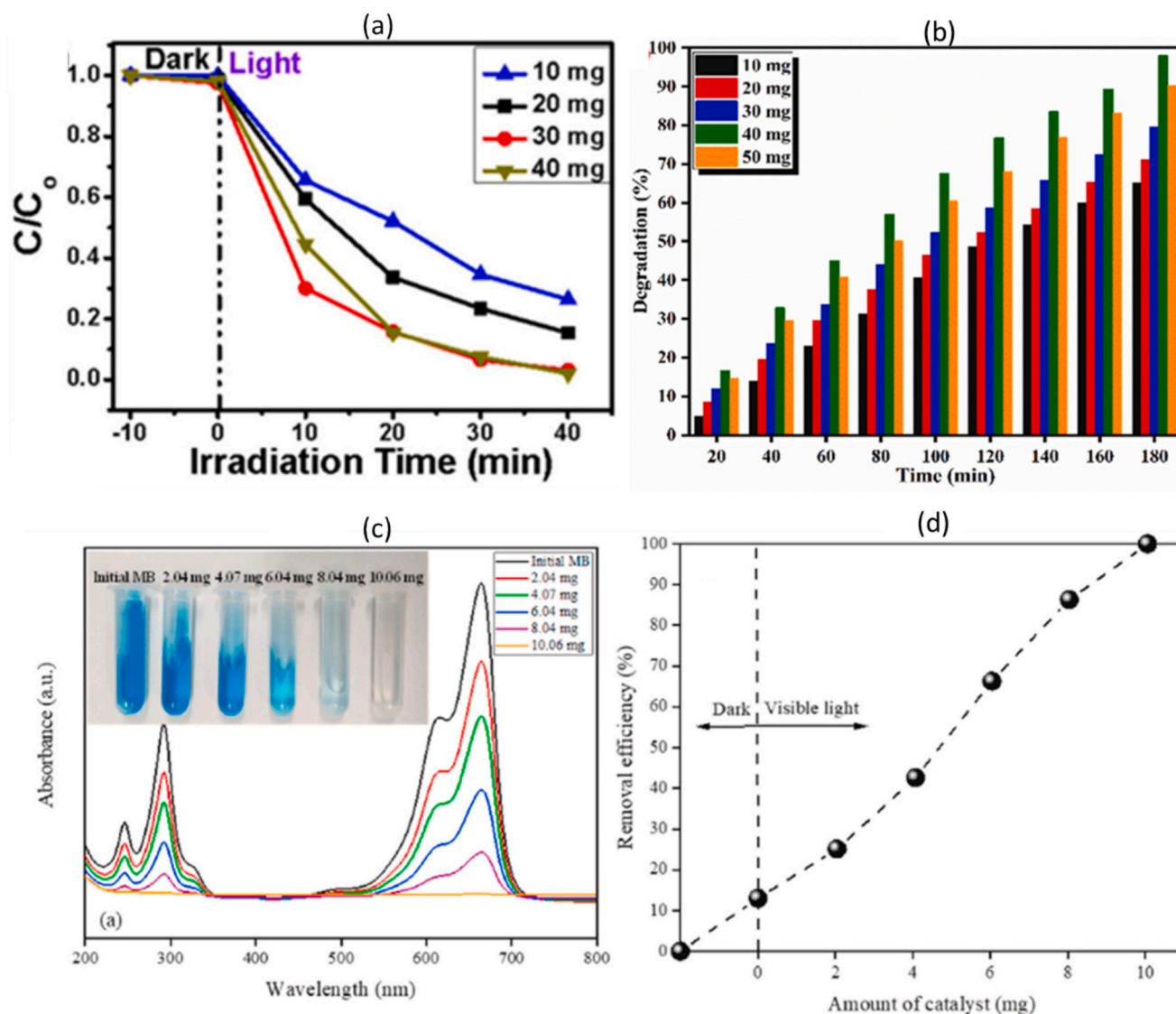
Fig. 8. Effect of radical scavenging on photocatalytic degradation of (a) MO using  $\text{Ti}_3\text{C}_2/\text{TiO}_2$  [107]. Reused with approval from Elsevier (Copyright 2021), (b) MB using  $\text{ZnO}/\text{Ti}_3\text{C}_2$  [87], (c) DR31 using MXene/Zn-MOF, (d) MB using MXene/Zn-MOF [113]. Reused with approval from Elsevier (Copyright 2023).

higher concentrations than RhB to react with the generated radicals, while the amount of EDTA is more than five times lower than that of  $\text{H}_2\text{O}_2$  to reduce the  $\text{h}^+$  of the VB.

In a similar study with  $\text{Fe}_2\text{O}_3/\text{ZnFe}_2\text{O}_4/\text{Ti}_3\text{C}_2$ , the evolution of reactive species was studied by adding 1 mM of BQ,  $\text{AgNO}_3$ , IPA, and EDTA-2Na to the RhB dye solution [106]. In this case, all the selected scavengers except IPA significantly reduced the degradation efficiency; it was suggested that the main RS were  $\cdot\text{O}_2^-$ ,  $\text{h}^+$ , and  $\text{e}^-$ . However, the result obtained with EDTA seems contradictory with respect to the other study. Thus, a plausible explanation for the result of this finding is that, as this material is not able to generate  $\cdot\text{OH}$  from the reaction of water with  $\text{h}^+$ . EDTA, as a good electron donor, could reduce  $\text{h}^+$ . In addition, the separation and transport efficiency of photo-generated electrons may be efficiently accelerated by the high conductivity of  $\text{Ti}_3\text{C}_2$  MXene due to its distinctive lamellar nanostructure [106]. However, this work does not show any mineralization study, which, as commented above, is crucial for the process's understanding. Same conclusions were obtained when Qu and Co-workers identified the reactive species during the photocatalytic degradation of MB with multilayer  $\text{Ti}_3\text{C}_2(\text{OH})_2$ 's as the catalyst [83]. This study showed that the major species were  $\text{h}^+$  and  $\cdot\text{O}_2^-$ ,

(Fig. 7c) with clear evidence from the scavenging study. It was further concluded that multilayer  $\text{Ti}_3\text{C}_2(\text{OH})_2$ 's valence band electrons were excited by visible light, which resulted in the production of  $\text{h}^+$ . Following the passage of photo-generated  $\text{e}^-$  to adsorbed  $\text{O}_2$ ,  $\cdot\text{O}_2^-$  is formed, which attacks and degrades MB completely. [83].  $\text{AgNO}_3$  as the scavenger revealed the role of electrons during the photocatalytic degradation of MB by MXene/ $\text{g-C}_3\text{N}_4$  (Fig. 7d) [104]. In this context, the activity of the MXene/ $\text{g-C}_3\text{N}_4$  was only moderately affected by the addition of other scavengers such as MeOH and IPA because the band gap of the material is not able to generate hydroxyl radicals from  $\text{h}^+$  of VB. The hydrated electron is expected to play a key function in the degradation of the MB dye pollutant by confirming the active sites contributing to the photocatalytic reaction; the degradation process is further supported by the  $\cdot\text{O}_2^-$ , and slightly by  $\cdot\text{OH}$  [104]. This study again does not perform any mineralization study and suggests the hydroxyl radicals generation from the VB of  $\text{g-C}_3\text{N}_4$ . However, this cannot occur because the VB value is 1 eV above the needed level to generate hydroxyl radicals from water.

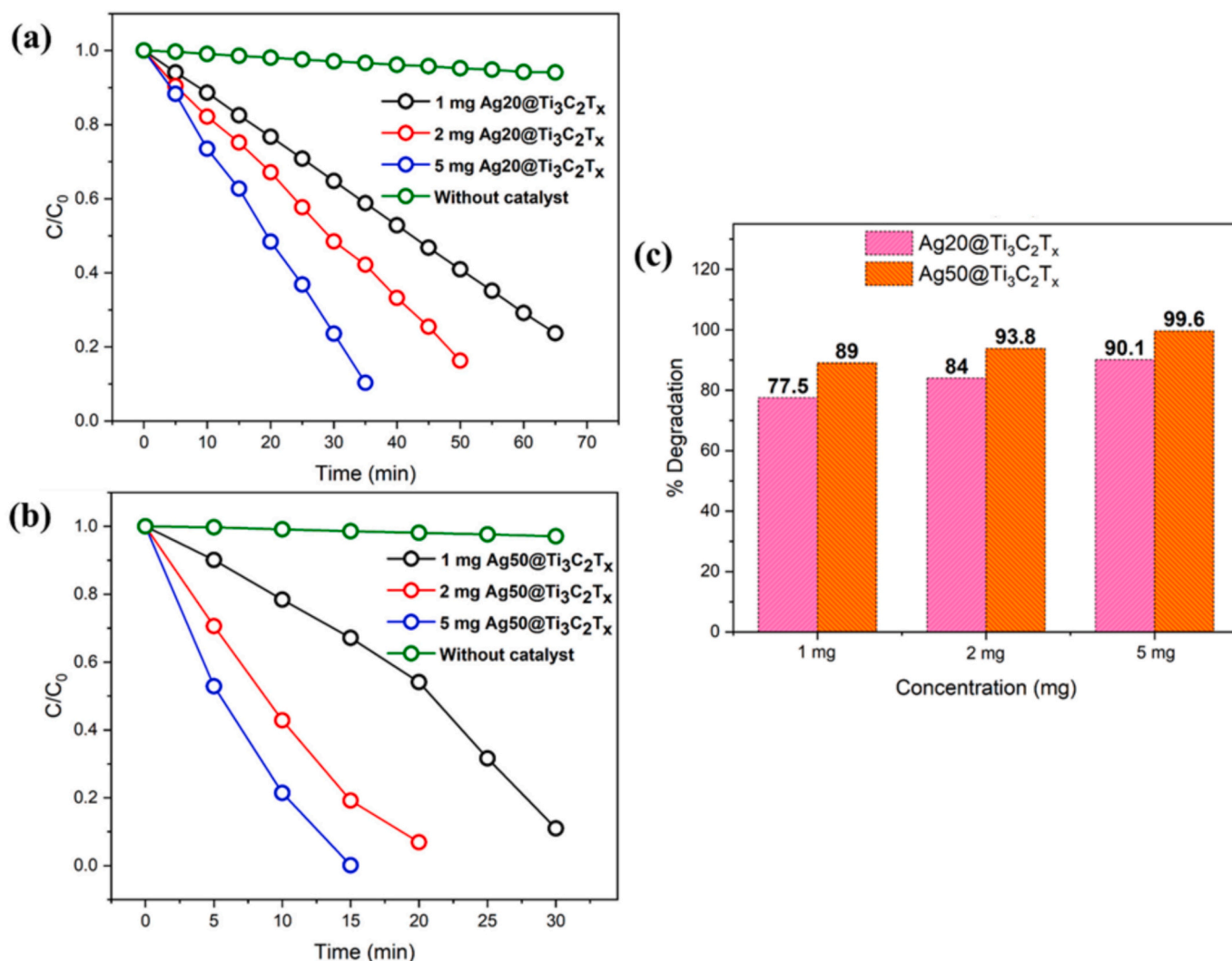
In general, the studies performed with MXene derivative materials containing  $\text{TiO}_2$  propose plausible mechanisms to explain the oxidation



**Fig. 9.** (a) Impact of dose on photocatalytic degradation of RhB utilizing  $\text{TiO}_2@/\text{Ti}_3\text{C}_2$  [82]. Reused with approval from Elsevier (Copyright 2020). (b) Effect of  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  dose on the degradation of MG [86]. Reused with approval from Elsevier (Copyright 2024). Effect of  $\text{ML-Ti}_3\text{C}_2(\text{OH})_2$  dose on the degradation of MB (c) UV-visible absorption spectra (d) % degradation efficiency [83]. Reused with approval from Elsevier (Copyright 2022).

of the dyes. In this context, the major species involved during the photoirradiation of  $\text{Ti}_3\text{C}_2-\text{TiO}_2$  for the degradation of MO were  $\bullet\text{O}_2^-$  and  $\bullet\text{OH}$ . Their contribution to the degradation mechanism was confirmed by scavenging experiments with EDTA, p-BQ, and *tert*-butanol. In this case, EDTA had only a little effect on the MO degradation efficiency. On the other hand, as shown in Fig. 8a, the MO photocatalytic performances drop sharply to around 15% and 40%, respectively, in the presence of p-BQ and *t*-BuOH, suggesting the significant role of corresponding radicals in the degradation [107]. Scavengers such as p-BQ, IPA, and EDTA-2Na were employed to scavenge  $\bullet\text{O}_2^-$ ,  $\bullet\text{OH}$ , and  $\text{h}^+$ , respectively, in  $\text{ZnO}/\text{Ti}_3\text{C}_2$  for the degradation of the MB solution. Fig. 8b shows that the inclusion of IPA and EDTA considerably inhibits the photodegradation of MB, although the presence of p-BQ somewhat reduces the process. These findings show that while  $\bullet\text{O}_2^-$  is not the major radical in this photocatalytic process,  $\bullet\text{OH}$  and  $\text{h}^+$  could be the primary active species [113]. Likewise, the major species in  $\text{MXene}/\text{Zn-MOF}$  for the photocatalytic degradation of DR31 dye was  $\bullet\text{O}_2^-$  as clear from the scavenging experiment with KI, IPA, and p-BQ, respectively, as shown in Fig. 8c. However, the major species involved in the degradation of MB with the same photocatalyst were  $\text{h}^+$ ,  $\bullet\text{OH}$ , and  $\bullet\text{O}_2^-$  as shown in Fig. 8d [87]. Conversely, in the case of photocatalytic degradation of RhB performed

by Ding et al. [116] using  $\text{TiO}_2@/\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$ , it was reported that  $\bullet\text{O}_2^-$  could be the major RS in this photocatalytic process as BQ inhibited the photocatalytic degradation of RhB dye compare to IPA ( $\bullet\text{OH}$  scavenger) and ammonium oxalate ( $\text{h}^+$  scavenger) and this is consistent with what was reported by [118]. In the same way, the degradation efficiency of MG using  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  was compared in the presence of scavengers like IPA, EDTA, BQ, and  $\text{AgNO}_3$ . The degradation efficiency was inhibited mainly with the addition of IPA, indicating the evolution of  $\bullet\text{OH}$  as the major RSs [86]. Many other studies identified the major RSs based on the scavengers and carried out the same protocols [122,126,127]. In general, it can be inferred that, depending on the targeted dye pollutants and the MXene material used, either of the RS ( $\text{h}^+$ ,  $\text{e}^-$ ,  $\bullet\text{OH}$ ,  $\bullet\text{O}_2^-$ ) can play a major role in the photocatalytic degradation operation. However, as commented before, to fully analyze the process effectiveness, it will be necessary to perform mineralization studies. Besides, for using the quenchers, one must take into account their concentrations, the reaction rate constant values between them and the reactive intermediates, and also the concentration of the dye.



**Fig. 10.** Effect of dose on photocatalytic action of (a) Ag<sub>20</sub>@Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> and (b) Ag<sub>50</sub>@Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub>; (c) % degradation efficiency [109]. Reused with open access approval from MDPI (Copyright 2024).

## 8. Effect of MXene-based photocatalyst dose

In terms of economic value, using the right amounts of photocatalyst in a photodegradation reaction can save treatment costs [82,86]. Thus, another crucial factor to be considered in the degradation of dye contaminants under the influence of irradiation is the quantity of the MXene employed, as this dictates the numerical strength of the active site available for interaction. For instance, Quyen's research team [82] investigated the impact of the dose of TiO<sub>2</sub>@Ti<sub>3</sub>C<sub>2</sub> heterojunction on RhB (10 mg/L) photocatalytic degradation efficiency. As expected, the RhB degradation efficiency for TiO<sub>2</sub>@Ti<sub>3</sub>C<sub>2</sub> quantity was increasing in a reasonable range of 143 to 429 mg/L (Fig. 9a). This rise in the rate of photocatalytic degradation results in an increased amount of active sites and increased exposure to RhB molecules. However, the photodegradation efficiency did not show any additional improvement with 571 mg/L of photocatalyst. This fact might be because of active surface overlapping at high doses, but it also could be attributed to the full absorption of the dose of light using 429 mg/L of TiO<sub>2</sub>@Ti<sub>3</sub>C<sub>2</sub> photocatalyst. On the other hand, while some positive effect of increased amount of catalysts, a negative effect of increased turbidity of the solution could "compensate" for it, thus not revealing an overall increase in the effectiveness. However, this point of view was not considered in this study (and also in the next), and they concluded that the ideal dosage may be thought of as 429 mg/L [82]. A similar trend was also

reported by Othman's team [84]. The MB and RhB photocatalytic degradation efficiency increased up to a certain catalyst dose, and beyond this dose, the dyes' photocatalytic degradation efficiency remained constant. In this case, the authors pointed out that this is because MXene materials have a shielding effect that can influence light penetration. This impact becomes more noticeable when catalyst loadings are large. However, the degradation productivity at larger MXBM loading remained the highest due to the increase in photocatalyst active sites. Furthermore, it was suggested that the SSA will be reduced due to the agglomeration of the material at higher doses [84]. Similar results were also observed with different doses of Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene for the degradation of MG (Fig. 9b) [86].

In another study, the MB (10 mg/L) photocatalytic degradation efficiency progressively increased from 0 % to 100 % with an increase in ML-Ti<sub>3</sub>C<sub>2</sub>(OH)<sub>2</sub> dose, as shown in Fig. 9c-d. The color of the reaction system may be used to visually measure the efficacy of MB degradation. The largest degradation rate (the slope of the tangent at each point) was observed at an ML-Ti<sub>3</sub>C<sub>2</sub>(OH)<sub>2</sub> dosage of 604 mg/L, even though MB was degraded fully at a dose of 1006 mg/L [83]. Doses of catalysts exceeding 100 mg/L seem to be impractical. Much lower optimal dose of MXBM was reported for Ag/Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> on the photocatalytic degradation of safranin (30 mg/L). As seen in Fig. 10a-c, an increase in the Ag/Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> mass to 5 mg/L yielded 99 % safranin degradation in 15 min, whereas 1 mg/L and 2 mg/L Ag/Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> loading demonstrated 90 % and 93 %

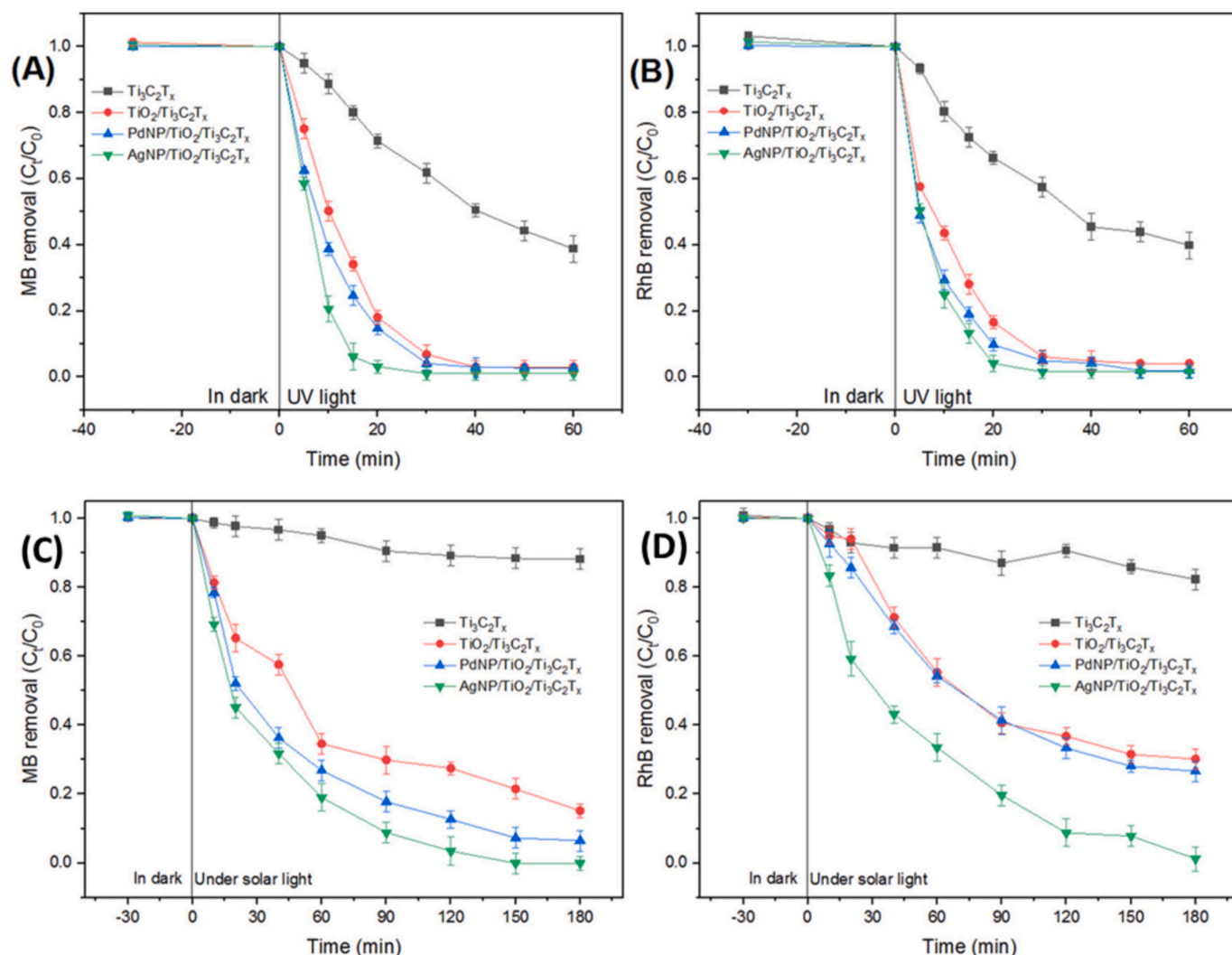


Fig. 11. Influence of irradiation time on the degradation of dyes by  $\text{Ti}_3\text{C}_2\text{Tx}$ -based nanocatalysts for (A) MB using UV, (B) RhB using UV, (C) MB using solar light, (D) RhB using solar light [84]. Reused with open access approval from the American Chemical Society (Copyright 2021).

photocatalytic degradation in 0.33 and 0.5 h, respectively [109]. A similar result was reported for the degradation efficiency of MO and CR with different doses of  $\text{BiVO}_4/\text{MXene}$ . Specifically, it was found that when the amount of  $\text{BiVO}_4/\text{MXene}$  is high, a much higher amount of  $\cdot\text{OH}$  and  $\cdot\text{O}_2^-$  radicals are produced, which leads to better photocatalytic activity, while the amount of catalyst is low, fewer oxidizing agents are present and this accounted for the increase in degradation efficiency with the increase in photocatalyst dose until optimum [145].

In short, most dye photocatalytic degradation operations were favored when the MXene dose was increased since more active sites for degradation were created, and there was an expansion of a large SSA for light harvesting. However, an excessively high dose may have negative effects due to the possibility of active surface overlapping, agglomeration, and shielding effects at high doses, which reduce the photocatalytic degradation efficiency. Depending on the characteristics of the MXene composite, there are variations in the dose level at which this happens. Typically, a dose in the range of 5–600 mg/L was reported as optimal, with 50 and 100 mg/L being the frequent optimal values. However, in these types of studies, light intensity is a key point that must be taken into account to determine the optimum photocatalyst dose. Too high a catalyst concentration can cause high turbidity and a significant drop in the photocatalytic process effectiveness.

## 9. Effect of irradiation time

The irradiation time greatly impacts the process effectiveness. Most degradation operations are completed at extended residence time. Additionally, residence time affects the kinetics of degradation [156]. The dye degradation with the irradiation time can also be observed based on the color change, as reported in many studies [83]. In a particular irradiation time study [84],  $\text{AgNPs}/\text{TiO}_2/\text{Ti}_3\text{C}_2\text{Tx}$  was discovered to have a fast kinetic response, and its photocatalytic degradation efficacy against MB and RhB rose very high in the first 60 min when exposed to solar irradiation and the first 15 min (Fig. 11). The reaction patterns with  $\text{PdNPs}/\text{TiO}_2/\text{Ti}_3\text{C}_2\text{Tx}$  and  $\text{TiO}_2/\text{Ti}_3\text{C}_2\text{Tx}$  were likewise comparable. According to the authors, it seems logical to assume that the dye molecule's interaction with photocatalysts will decline along with a drop in MB or RhB concentration. Additionally, according to the authors, when these dyes degrade, by-products like benzoic acid, 1,2-benzenedicarboxylic acid, thionin, and ethanedioic acid compete with the primary pollutant for the active sites on the MXBM's surface [84]. A similar trend was reported for the photocatalytic degradation of different dyes using  $\text{ZnO}/\text{Ti}_3\text{C}_2\text{Tx}$  (Fig. 12) [129] and other types of MXene-based materials [126,150,151].

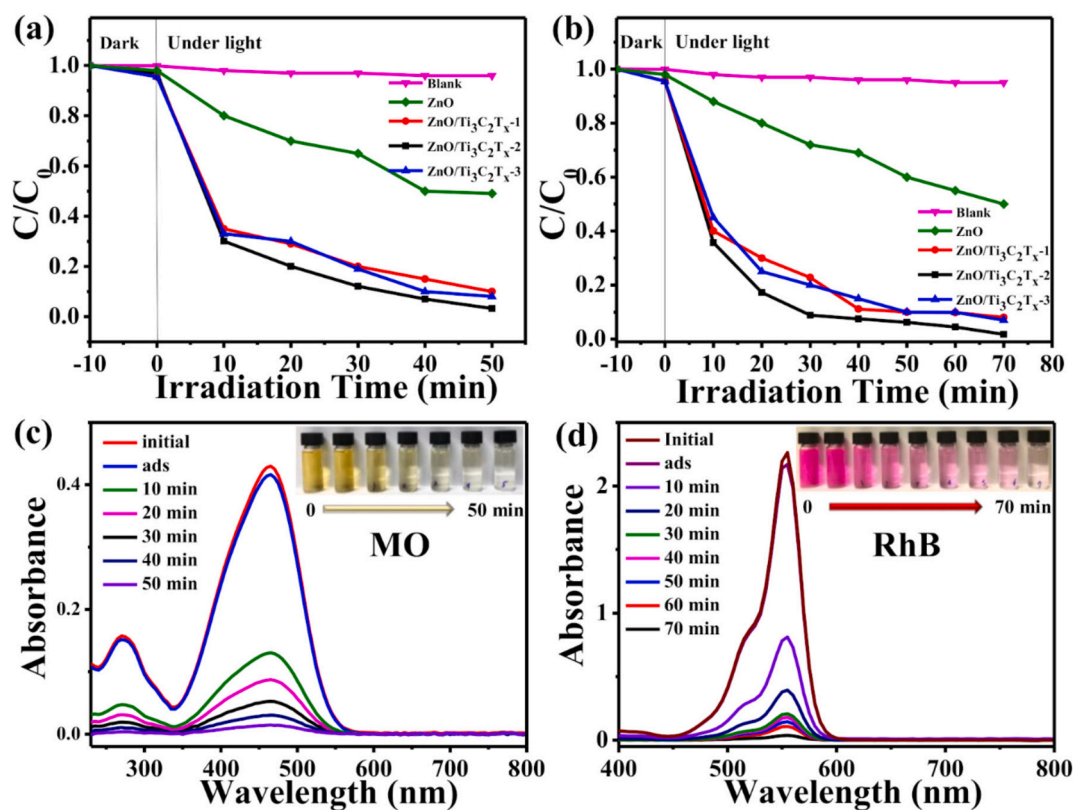


Fig. 12. Influence of irradiation resident time on ZnO/Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub> for the degradation of rhodamine blue and methyl orange and their corresponding UV-Visible absorption spectra [129]. Reused with open access approval from MDPI (Copyright 2020).

## 10. Effect of initial dye concentration

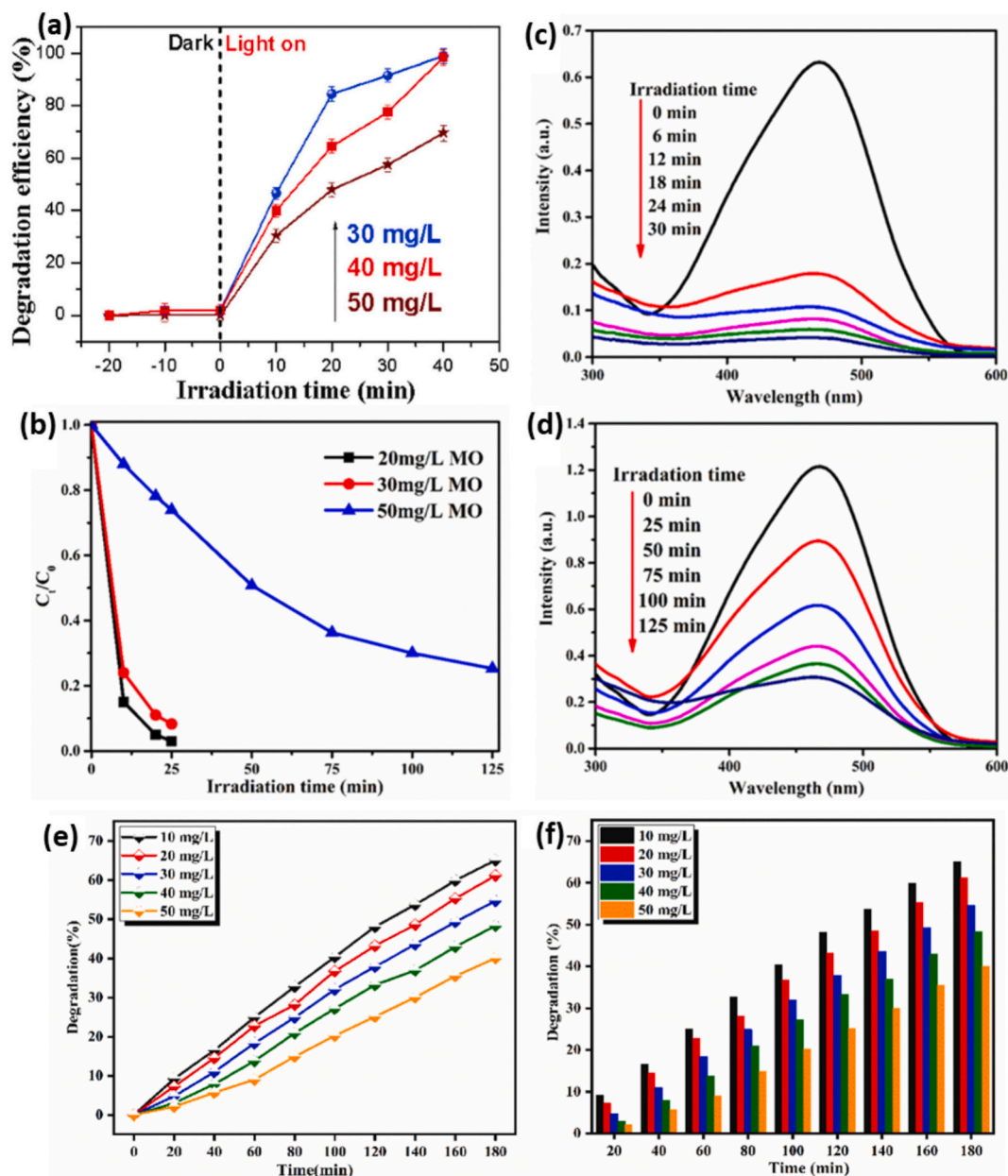
Another important parameter that affects photocatalytic degradation is the initial pollutant concentration. In this context, many studies have been performed from a low to a higher concentration to observe the effect. It is reported that higher dye concentrations diminished the degradation output due to the oversaturation of the employed nanomaterial [157]. There are a large number of studies that analyze the efficiency of a process by comparing the results of the percentage of degraded dye at different dye concentrations vs the irradiation time. However, this is not the correct way to evaluate this parameter because it must be compared to the amount of degraded dye vs irradiation time. For instance, in a particular study, the effect of the initial concentration of MO dye on the photocatalytic activity of Ti<sub>3</sub>C<sub>2</sub>-TiO<sub>2</sub> was explored by comparing the efficiency of MO degradation at varying concentrations using the percentage of degraded dye vs irradiation times, as it is shown in Fig. 13a. Thus, the MO degradation efficiencies decreased to 99 %, 98.5 %, and 70 %, respectively, when the starting concentrations were increased from 30 mg/L to 40 mg/L and 50 mg/L. However, calculating the amount of degraded dye from these data, it can be observed that close efficiencies are obtained at the three dye concentrations.

Thus, as in the study commented, most MO molecules interface and are exposed completely to the MXBM nanocatalyst at MO concentrations of 30 mg/L and 40 mg/L, which results in a full reaction and achieves optimal efficiency. Hence, it was suggested that the ideal MO concentration was 40 mg/L [107]. In this context, the degradation rate of RhB gradually decreases with increasing the initial concentration of RhB in their photocatalytic degradation by g-C<sub>3</sub>N<sub>4</sub>/Ti<sub>3</sub>C<sub>2</sub> [136]. However, although this is true, the efficiency of the dye photodegradation was similar. In this case, it was represented as  $C/C_0$  vs irradiation time, but the amount of degraded dye was very similar. It is clear that in photocatalyzed dye degradation at a fixed intensity of light and amount of photocatalyst, the efficiency of the process increases by adding more

amount of dye to reach a limit where all the surface of the activated photocatalyst is fully occupied by interactions with the dye, and this will be the optimum, higher amount of dye or do not change the efficiency of the reaction or decreases it, if other phenomena affecting the photocatalytic process are occurring such as light absorption of the dye.

In another research, Yao's group [119] investigated the effect of initial MO concentration on the photocatalytic degradation effectiveness of Ti<sub>3</sub>C<sub>2</sub>/MoS<sub>2</sub>. As displayed in Fig. 13b, as the concentration of the MO solution rises from 20 to 30 to 50 mg/L, the MXene's degrading efficiency generally falls. It is evident that after 25 min, almost 90 % of the MO solution at lower concentrations degrades. The variations in the UV absorption spectra of the 30 and 50 mg/L MO solutions are displayed in Fig. 13c and d, respectively. The photodegradation impact of MXene causes the high absorption spike of the MO solution at 554 nm to progressively decline. Additionally, the MXene sample shows a high ability to degrade MO (50 mg/L) in 125 min, with a degradation efficiency of around 80 %. The aforementioned findings demonstrate the potential for MXene photocatalysts to degrade organic pollutants at high concentrations. Similar observation was reported for other MXBM such as Ti<sub>3</sub>C<sub>2</sub>/TiO<sub>2</sub>/g-C<sub>3</sub>N<sub>4</sub>, Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene (Fig. 13e-f), and Ti<sub>3</sub>C<sub>2</sub>T<sub>x</sub>@Chitosan@g-C<sub>3</sub>N<sub>4</sub> [130,132,145,147,151].

This can be attributed to the following factors: (i) As the initial concentration of pollutants increases, the transmittance of the solution could be weakened, which extends the pathlength of photons to the increased surface and diminishes the number of photons absorbed by the photocatalyst; (ii) Higher initial pollutants concentrations, a larger number of pollutant molecules covered the active sites on photocatalyst surface, resulting in the reduction in radicals generation; (iii) at the fixed dosage of photocatalyst, the ability of photocatalytic system for radicals production is not increasing while increase of pollutant dose, thus the ratio between radical species and pollutant decreases. Finally, as the concentration of pollutants increases, more pollutants and degradation intermediates compete with the limited radicals [147].



**Fig. 13.** Effect of different initial concentrations of dye pollutant in the matrix on photocatalytic degradation of MO (a) using  $\text{Ti}_3\text{C}_2\text{-TiO}_2$  [107]. Reused with approval from Elsevier (Copyright 2021), (b) using  $\text{Ti}_3\text{C}_2/\text{MoS}_2$  (c) UV-Vis absorption bands for 30 mg/L and (d) for 50 mg/L of MO solutions following irradiation by a 400W metal halide lamp, respectively [119]. Reused with open access approval from Springer (Copyright 2020) (e-f). Effect of different initial concentrations of dye pollutant in the matrix on photocatalytic degradation of MG using  $\text{Nd-NiCo}_2\text{O}_4/\text{MXene}$  [86]. Reused with approval from Elsevier (Copyright 2024).

### 11. Effect of ionic species/co-existing ions

It has been established that the presence of ionic species like NaCl and KCl, which are often found in industrial runoffs and are often used in some stages of dyeing processes, affects their degradation [134,158,159]. In photocatalysis, the existence and increase of ionic species can either enhance or reduce the degradation of dye pollutants or have no discernible effect [134]. The influence of the ionic strength of the solution on the photocatalytic elimination of RhB on  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  MXBM was studied by Tu et al. [124]. The outcome demonstrated that the NaCl greatly decreased the photocatalytic degradation of RhB, equivalent to the  $k$  values spanning from  $0.0483\text{ min}^{-1}$  to  $0.0139\text{ min}^{-1}$  with NaCl concentration between 0 and 0.08M. Here, as shown in Fig. 14, the inorganic ions inhibited both the photocatalytic and adsorption processes. This may have been caused by the removal of free

radicals and photo-excited holes.

There are two possibilities in the presence of inorganic salt during the photocatalytic oxidation. One is the competitive reaction between the ions and reactive species. The second possibility is the adsorption on the reaction site. The electrostatic interaction relationship caused the negatively charged  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  MXBM to attract the  $\text{Na}^+$ , eventually exhausting its active sites. The reaction between  $^{\bullet}\text{OH}$  and  $\text{Cl}^-$  formed the  $\text{HOCl}^{\bullet-}$  and next the  $\text{Cl}^{\bullet}$ , i.e., served as an  $^{\bullet}\text{OH}$  scavenger to inhibit the entire degradation process. So, to maintain higher photocatalytic capabilities in future applications, it is imperative to lower the salt content [124].

Tan's research group [134] achieved a different result on the effect of ionic species on the photocatalytic degradation of dye pollutants by  $\text{Ag}_3\text{PO}_4/\text{Ti}_3\text{C}_2\text{-TiO}_2$  MXBM. Notably, the addition of potassium chromate significantly decreased the photocatalytic effectiveness of  $\text{Ag}_3\text{PO}_4/$

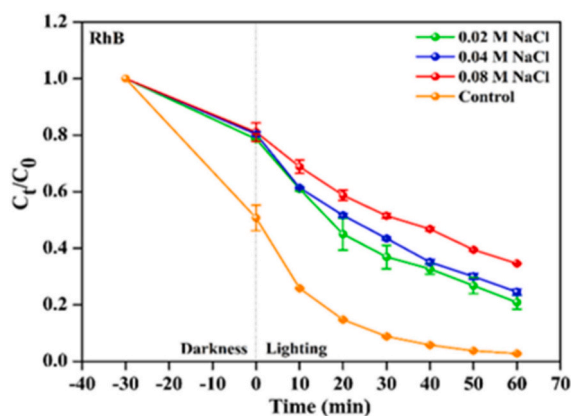


Fig. 14. Influence of ionic species on degradation of RhB using  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  under light illumination [124]. Reused with approval from Elsevier (Copyright 2022).

$\text{Ti}_3\text{C}_2\text{-TiO}_2$ , as seen in Fig. 15. This is because the strong oxidizing ability of potassium dichromate and the strong activity of loaded  $\text{Ag}_3\text{PO}_4$  impact the photocatalytic performance. The fact that  $\text{Ag}_3\text{PO}_4$  has high oxidation and will reduce to elemental Ag when exposed to

irradiation, but the introduction of inorganic species suppresses this process, may be the reason why the introduction of supplementary inorganic ions accelerated the degradation of dye contaminants. Thus, enhancing the material's photocatalytic activity in the process. The aforementioned findings demonstrate that  $\text{Ag}_3\text{PO}_4/\text{Ti}_3\text{C}_2\text{-TiO}_2$  composite has strong applicability and future application prospects. It can adjust/adapt to most inorganic salts and the humic acid-containing environment, with the exception of  $\text{Cr}_2\text{O}_7^{2-}$  [134].

Overall, it can be concluded that degrading dye pollutants from effluent is more challenging than degrading a one-dye-containing solution, because wastewater comprises a variety of inorganic ions and heavy metals in addition to organic pollutants. Inorganic ions can change the ionic strength of the solution and impact its photocatalytic degradation capabilities [158,160]. These ions may directly interact with active sites on the MXBM surface [161]. In certain instances, though, the effect can be advantageous. Competitive tests are especially crucial for assessing the MXBM's capability for industrial wastewater treatment.

## 12. Effect of pH

The initial pH of the solution is critical in any of the water treatment processes [162,163]. In photocatalysis, pH can affect the degree of ionization of the dye pollutant as well as functional groups, thus the

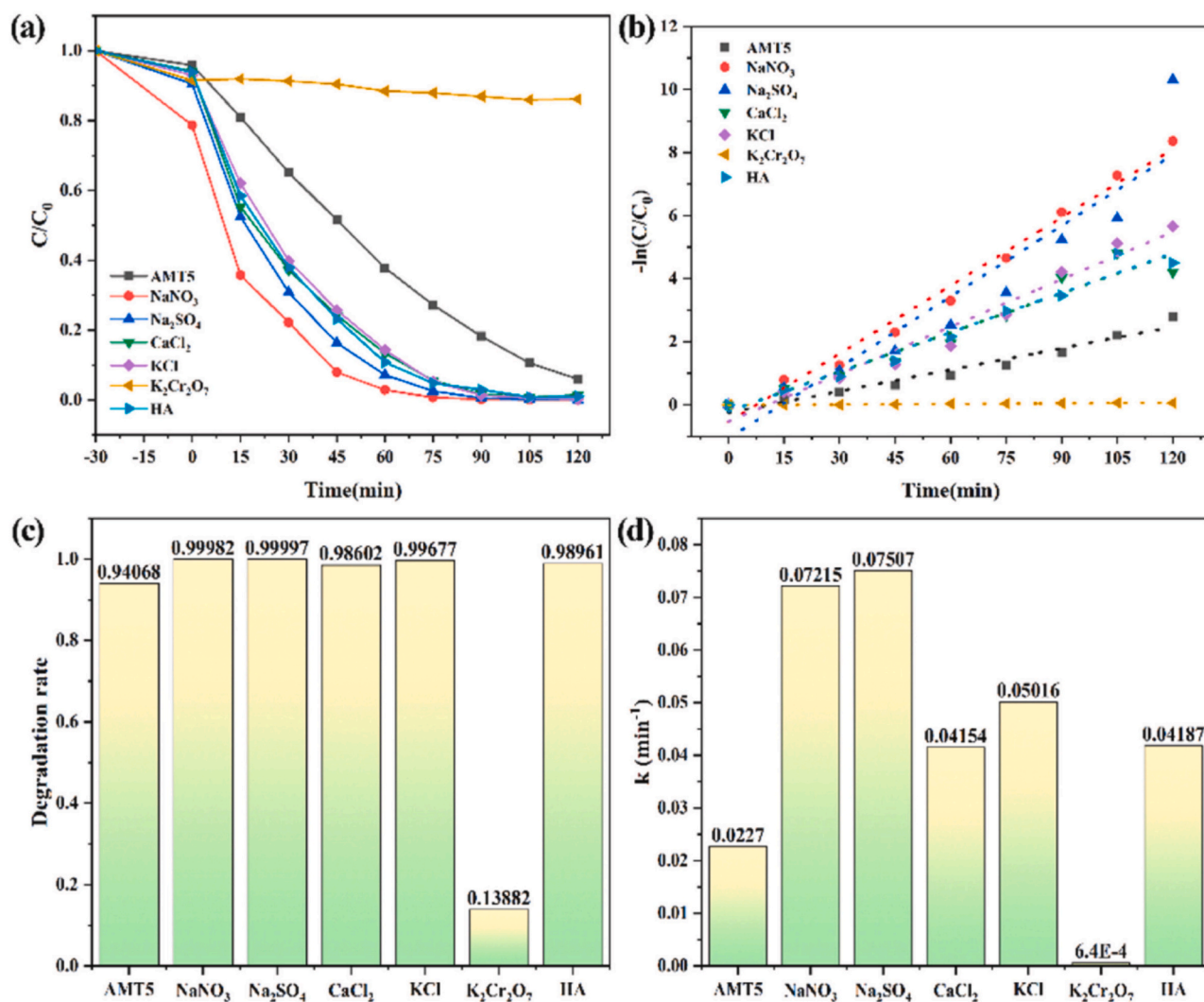


Fig. 15. Influence of diverse inorganic salts on the photocatalytic efficiency of  $\text{Ag}_3\text{PO}_4/\text{Ti}_3\text{C}_2\text{-TiO}_2$  (a) the equivalent degradation output of dye contaminants; (b) Rate constant; (c) and (d) are histograms of equivalent values [134]. Reused with approval from Elsevier (Copyright 2022).

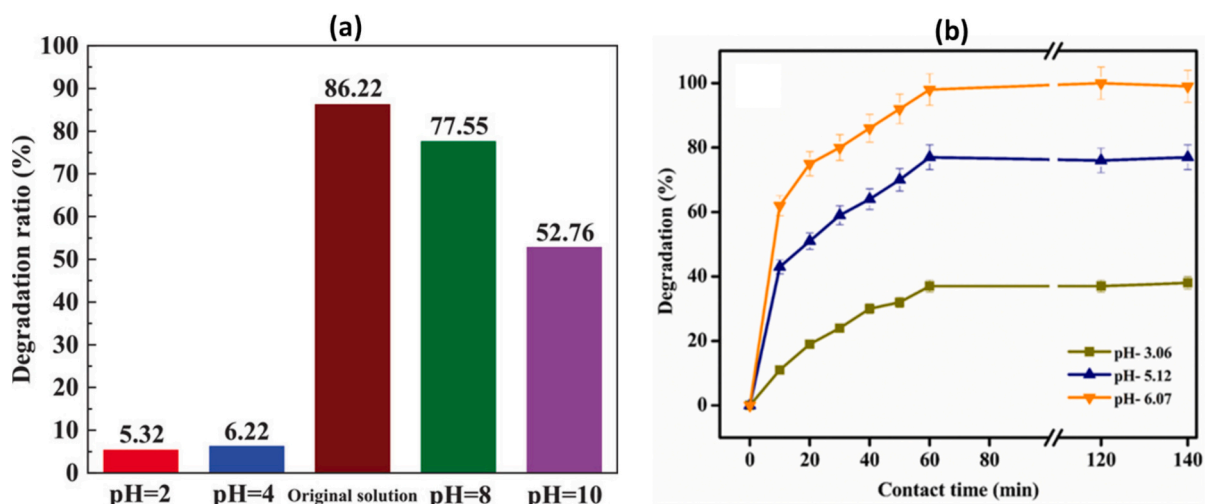


Fig. 16. Effect of different initial pH on photocatalytic degradation of (a) MB dye pollutant solution using ZnO/Ti<sub>3</sub>C<sub>2</sub> [113]. Reused with approval from John Wiley and Sons (Copyright 2022), (b) RhB dye pollutant solution using NiCo<sub>2</sub>S<sub>4</sub>/MXene [132]. Reused with approval from Elsevier (Copyright 2020).

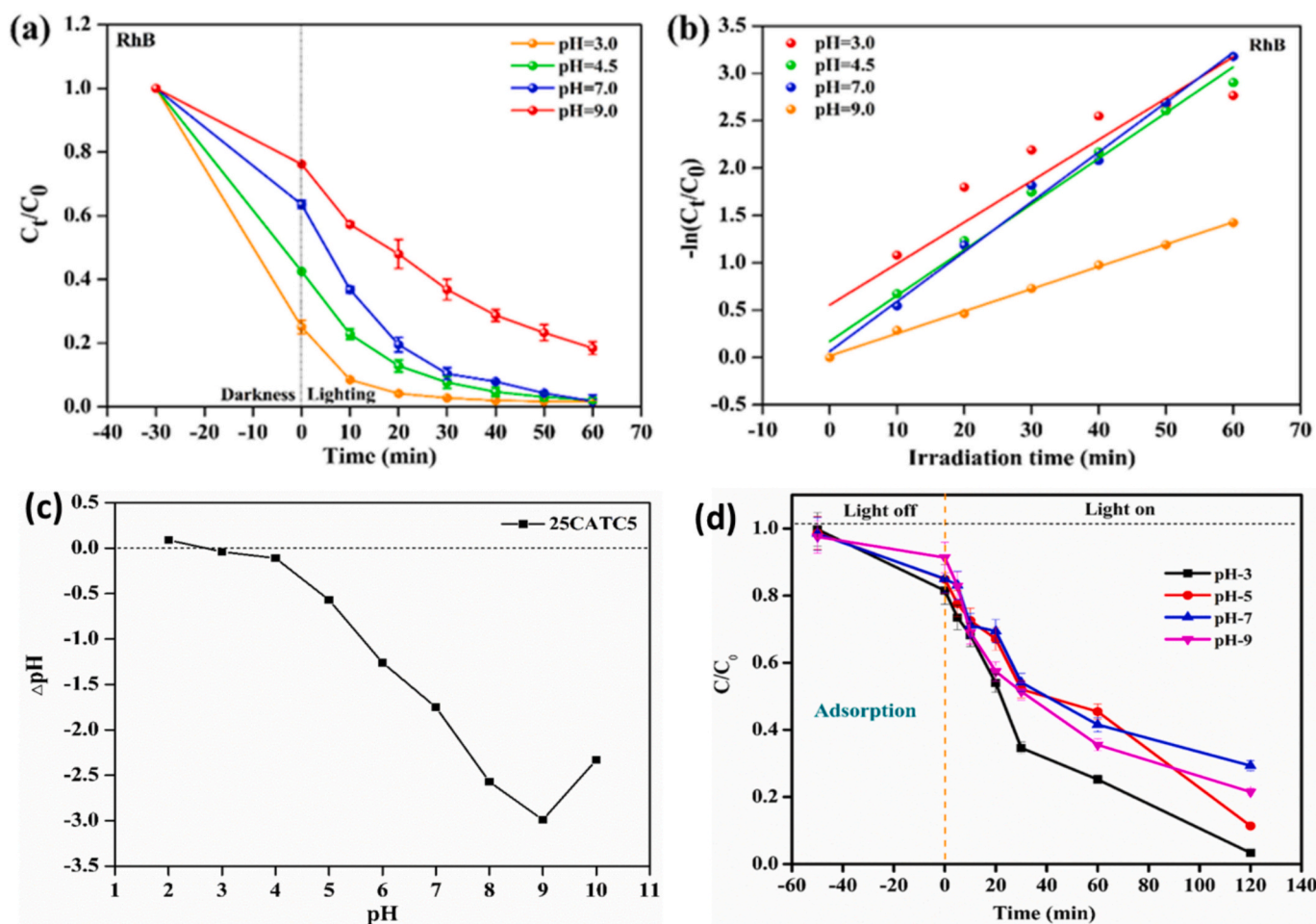


Fig. 17. (a) Effect of pH on RhB degradation using Ti<sub>3</sub>C<sub>2</sub>/g-C<sub>3</sub>N<sub>4</sub> when exposed to visible light illumination, (b) Pseudo-1st order rate constants of rhodamine blue when degraded by Ti<sub>3</sub>C<sub>2</sub>/g-C<sub>3</sub>N<sub>4</sub> at different initial pH, (c) Dependence of pH on  $\Delta pH$  during the equilibration of Ti<sub>3</sub>C<sub>2</sub>/g-C<sub>3</sub>N<sub>4</sub> [124]. Reused with approval from Elsevier (Copyright 2022), (d) Effect of pH on MB and RhB degradation using Ti<sub>3</sub>C<sub>2</sub>Tx@Chitosan@g-C<sub>3</sub>N<sub>4</sub> [130]. Reused with approval from Elsevier (Copyright 2020).

adsorption capacity of the photocatalyst, etc. [16,124,164,165]. Also, pH governs the photocatalytic mechanism process since the operation involves using radicals like  $\cdot OH$ . However, only a few investigations are

available on this effect. For example, Luo's research group investigated the influence of pH on the catalytic degradation of MB using ZnO/Ti<sub>3</sub>C<sub>2</sub> under light illumination. Fig. 16a displays the degradation efficiency of

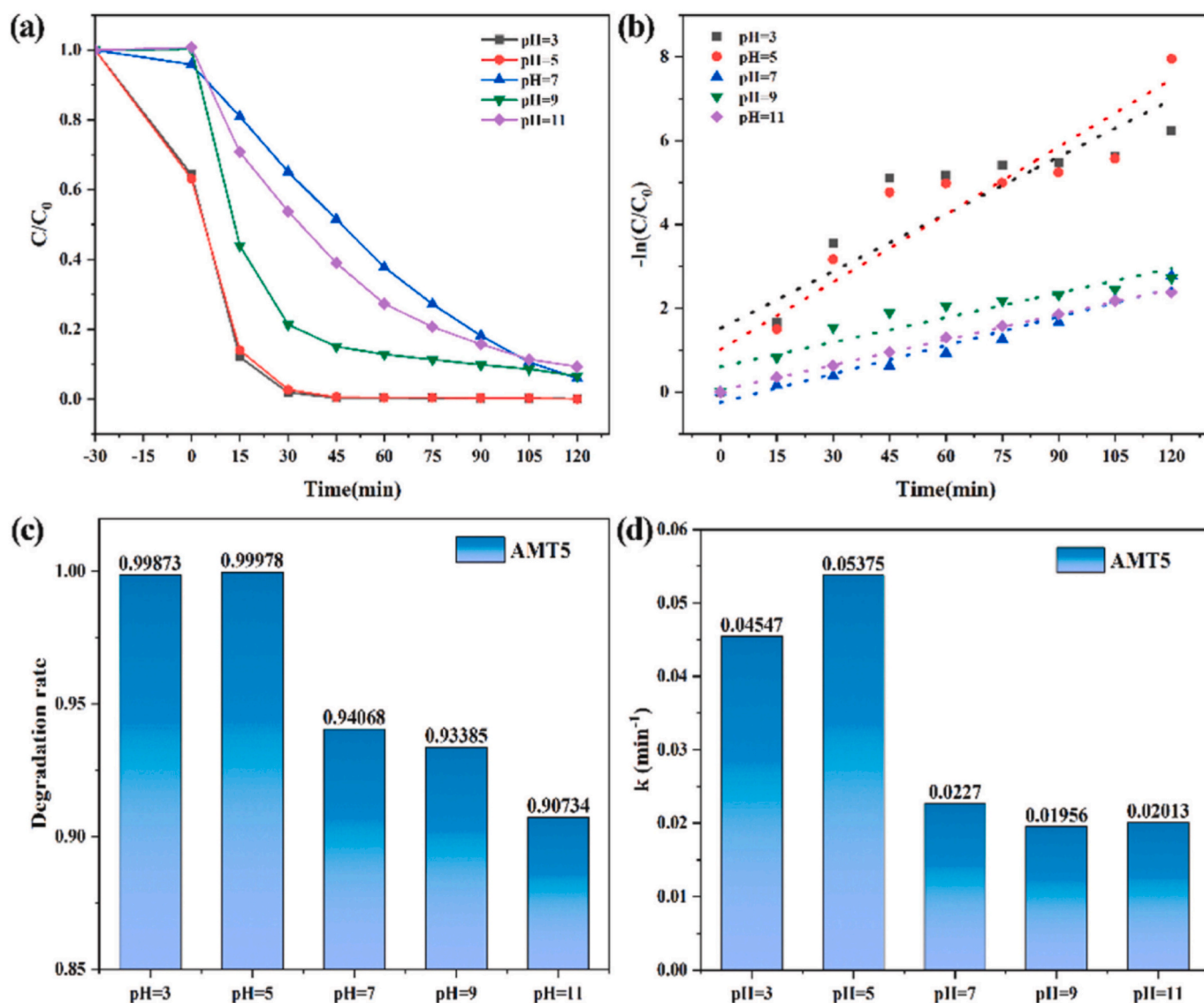


Fig. 18. Photocatalytic efficiency of  $\text{Ag}_3\text{PO}_4/\text{Ti}_3\text{C}_2\text{-TiO}_2$  under various pH (a) the equivalent degradation rate of RhB; (b) Rate constant; (c) and (d) are histograms of equivalent values [134]. Reused with approval from Elsevier (Copyright 2022).

MB over  $\text{ZnO}/\text{Ti}_3\text{C}_2$  in 60 min at various pH values. The dissolution of  $\text{ZnO}$  in acidic conditions causes the degradation efficiency of MB to drop drastically between pH 2 and 4. Moreover, for pH values of 8 and 10, the photocatalytic efficiencies also drop because the basic solution's  $\text{OH}^-$  might lessen the generation of  $\cdot\text{OH}$  [113]. This claim about better photocatalytic degradation performance of MXene at a neutral pH region is consistent with what was reported by Vigneshwaran research group [132] and Khaneghahi's team [151] for the degradation of RhB using  $\text{NiCo}_2\text{S}_4/\text{MXene}$  under 250 W Xe lamp illumination for 140 min as shown in Fig. 16b and for the degradation of MB using  $\text{Ti}_3\text{C}_2(\text{OH})_2$  under 150 W Xenon Arc Lamp irradiation for 80 min respectively.

In another study [124], the photocatalytic degradation of 20 mg/L RhB employing 20 mg of  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  is explored at different pH levels. As shown in Fig. 17a, RhB's degradation process showed a distinct trend, with a clear decline in RhB's degradation efficiency at pH 9. At pH levels of 3, 4.5, 7, and 9, the  $k$  values were  $0.044 \text{ min}^{-1}$ , 0.048, 0.052, and  $0.023 \text{ min}^{-1}$ , respectively, as shown in Fig. 17b. Specifically, the data shown in Fig. 17 indicate that RhB degradation favored an acidic and neutral reaction environment. The charge characteristics for  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  MXBM were found by dipping the material into 0.1 mol/L KCl solutions and stirring it for 24 h. The  $\Delta\text{pH}$  values were related to these

surface charge characteristics. Fig. 17c illustrates that the  $\Delta\text{pH}$  was negative from pH 3 to 11, and below this pH range, it was positive. These findings suggested that the surface of  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  is negatively charged. This explains why the degradation of the targeted contaminant, cationic RhB dye (which is positively charged), is very high between pH 3–11 but very poor below this pH range. Nonetheless, the outcome of the experiment showed that the effectiveness of photocatalysis and adsorption in an acidic environment was greater than in an alkaline one. Therefore, the adsorption and degradation of the RhB were not primarily driven by the electrostatic attraction relationship. However, the H-bond that forms between the protonated  $\text{Ti-O}$ , the unprotonated  $\text{Ti-OH}$ , and the rhodamine blue present in aqueous solution may improve  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$ 's adsorption ability. Through the noncovalent surface- $\pi$  attraction, the aromatic groups on RhB and the peroxy groups on  $\text{Ti}_3\text{C}_2/\text{g-C}_3\text{N}_4$  material interacted. Because of the attack of photoexcited  $h^+$ , the acidic aqueous solution further speeds up the reaction process. Conversely, the reduction in photodegradation efficiency at higher pH values is caused by the generation of less active free radicals ( $\text{HO}_2\cdot$ ). As shown in Fig. 17d, a similar acidic pH positive effect for the photocatalytic degradation of MB and RhB using  $\text{Ti}_3\text{C}_2\text{Tx@Chitosan@g-C}_3\text{N}_4$  was observed [130]. Tan and co-workers also found a similar observation when exploring the

Table 2

Regeneration and recyclability of MXene towards photocatalytic degradation of various dye pollutants in brief.

MXenes-based photocatalyst	Dye degraded	Eluent/Regenerant	% effectiveness at n = 1	Number of cycles(n)	% effectiveness after n-th cycle	References
Sm <sub>2</sub> O <sub>3</sub> /MXene/g-C <sub>3</sub> N <sub>4</sub>	Orange G	Washed and dried	~100	5	~95	[93]
MXene@Cu/Tb/MgFe <sub>2</sub> O <sub>4</sub>	Crystal violet	DI H <sub>2</sub> O	95	5	79	[94]
NiMoO <sub>4</sub> /MXene	Methylene blue	99.5 % pure EtOH	97	5	91	[96]
MXene@V-CuO	Malachite green	Washed and dried	~88	5	78	[97]
g-C <sub>3</sub> N <sub>4</sub> /MXene	Methylene blue	Washed and dried	100	4	~85	[98]
MXene@g-C <sub>3</sub> N <sub>4</sub> /TiO <sub>2</sub>	Rhodamine B	Washed	~99	4	>90	[99]
MXene/CuO/B-g-C <sub>3</sub> N <sub>4</sub>	Malachite green	Washed	>90	5	>90	[100]
MnWO <sub>4</sub> /MXene/carbon xerogel	Crystal violet	Washed	95	5	86	[101]
MXene/ZnO-tetrapod	Rhodamine B	Washed	~78	4	>62	[102]
ZnO/NiWO <sub>4</sub> /V <sub>2</sub> C MXene	Rhodamine B	Washed	~92	4	86	[103]
ZnO/NiWO <sub>4</sub> /V <sub>2</sub> C MXene	Methylene blue	–	~93	4	89	[103]
Ti <sub>3</sub> C <sub>2</sub> /TiO <sub>2</sub>	Rhodamine B	–	~94	4	90	[81]
TiO <sub>2</sub> -Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	Washed	97	5	97	[82]
MXene-COOH@(PEI/PAA) <sub>10</sub>	Methylene blue	EtOH	85	8	65	[173]
MXene@Fe <sub>3</sub> O <sub>4</sub>	Methylene blue	EtOH	>80	5	77	[174]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub>	Methyl orange	deionized H <sub>2</sub> O	~97	4	~93	[110]
Fe <sub>2</sub> O <sub>3</sub> /ZnFe <sub>2</sub> O <sub>4</sub> @Ti <sub>3</sub> C <sub>2</sub>	Rhodamine B	–	97	4	>90	[106]
TMAOH-MXene	Bromocresol green	Deionized H <sub>2</sub> O	100	2	70	[89]
TiO <sub>2</sub> /Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	–	92	5	~88	[111]
p-MXene/SWCNTs	Methylene blue	H <sub>2</sub> O and EtOH	95	5	~95	[175]
V <sub>2</sub> CT <sub>x</sub>	Methylene blue	EtOH	~76	4	>50	[176]
NiCo <sub>2</sub> S <sub>4</sub> /MXene	Rhodamine B	0.1 M HCl	99	5	~98	[132]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Methylene blue	Double-distilled H <sub>2</sub> O and 99.9 % pure EtOH	99	5	~93	[130]
Sodium alginate aerogel/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Congo red	Pure EtOH and deionized H <sub>2</sub> O	~96	5	70	[177]
Ti <sub>3</sub> C <sub>2</sub> (OH) <sub>2</sub>	Methylene blue	Double-distilled H <sub>2</sub> O	~97	3	~89	[151]
Sodium alginate aerogel/d-Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methylene blue	5 % EtOH and deionized H <sub>2</sub> O	88	10	84	[178]
Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Methyl orange	–	99	5	~97	[107]
g-C <sub>3</sub> N <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> /TiO <sub>2</sub>	Rhodamine B	EtOH and deionized H <sub>2</sub> O	>95	5	~95	[179]
TiO <sub>2</sub> /MXene	Methylene blue	–	~99	5	~84	[85]
MXene/Zn-MOF	Direct red 31	EtOH and acetone (50:50)	35	5	~35	[87]
TiO <sub>2</sub> /MXene	Methyl orange	–	~91	5	~87	[108]
MXene/Zn-MOF	Methylene blue	EtOH and acetone (50:50)	62	5	~62	[87]
Ti <sub>3</sub> C <sub>2</sub> /BiFeO <sub>3</sub>	Congo red	–	100	4	~100	[88]
TMAOH-MXene	Methylene blue	Deionized H <sub>2</sub> O	100	2	70	[89]
ZnO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	Deionized H <sub>2</sub> O	~86	5	~70	[113]
ZnO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	–	~95	3	>90	[114]
Ti <sub>3</sub> C <sub>2</sub> /MoS <sub>2</sub>	Methyl orange	Centrifugation	~97	3	>75	[119]
Ti <sub>3</sub> C <sub>2</sub> /ZnO	Rhodamine B	–	~98	7	>90	[59]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Methyl orange	–	~99	5	>90	[78]
ZnO@Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	Without treatment	~93	6	~89	[121]
ZnO/V <sub>2</sub> C	Methylene blue	Centrifugation	~100	4	~97	[60]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Methylene blue	–	~100	5	>95	[78]
Ag <sub>3</sub> PO <sub>4</sub> /Ti <sub>3</sub> C <sub>2</sub> -TiO <sub>2</sub>	Rhodamine B	Centrifugation	~94	8	80	[134]
ZnO/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Methyl orange	–	~100	5	>95	[129]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> /Ti <sub>3</sub> AlC <sub>2</sub> @Ag	Rhodamine B	–	~99	5	>95	[78]
Nb <sub>2</sub> C/CdS	Rhodamine B	–	99	3	~80	[128]
Nd-NiCo <sub>2</sub> O <sub>4</sub> /MXene	Malachite green	DI H <sub>2</sub> O and EtOH	98	5	>80	[86]
ZnO/Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub>	Rhodamine B	Washed with deionized H <sub>2</sub> O	~100	5	>95	[129]
ZnO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	Centrifugation	~99	4	>97	[149]
Ti <sub>3</sub> C <sub>2</sub> T <sub>x</sub> @Chitosan@g-C <sub>3</sub> N <sub>4</sub>	Rhodamine B	DD H <sub>2</sub> O and EtOH	~99	5	~92	[130]
CuO/Ti <sub>3</sub> C <sub>2</sub>	Methylene blue	–	~100	3	>80	[131]
NiFe <sub>2</sub> O <sub>4</sub> /MXene	Methylene blue	–	74	3	>70	[137]
MXene/PbCrO <sub>4</sub>	Methylene blue	DI H <sub>2</sub> O and EtOH	~93	3	90	[150]

effect of different pH conditions on the photocatalytic efficacy of Ag<sub>3</sub>PO<sub>4</sub>/Ti<sub>3</sub>C<sub>2</sub>-TiO<sub>2</sub> for the degradation of RhB [134]. In this study, the solution pH was adjusted to 3, 5, 9, and 11. Fig. 18 demonstrates that the photocatalytic degradation and adsorption capacity are substantially higher in acidic conditions as opposed to neutral and alkaline conditions. This could be due to the increased incidence of hydrogen ions in acidic solutions, which react with •O<sub>2</sub><sup>-</sup> in the CB to form H<sub>2</sub>O<sub>2</sub>. This H<sub>2</sub>O<sub>2</sub> then reacts with •O<sub>2</sub><sup>-</sup> to yield •OH radicals, boosting the efficiency of the photocatalytic process. Ag<sub>3</sub>PO<sub>4</sub>'s strong oxidizing ability allowed dyes to oxidize more readily in the presence of H<sup>+</sup>. Nevertheless, the photocatalytic performance will not significantly improve with additional pH lowering. Since the Ag<sub>3</sub>PO<sub>4</sub>/Ti<sub>3</sub>C<sub>2</sub>-TiO<sub>2</sub> reaction mechanism has a restricted rate of H<sup>+</sup> consumption, too much H<sup>+</sup> will cause the reaction to be restricted. In comparison to neutral conditions, the degradation

rate was marginally lower under basic conditions during the first hour of photocatalytic degradation. This might be the result of higher OH<sup>-</sup> concentration, which increased Ag<sub>3</sub>PO<sub>4</sub>/Ti<sub>3</sub>C<sub>2</sub>-TiO<sub>2</sub> oxidation potential, and an interaction between the h<sup>+</sup> in the VB and OH<sup>-</sup> in the solution that produces •OH, which facilitates the photocatalytic operation [134]. Many other researchers [145,147] also established the ability of MXene and its composite to provide high effectiveness in an acidic medium rather than an alkaline medium, and this proves the chemical stability of this emerging photocatalyst in difficult environmental conditions.

### 13. Regenerability and recyclability

From an economic, ecological, and sustainability perspective, the reusability of MXenes should be ultimately considered for an all-around

environmental application as per pollutant elimination, especially in real wastewater treatment scenarios and industrial/commercial scale level [166–168]. Regeneration-recyclability is an adsorption-desorption process of recovering the expended photocatalytic material (MXene) from the reaction mixture, restoring the photocatalyst integrity/capacity by removing pre-adsorbed pollutant (adsorbate) from the surface with the aid of eluent and then reusing the nanocatalyst in the next photocatalytic degradation operation [167,169–172]. This is the desired way; however, in many cases, additional operations to recover full catalytic properties can also be needed (for example, heating in high temperatures, calcination, or treatment with specific chemicals).

Various research works on the reusability of MXenes for the degradation of dye contaminants are analyzed and summarized in Table 2. For example, in a regeneration-reusability experiment, Luo's research group [113] used deionized H<sub>2</sub>O to wash the photocatalyst (desorption of remaining pollutants) of ZnO/Ti<sub>3</sub>C<sub>2</sub> MXene that had degraded MB during five consecutive cycles. As presented in Table 2, the MB photocatalytic degradation effectiveness of MXenes declined a little after the 5th regeneration-reusability round. The authors highlighted that this may be due to slight structural deterioration and a decrease in functional groups on the surface of the MXenes [113]. These findings resonate with the conclusion drawn by Yao's team [103] and Zhang's group [85], which reported a minor drop in the degradation effectiveness after several regeneration cycles of spent MXenes. It may be ascribed to the loss of the number of functional groups on the surface of the spent MXenes as a result of reactions occurring during their usage. In such cases, additional regeneration steps should be developed to ensure recovery of photocatalyst properties for long-term applications.

In another stability study [87], expended MXene/Zn-MOF after recovery was regenerated using EtOH and acetone (50:50 v/v). It was discovered that the photocatalytic degradation efficiency of the MB pollutant was around 62 % in the 1st round, and there was no slight drop in effectiveness trend until the 5th round [87]. This excellent reusability and stability performance is consistent with that reported for g-C<sub>3</sub>N<sub>4</sub>/Ti<sub>3</sub>C<sub>2</sub>/TiO<sub>2</sub> by Diao's research group [179] towards photocatalytic degradation of RhB. The authors concluded that the catalyst amount lowered only minimally, which might be affected by the MXBM cleaning step with EtOH and H<sub>2</sub>O during the reusability test.

Vigneshwaran and co-workers also reported regeneration-reusability experiments. In this experiment, H<sub>2</sub>O and EtOH were used to achieve desorption of MB degradation by-products from the surface of Ti<sub>3</sub>C<sub>2</sub>Tx@Chitosan@g-C<sub>3</sub>N<sub>4</sub> and to restore the photocatalytic activity. The regeneration experiment was performed for 5 rounds, and there was no significant drop in the photocatalytic degradation capacity towards MB removal, indicating that Ti<sub>3</sub>C<sub>2</sub>Tx@Chitosan@g-C<sub>3</sub>N<sub>4</sub> has outstanding stability [130]. This is comparable to the reusability study of Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene towards the degradation of MG. In this study, the spent Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene nanocomposites were easily regenerated upon contact with DI H<sub>2</sub>O and EtOH. The regenerated Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene was then subjected to the next round of MG degradation. Notably, the photocatalytic degradation efficiency did not exhibit any significant loss until the third recycle, but a very slight decrease (about 12 %) in the photocatalytic degradation efficiency was found in the fifth cycle, and this was suggested to be due to incomplete desorption. However, the fifth time regenerated spent Nd-NiCo<sub>2</sub>O<sub>4</sub>/MXene still had >80 % photocatalytic degradation efficiency [86].

In another work [175], to assess the practical applicability, the reusability of the porous MXene/SWCNTs film for adsorbing MB was investigated. Using the same material, the adsorption and desorption procedures were carried out five times. The outcome implies that the p-MXene/SWCNTs film in its current state may find application as a stable and reusable electro-sorption film. Washing many times with H<sub>2</sub>O and EtOH allowed the MB to desorb. The authors [175] affirmed that, from the theoretical point of view, because of the incomplete desorption of the organic dye or partial oxidation of the film during recycling, the removal effectiveness would gradually decline. On the other hand, MB's

degradation effectiveness starts at 95 % and rises to 97.8 % in the third cycle. The intercalation of MB during the electro-sorption process may have been the cause of this event. This might encourage the MXenes to further delaminate, exposing more active sites and functional groups and enhancing the SSA, hence improving the degradation performance of MB. However, the fifth cycle yielded a degradation effectiveness of up to 95.2 %, suggesting that the developed MXene-based film might serve as a cost-effective and recyclable adsorbent for the removal of MB [175]. The above-mentioned reason is consistent with what was suggested by [150] in the recyclability study of MXene/PbCrO<sub>4</sub> towards the photocatalytic degradation of MB using DI H<sub>2</sub>O and EtOH as the eluent.

Notably, it can be seen from the regenerability and reusability data collected in Table 2 that the least number of times that MXene can be reused was 3, the highest number of times was 8, the average MXene pollutant degradation efficiency after n-th round of application is >70 %, and different eluents/regeneration agents that have been explored for regeneration of the MXene photocatalyst and desorbing dye pollutants from expended MXene include EtOH, and deionized/distilled H<sub>2</sub>O.

It is worth highlighting that real-case scenario applications would demand a 1–4 year constant stability over continuous operation of the catalyst, which still is not confirmed for MXene-based materials.

#### 14. Challenges and areas for future work

The review has identified main strategies for enhancing the photocatalytic effectiveness of MXBM for dye pollutant degradation in aqueous conditions. Notably, MXBM has proven to offer several advantages and become a real game-changer in environmental remediation. Nevertheless, based on the review of the literature, some problematic issues must be resolved before its pilot-scale implementation. The following are some of these issues and possible improvements that would be necessary to promote MXBM applicability in the research area, especially in the context of dye pollutant degradation.

Firstly, much of the research that is currently accessible on MXBM for dye pollutant photocatalytic degradation is conducted on a laboratory scale for model pollutant solutions. As a result, it is recommended that attempts on the industrial scale application level be considered in future studies. Secondly, the safe disposal of spent MXBM even after multiple reusability is another up-to-date topic that is worth looking into because the spent MXBM will still retain traces of dyes along with their degradation products and thus might still be potentially harmful. The presence of adsorbed degradation by-products also poses risks for the environment. Therefore, it is important to develop a method for the management of the spent MXBM.

Secondly, to significantly reduce MXenes' stability disadvantage, especially under humid and oxidative conditions, which is key for long-term potential use as well as pilot-scale implementation, we suggest that future researchers should look into surface modification or coating MXene with stabilizing materials like biopolymers, as this can create protective layers that inhibit oxidation and structural degradation. Additionally, elemental doping with N or Fe, and constructing heterostructures with stable semiconductors like g-C<sub>3</sub>N<sub>4</sub> can reinforce the MXene lattice, boost the stability, and modulate its band structure, thus enhancing durability as well as photocatalytic efficiency.

Thirdly, in the future, researchers may examine the circular economy potential of spent MXBM, as they may be recycled as raw materials for supercapacitors. Moreover, it is essential to perform a thorough life cycle analysis and ecotoxicological evaluation of MXBM [180].

Also, the relationship between the MXene structure and its photocatalytic activity can take many different forms, and many of the underlying processes are still unknown. Its photocatalytic properties and electrical activity, for instance, will also be influenced by surface alteration and quantum dot doping. There hasn't been much related study done yet on these aspects. Moreover, computational and density functional theory simulation studies can also be employed to predict the photocatalytic degradation mechanism of dye pollutants. Therefore,

**Table 3**

MXenes vs other common materials with respect to toxicity, cost-effectiveness, and sustainability as gathered from open literature [185–189].

Characteristics	MXenes	Graphene	Carbon nanotubes	Metal and metal oxide nanomaterial	Activated carbon
Toxicity	Moderate to low; depends on surface termination and metal core. In particular, MXenes produced using the HF etching method are considered to have a toxic effect because of HF involvement. However, there are now better synthesis methods with low toxicity been developed	Generally low; however, reduced graphene oxide can exhibit longer toxicity because of oxidative stress	Moderate; long carbon nanotubes may cause lung toxicity	Low to moderate; e.g., ZnO, TiO <sub>2</sub> may induce oxidative stress, phototoxicity, and genotoxicity. Moreover, heavy metal leaching is still a problem	Low; generally considered safe and non-toxic. Although, depending on the source
Cost-effectiveness	Moderate. Although still perceived as expensive, depending on the synthesis method. For example, the CVD method is considered costly	Moderate to high; depends on type. For instance, CVD graphene and graphene oxides are somewhat expensive	High production is complex and energy-intensive	Moderate; bulk synthesis is cheap and scalable	Moderate to high; Those made from biomass waste widely available are considered cheaper
Sustainability	Medium; relies on MAX phase precursors and HF or safer alternatives	Medium; synthetic routes use chemicals and high energy	Low; petroleum-based, difficult to recycle	Medium to high; depending on source and lifecycle	High; derived from renewable sources, recyclable

MXBM photocatalytic study in the future should include computational chemistry approaches. Also, future research projects can further explore the possibility of combining MXBM with different adsorbents of semiconductors. We believe this can deliver better degradation performance synergistically. We also identify a need for more studies on the effect of temperature and ionic species as part of technical operating parameters. On the other hand, each study on the application of MXBM for environmental remediation should include aspects of metal leaching into the treated environment.

Right now, large-scale production strategies are still far from ideal. Producing MXene in large amounts that are both of constant size and good quality will become essential in the coming years. Also, the use of hybrid technology like tribo-photocatalytic degradation can enhance the generation of radicals, which can, in turn, improve dye degradation and thus should be explored in the future. In addition, future researchers should look into the financial analysis aspect of this subject, from the production of MXBM to their usage in pollution remediation, as this is a crucial topic for engineers, investors, industries, and other stakeholders. Notably, for many materials, a size-dependent toxicity was found. Particles at nano- or sub-nano-scale can show higher toxicity compared to micro-scale particles of the same material [181–184]. The same studies on size-dependent toxicity should be performed in the future on different types of photocatalytic MXene-based materials. Finally, as summarized in Table 3, future research work will need to take note of the comparison of MXenes with other commonly used materials in terms of toxicity, cost-effectiveness, and sustainability for scalability advancement.

## 15. Conclusion

This review has critically explored the recent innovations and advances in the application of MXene-based materials (MXBM) for photocatalytic degradation of dye pollutants in wastewater and the experimentally confirmed effect of process parameters. A key innovation is the evaluation of photocatalytic degradation performance of MXenes integrated with other functional materials, resulting in hybrid materials with improved surface activity, efficient charge separation, and reduced recombination of electron-hole pairs. These properties significantly enhance photocatalytic efficiency, reusability, and ability to work under various operating conditions, crucial for sustainable water remediation. The findings show that MXBM typically achieves degradation efficiencies greater than 90 % within a treatment time of 60–80 min under optimal operating conditions of 10–100 mg/L photocatalyst dose and solution pH 2–7.5. The major reactive species in the photocatalytic reaction of MXBM were hydroxyl radicals ( $\cdot\text{OH}$ ) and superoxide radical anions ( $\cdot\text{O}_2^-$ ), which are accountable for the degradation of the dye

contaminants. The reusability of the materials is an important advantage of MXBM, as it can be reused up to 3–8 cycles while still maintaining degradation efficiency of >75 %, demonstrating both environmental friendliness and practical applicability. In conclusion, MXene-based photocatalysts represent a transformative class of emerging materials in photocatalytic water treatment, combining efficiency, tunability, and reusability. However, to fast-track their practical applications like for other existing materials that have been reviewed, future research should focus on intermediate products (by-products) formation, long-term toxicity assessments, cost analysis, sustainable scale-up strategies under real-case scenario conditions, and perform studies on particle size-dependent toxicity of different types of photocatalytic MXene-based materials. Bridging these gaps will help translate MXene from a promising laboratory remediation system to an impactful environmental material.

A still overlooked potential of MXBMs relates to their application in Advanced Reduction Processes (ARPs). Due to utilisation of hydrated electrons in specific conditions MXBMs could find applicability also in ARPs, which provide high effectiveness towards pollutants persistent to degradation via AOPs. Future research should include experiments in anoxic conditions (to favour hydrated electron driven reduction reactions), activation of reductants like sulphite or dithionate. MXBNs containing metals having affinity to hydrogen ( $\text{H}_2$ ), could catalyse reduction reactions by molecular hydrogen as well as hydrogen radical.

## CRedit authorship contribution statement

**Stephen Sunday Emmanuel:** Writing – review & editing, Writing – original draft, Project administration, Conceptualization. **Ademidun Adeola Adesibikan:** Writing – original draft, Visualization, Data curation. **Shepherd M. Tichapondwa:** Writing – original draft, Visualization, Data curation. **Manoj P. Rayaroth:** Writing – review & editing, Validation. **Francisco Bosca:** Writing – review & editing, Writing – original draft, Validation. **Maria Luisa Marin:** Writing – review & editing. **Bakhtiar Ali Samejo:** Writing – review & editing, Writing – original draft, Visualization. **Grzegorz Boczkaj:** Writing – review & editing, Writing – original draft, Validation, Supervision, Project administration, Conceptualization.

## Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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## Data availability

No primary research results, software, or code have been included, and no new data were generated or analyzed as part of this review.

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