

RESEARCH ARTICLE

Hybridization in an isolated population of blesbok and red hartebeest

Anna M. van Wyk^{1,2} | Erika Schulze³ | Kim Labuschagne¹ | Seeng Thamae¹ |
Antoinette Kotzé⁴ | Desiré Lee Dalton^{1,5} 

¹South African National Biodiversity Institute, Pretoria, South Africa

²Molecular Ecology and Evolution Program (MEEP), Department of Biochemistry, Genetics and Microbiology, University of Pretoria, Pretoria, South Africa

³Department of Economic, Small Business Development, Tourism and Environmental Affairs, Bloemfontein, South Africa

⁴Department of Genetics, University of the Free State, Bloemfontein, South Africa

⁵School of Health and Life Science, Teesside University, Middlesbrough, UK

Correspondence

Desiré Lee Dalton, South African National Biodiversity Institute, P.O. Box 754, Pretoria 0001, South Africa.
Email: d.dalton@tees.ac.uk

Abstract

Hybridization in antelope species has been widely reported in South African national parks and provincial reserves as well as on private land due to anthropogenic effects. In a closed management setting, hybridization may occur due to the crossbreeding of closely related species with unequal sex ratios, resulting in either sterile or fertile offspring. In this study, we used molecular techniques to evaluate the risk of anthropogenic hybridization between blesbok (*Damaliscus pygargus phillipsi*) and red hartebeest (*Alcelaphus buselaphus caama*) in an isolated group that purposely included the two species with unequal sex ratios (one red hartebeest male and 19 male and female blesbok). Genetic analysis based on microsatellites confirmed the presence of seven hybrid individuals. Mitochondrial analysis verified that hybridization occurred between blesbok females and the red hartebeest male. STRUCTURE and NEWHYBRIDS classified the hybrids as F1. It is suspected that the hybrid individuals were sterile as the males had undeveloped testes and only F1 hybrids were detected. Thus, the risk of hybridization between these two species may be limited in the wild. In captive settings, genetic monitoring should be included in management plans for blesbok and red hartebeest to ensure that the long-term consequences of wasted reproductive effort are limited.

KEYWORDS

blesbok, hybridization, microsatellites, red hartebeest, sterile offspring

TAXONOMY CLASSIFICATION

Conservation genetics

1 | INTRODUCTION

South Africa, as a signatory to the Convention on Biological Diversity (CBD) has implemented legislative measures for environmental protection. Establishment of national parks, game reserves and other protected areas of natural habitat is one of the principle

strategies used to conserve unique biodiversity. Approximately 9% of the country's land has been identified as protected areas. However, the private sector also plays a major role in the protection and management of antelopes. It is estimated that the private wildlife industry is 2.2 times greater than the state protected area network of the country (Els, 2017). Privately owned wildlife are

This is an open access article under the terms of the [Creative Commons Attribution](https://creativecommons.org/licenses/by/4.0/) License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2024 The Authors. *Ecology and Evolution* published by John Wiley & Sons Ltd.

used for recreational hunting, trophy hunting, wildlife meat production, breeding of wildlife and ecotourism (game viewing, walking safaris and photographic safaris) (Du Plessis, 1997). Antelopes in protected areas and on private land, face several threats including habitat degradation and destruction, disease, small population size and lastly hybridization.

Natural hybridization refers to the interbreeding of individuals from two genetically distinct species or populations (Arnold, 1997). This natural process is considered rare and may have substantial evolutionary significance, increasing adaptive capacity and species diversity (Abbott et al., 2013). However, others consider natural hybridization to be maladaptive due to reduced frequency and fitness of hybrids in comparison to parental genotypes (Moore & Price, 1993). In contrast, anthropogenic activities such as introduction of plants or animal (eg. translocation), habitat modification and/or fragmentation has dramatically increased the rates of hybridization worldwide (Stronen & Paquet, 2013). Anthropogenic hybridization is reported to have resulted in the extinction of several taxa and it has been suggested that conservation policies that are designed to reduce anthropogenic hybridization should be adopted (Allendorf et al., 2001). Hybridization between species or sub-species can have a variety of outcomes, including hybridization with introgression (Rhymer & Simberloff, 1996) where offspring are fertile. In these cases, there is a risk of extinction via introgression resulting in complete admixture (Allendorf et al., 2001; Rhymer et al., 1994; Rhymer & Simberloff, 1996). Hybridization without introgression can occur where hybrid offspring are sterile. Here, populations may decline due to wasted reproductive effort (Allendorf et al., 2001). However, the long-term consequences of wasted reproductive effort in small isolated populations are currently unknown. Hybridization is considered a real threat in South Africa as wildlife species are extensively translocated outside of their historic distribution ranges onto private land as part of wildlife management and commercial breeding (Spear & Chown, 2009). In closed management settings, multiple species may occur on the same property where there are few or no conspecific mates (Dalton et al., 2014; Grobler et al., 2011) which may contribute to an increased risk of hybridization. In addition, hybridization can occur due to loss of reproductive barriers between previously isolated evolutionary lineages (Green & Rothstein, 1998).

Red hartebeest (*Alcelaphus buselaphus caama*) is suspected to hybridise with blesbok (*Damaliscus pygargus phillipsi*), bontebok (*D. p. pygargus*) or tsessebe (*D. lunatus lunatus*) on private land (Venter & Child, 2016). Hybridization between red hartebeest and blesbok has been previously confirmed based on cytogenetic analyses (Robinson et al., 1991). The authors identified F1 hybrids with the number of chromosome being intermediate ($2n=39$) in comparison to parental species (red hartebeest $2n=40$ and blesbok, $2n=38$). One of the hybrid males was further reported as sterile based on azoospermia and lack of germ cells in seminiferous tubule cross-sections (Robinson et al., 1991). The blesbok is an abundant antelope species that is listed as least concern on the International Union for Conservation of Nature (IUCN) Red List of Threatened Species (Lloyd & David, 2008). Blesbok were historically distributed across

the Highveld grasslands of the Free State, Gauteng, north-western KwaZulu-Natal and parts of the Karoo in the Eastern and Northern Cape provinces of South Africa (Skinner & Chimimba, 2005). Red hartebeest occurs in dry, arid regions of Namibia, the Kalahari, southern Botswana, and is widespread in South Africa, with the exception of the Lowveld of Mpumalanga and Limpopo and Northern KwaZulu-Natal. The species is listed as least concern (IUCN SSC Antelope Specialist Group, 2017; Venter & Child, 2016). Currently, both species are found on private land and in national parks and provincial reserves within and outside of their natural distribution range (Power, 2014). In South Africa, the total number of mature red hartebeest and blesbok were estimated at 38,511 (2013–2014 counts) and 54,426 (2010–2016 counts) respectively with 14,849 red hartebeest and 17,235 blesbok found in protected areas across the country (Dalton et al., 2017; Venter & Child, 2016).

Thus far, identification of hybridization between the red hartebeest and blesbok based on genetic markers has not been reported. Molecular analysis has been identified as an accurate method to identify hybrids and detect introgression between taxa (Avisé & Hamrick, 1996). Various nuclear and uniparental molecular markers can be used including; mitochondrial DNA (mtDNA; Abbott et al., 2013), Y-chromosome markers (Petit et al., 2002), microsatellites (Haas & Payseur, 2011) and single nucleotide polymorphisms (SNPs) (Herrero-Medrano et al., 2013). Identification of anthropogenic hybridization based on genetic markers has been reported in a number of antelope species in South Africa including the blue (*Connochaetes taurinus*) and black wildebeest (*C. gnu*; Grobler et al., 2011, 2018), bontebok and blesbok (van Wyk et al., 2013) nyala (*Tragelaphus angasii*) and kudu (*T. strepsiceros*; Dalton et al., 2014). In this study, we used nuclear and mtDNA loci to detect hybridization in an isolated population established in 2013 where blesbok and red hartebeest were deliberately mixed with unequal sex ratios. Here, the aim was to determine if hybridization would occur between these two species and to determine the extent of hybridization. In addition, we explore the potential for hybrid fertility, in order to assess the threat that this phenomenon currently poses to both species.

2 | MATERIALS AND METHODS

2.1 | Study site and sample collection

Rustfontein Dam Nature Reserve (29°16'15" S, 26°37'1" E) is situated 10km south of the N8 national road between Bloemfontein and Thaba Nchu in the Free State Province, South Africa. The dam provides water for domestic, industrial and irrigation purposes. The founder group, introduced in 2008, consisted of five blesbok. One adult red hartebeest male was released into the herd in June 2013 and 19 blesbok were added from an adjacent reserve in July 2013 (sex of blesbok was not recorded). The mixing of two species with unequal sex ratios is a common occurrence in South Africa on private land and in this study was done deliberately in order to determine if hybridization would occur. In 2018, the entire group was culled

and tissue samples were taken from the remaining 26 individuals. Further details of the experimental population per year is indicated in [Figure S1](#). In addition, reference samples of red hartebeest ($n=13$) and blesbok ($n=20$) were obtained from the South African National Biodiversity Institute (SANBI) Biobank. The reference samples are part of the Biobank collection and represent various localities and populations within the current distribution ranges of both species.

2.2 | Genetic hybrid identification

2.2.1 | Microsatellite DNA analysis

DNA was extracted from tissue samples using the ZR Genomic DNA™ Tissue MiniPrep (Zymo Research Corporation) following the extraction protocol. Cross-species autosomal microsatellites markers ($n=21$, [Table S1](#)) developed for bovids (cattle, goat, sheep or blesbok) were used to genotype all individuals (Bhebhe et al., 1994; Bishop et al., 1994; Buchanan et al., 1994; Dalton et al., 2011; Ede et al., 1995; FAO, 2011; Massey & Georges, 1992; Sonstegard et al., 1997; Sunden et al., 1993; Toldo et al., 1993; Vaiman et al., 1992, 1994). Based allelic size range and fluorescent dye of individual primer pairs, sets of primers were combined and tested in multiplexed PCR reactions. Polymerase Chain Reaction (PCR) amplification was conducted in a 12.5 micro litre (μL) reaction volume consisting of Ampliqon Taq DNA polymerase RED (Lasec, Cape Town, SA), 5'-fluorescent dye labelled forward primer and reverse primer (0.5 micro Molar (μM) each) and 50 nano gram (ng) genomic DNA template. The conditions for PCR amplification were as follows: 5 min at 95°C denaturation, 35 cycles for 30s at 95°C, 30s at 50–62°C (refer to [Table S1](#) for annealing temperatures) and 30s at 72°C, followed by extension at 72°C for 10 min in a T100™ Thermal Cycler (Bio-Rad Laboratories, Inc. Hercules, CA, USA). PCR products were run against a Genescan™ 500 LIZ™ internal size standard on an ABI 3130 genetic analyser (Applied Biosystems, Inc., Foster City, CA, USA) and were genotyped using GeneMapper® v. 4.0 (Applied Biosystems, Inc., Foster City, CA, USA).

Analysis was performed for red hartebeest, blesbok and putative hybrids separately. In order to detect genotyping errors, allele drop-out and null alleles in the reference samples (red hartebeest and blesbok) MICRO-CHECKER (van Oosterhout et al., 2004) was used. In addition, in order to identify deviations from Hardy–Weinberg (HW) proportions of genotypes, analysis was conducted in Arlequin 3.5 (Excoffier et al., 2005; Excoffier & Lischer, 2010) using the following settings: Markov Chain length of 105 and 100,000 dememorization steps. Lastly, the presence or absence of linkage disequilibrium was verified using exact test described by Guo and Thompson (1992). A total of 100 initial conditions followed by ten permutations, based on the Sequential Bonferroni correction was used to adjust for multiple tests at a significant level of .05 (Rice, 1989). Mean number of alleles per locus (A), effective number of alleles (N_e), observed heterozygosities (H_o) and unbiased heterozygosities (H_z = expected heterozygosity adjusted for unequal sample sizes; Nei, 1987) were used

to estimate the level of genetic diversity (GenAIEx 6.5; Peakall & Smouse, 2006, 2012). In order to assess the level of genetic differentiation between the two species and F_{ST} -based hierarchical analysis of molecular variance (AMOVA; Excoffier et al., 1992) was calculated using Arlequin 3.5 (Excoffier et al., 2005; Excoffier & Lischer, 2010).

In order to identify hybrid individuals STRUCTURE 2.3.4 (Falush et al., 2007; Hubisz et al., 2009; Pritchard et al., 2000) and NEWHYBRIDS 1.1 (Anderson & Thompson, 2002) was used. The STRUCTURE analysis was performed using a model that assumes admixture, correlated allele frequencies and without prior population information for ten replicates each from $K=1-2$, with a run-length of 700,000 Markov Chain Monte Carlo repetitions, following a burn-in period of 200,000 iterations. NEWHYBRIDS determines the probability of each individual belonging to one of six categories including the two parental species, F1 and F2 hybrids, and backcrosses to each of the parental species. The NEWHYBRIDS analysis was performed in ten replicates runs for each prior (Jeffreys and Uniform) with 200,000 burn-in iterations and a total run length of 1,000,000. CLUMPAK (Cluster Markov Packager Across K ; Kopelman et al. (2015)) was used to graphically represent the runs for both STRUCTURE and NEWHYBRIDS. CLUMPAK obtains the membership coefficient matrices (Q-matrices) of replicate runs using CLUMPP (Jakobsson & Rosenberg, 2007), employing as a default, the LargeKGreedy algorithm with 2000 random input sequences. CLUMPAK employs DISTRUCT (Rosenberg, 2004) to graphically display the average run representing each cluster. For STRUCTURE, the average proportion of membership (q_i) of the sampled groups to the inferred clusters was assessed. Individuals were assigned to the inferred clusters using an threshold of $q_i > 0.9$ (Barilani et al., 2007). Thus, individuals with q_i values higher than 0.9 were identified as pure whereas individuals with values lower than 0.9 were classified as hybrid. Vähä and Primmer (2006) indicated that NEWHYBRIDS is less sensitive than STRUCTURE to differentiate between nonadmixed and admixed individuals, therefore a 0.7 probability of an individual belonging to a single category was selected as a threshold value (Costa et al., 2017; Gagnaire et al., 2009; Gunnell et al., 2008; Marie et al., 2011; Uwimana et al., 2012).

In order to maximise the accuracy of assignment, simulated genotypes were created using HYBRIDLAB (Nielsen et al., 2006). The genotypes of reference red hartebeest ($n=13$) and blesbok ($n=21$) were used to create 50 of each parental group and in turn to create the simulated hybrid genotypes. A dataset consisting of 300 individuals were produced consisting of 50 genotypes each belonging to red hartebeest, blesbok, F1 hybrids, F2 hybrids, backcross to red hartebeest and backcross to blesbok. The simulated dataset was analysed with STRUCTURE and NEWHYBRIDS as described above.

Pairwise relatedness was calculated between all individuals in the putative hybrid group, using the Lynch and Ritland estimator in GenAIEx version 6.5 (Peakall & Smouse, 2006). The relatedness coefficients (r) varies from 0 to 1, where .5 indicates that individuals are first-order relatives (parent and offspring or full-siblings) and a relatedness coefficient of .25 indicates second-order relatedness (half-siblings, grandparents, aunts/uncles or to niece/nephews).

2.2.2 | Mitochondrial DNA sequencing

The hybrid individuals identified through STRUCTURE and NEWHYBRIDS were further classified by means of cytochrome b (Cytb) amplification and sequence analyses to confirm maternity. Primers L14724 and H16498 (Hsieh et al., 2001) were used to target a 427 base pair (bp) region of the gene and amplification was conducted in a final reaction volume of 25 μ L using DreamTaq Green PCR Master Mix (Thermo Scientific), forward and reverse primers (10 pico mole (pmol) each), and template DNA (20 ng). The cycling conditions were as follows; 2 min at 94°C, 5 cycles for 30s at 94°C, 40s at 45°C, 1 min at 72°C, followed by 35 cycles for 30s at 94°C, 40s at 51°C, 1 min at 72°C and finally 72°C for 10 min in a T100™ Thermal Cycler. PCR purification was conducted in a final reaction volume of 22.5 μ L consisting of FastAP™ Thermosensitive Alkaline Phosphatase (Thermo Fisher Scientific Inc.) (1 U/ μ L), exonuclease I (20 U/ μ L) and PCR product (20 μ L) in a T100™ Thermal Cycler at the following conditions; 37°C for 15 min followed by 85°C for 15 min. After PCR purification, cycle sequencing was conducted using BigDye® Terminator v3.1 Cycle Sequencing Kit (Life Technologies, USA) in a final reaction volume of 10 μ L, including purified PCR product (5 μ L) and forward or reverse primer (10 pmol). Cycle sequencing was carried out in a T100™ Thermal Cycler at the following conditions; 94°C for 2 min followed by 40 cycles for 10s at 85°C, 10s at 50°C, 2 min 30s at 60°C. BigDye® X Terminator™ Purification Kit (Applied Biosystems, Inc., Foster City, CA, USA) was used for sequencing clean-up according to the manufacture's protocol. Analysis of the sequences was carried out on the ABI 3130 genetic analyser. Sequence-Analysis v. 4.0 (Applied Biosystems, Inc., Foster City, CA, USA) was used to visualise the sequences. Resulting sequence chromatograms were viewed and edited in the Chromas program embedded in MEGA5 (Tamura et al., 2011) prior to performing a BLAST nucleotide search (www.ncbi.nlm.nih.gov/blast). Maximum likelihood (ML) analyses were inferred in MEGA5 while the best fit model of sequence evolution was selected under the Akaike Information Criterion (AIC) in jModeltest (Posada, 2008). Nodal support for the Likelihood (ML) tree was assessed through 10,000 non-parametric bootstrap replications.

3 | RESULTS

3.1 | Genetic analysis and assignment testing

Amplification was obtained for all 21 cross-species autosomal microsatellites markers. However, three markers (SRCRSP8, BB05 and SPS113) were removed from the final data set as amplification was not achieved for more than 30% of the genotypes in the red hartebeest reference population. Null alleles were not detected in the red hartebeest population (Table S2). In addition, seven markers (BB10, BB08, OARCP26, BM2113, BB04, ETH10 and INRA128) in the blesbok population showed null allele frequencies higher than 0.2 at all four algorithms (Oosterhout, Chakraborty, Brookfield 1 and Brookfield 2). Significant deviations from HW

equilibrium were observed in one marker (BB10) in the blesbok following Bonferroni corrections. Significant linkage disequilibrium (LD) was only observed in the putative hybrid group which may be a consequence of it being admixed. Markers containing possible null alleles and deviations from HW equilibrium were removed from the final dataset. Thus, the final data set included 11 cross-species microsatellite.

Genetic diversity for each species and putative hybrids is summarised in Table 1. The genetic diversity in the red hartebeest was higher compared to the blesbok with the putative hybrid individuals being intermediate. Mean number of alleles per locus was 5.80, 2.80 and 4.00 while the average N_e was 3.58, 2.12 and 2.31 for red hartebeest, blesbok and putative hybrid individuals respectively. Observed heterozygosity was 0.55 (red hartebeest), 0.40 (blesbok) and 0.46 (putative hybrids) while H_e was 0.56 (red hartebeest), 0.46 (blesbok) and 0.53 (putative hybrids) (Table 1).

In the reference population, a total of 91 alleles were observed with 23 specific to blesbok, 60 to red hartebeest and eight shared (Table S3). Analysis by AMOVA (45%) and pairwise F_{ST} ($F_{ST}=0.45286$, $p < .001$) indicated a high proportion of genetic differentiation among populations. This was further supported by phylogenetic analysis where distinct clades for blesbok and red hartebeest were recovered with 100% bootstrap support (Figure 1). Hybrid individuals clustered with blesbok indicating that hybridization involved mating of blesbok females with the male red hartebeest.

Both STRUCTURE and NEWHYBRIDS showed similar results, identifying seven F1 hybrids and classifying the rest of the individuals (19) as blesbok (Figure 2; Table S4). In STRUCTURE, the average proportion of membership for both pure populations was $q_i > 0.989$. Using the criterion of $q_i > 0.90$ suggested by Barilani et al. (2007) to identify individuals as either pure or hybrid, 19 individuals were classified as blesbok with $q_i > 0.983$ and seven F1 hybrids were identified with $q_{iblesbok}$ ranging from 0.373 to 0.482 (Table 2, Table S4). NEWHYBRIDS indicated similar results for both Jeffreys and Uniform priors using a threshold value of $p > .7$ to assign individuals into categories (Figure 2, Table S4). All the reference samples were correctly assigned to the parental species ($p > .997$) and only F1 hybrids were identified ($p > .780$) from the putative hybrid population, the remainder of the individuals were identified as blesbok ($p > .999$) (Table 2; Table S4).

STRUCTURE identified all the simulated genotypes correctly except for one backcross to red hartebeest individual when applying a threshold value of $q_i > 0.90$ (Figure 3; Table S5). For NEWHYBRIDS the Jeffreys prior performed slightly better in assigning individuals to the correct category when using a threshold of $p > .7$ (Figure 3; Table S5). Analysing the simulated dataset using Jeffreys prior incorrectly categorised three F2 individual (Table S5) while applying the Uniform prior did not correctly identify eight F2 hybrids, two backcross to red hartebeest and three backcross to blesbok individuals (Table S5).

The average pairwise coefficient of relatedness in the putative hybrid group was -0.039 . Within the seven F1 hybrids identified

TABLE 1 Genetic diversity estimates for red hartebeest, blesbok and putative hybrid populations.

Population	Sample size (N)	Mean number of alleles per locus (A)	Effective number of alleles (N_e)	Observed heterozygosity (H_o)	Unbiased heterozygosity (H_z)
Red hartebeest	13	5.80	3.58	0.55	0.56
Blesbok	20	2.80	2.12	0.40	0.46
Putative hybrids	26	4.00	2.31	0.46	0.53

Abbreviations: A, Mean number of alleles per locus; H_o , Observed heterozygosity; H_z , Unbiased heterozygosity; N, sample size; N_e , effective number of alleles.

FIGURE 1 Maximum likelihood (ML) phylogenetic tree of blesbok, red hartebeest and hybrids (001, 002, 004, 010, 015 and 019) based on cytochrome b in combination with reference samples acquired from Genbank. All reference samples are prefixed with relevant Genbank accession numbers, while reference samples (generated in this study) are indicated with BL or RH. ML bootstrap support values given at the nodes. Red hartebeest and blesbok fall into two distinct clades and all hybrid animals have blesbok mitochondrial lineages.

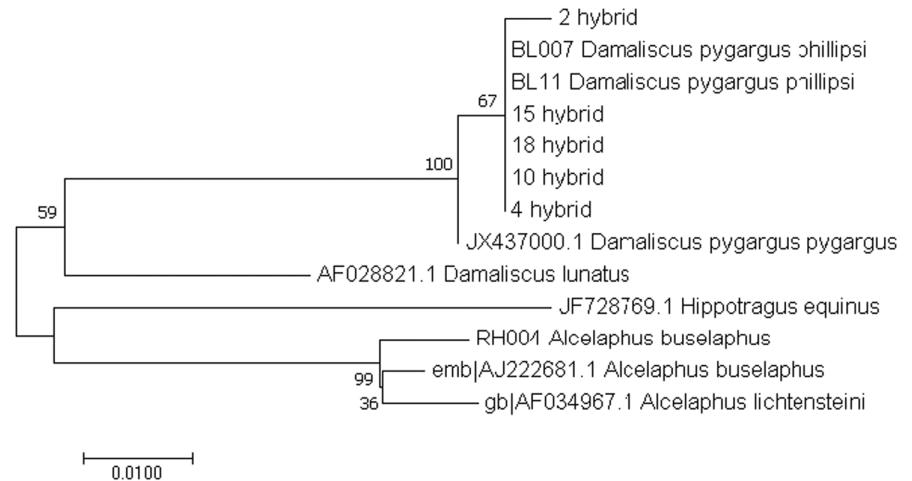
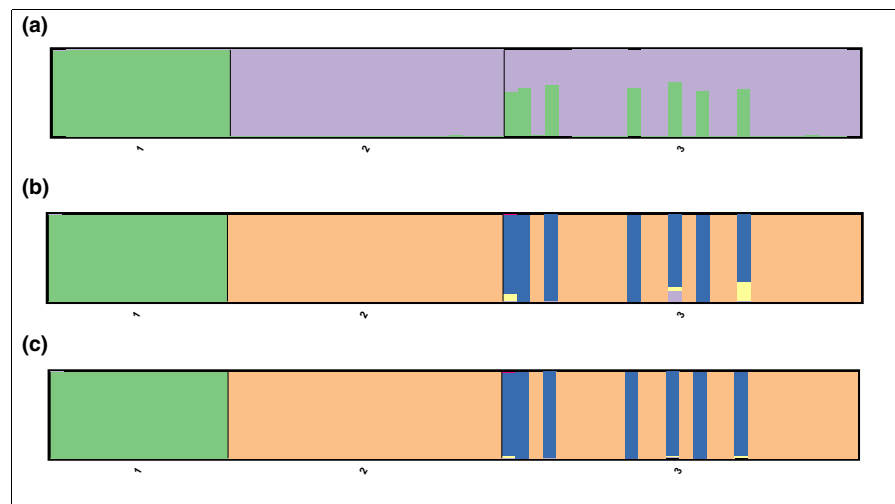


FIGURE 2 Genetic differentiation analysis between populations based on (a) STRUCTURE analysis (performed with $K=2$) and NEWHYBRIDS using (b) Jeffreys prior and (c) Uniform prior of 1 = red hartebeest, 2 = blesbok and 3 = putative hybrids.



using STRUCTURE and NEWHYBRIDS, first-order relatives (0.48–0.58) and second-order relatives (0.19–0.38) were identified indicating that they were either full-siblings or half-siblings. In all cases the parents could not be assigned indicating that they were culled prior to sampling.

4 | DISCUSSION

Here, we describe the first application of genetic tools to identify hybrids, due to mating between a male red hartebeest and

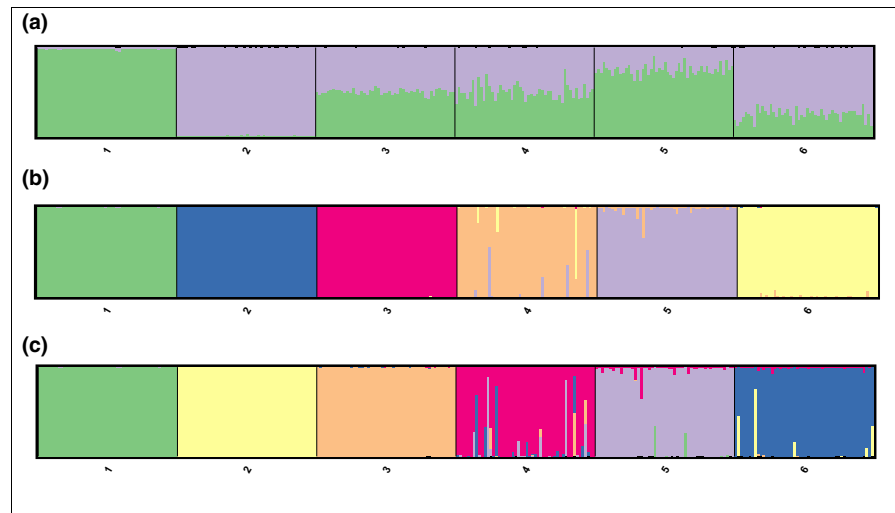
female blesbok. Hybridization is generally prevented due to prezygotic barriers where behaviour such as courtship displays and chemical signals play a fundamental role in species-specific mate recognition (Smadja & Butlin, 2009) and post-zygotic barriers where hybrids are either sterile or have reduced fertility (Coyne & Orr, 2004; Dobzhansky, 1937; Mayr, 1963). In this study, hybridization was forced with the shortage of conspecific mates, specifically for the red hartebeest. Hybridization events due to changes in population demographics such as skewed sex ratios is well described in other species including Grevy's zebra (*Equus grevyi*) and plains zebra (*E. burchelli*; Cordingley et al., 2009), European mink

TABLE 2 List of samples indicating assignment of animals.

Sample ID	Age estimate	Sex	Genetic result/assignment	STRUCTURE		NEWHYBRIDS			
				Red hartebeest (q_1)	Blesbok (q_1)	Blesbok Jeffreys prior (p)	F1 Jeffreys prior (p)	Blesbok Jeffreys prior (p)	F1 uniform prior (p)
1	Adult	Male	Hybrid	0.518	0.482	0.000	0.9071	0.0000	.9726
2	Adult	Male	Hybrid	0.562	0.438	0.000	1.0000	0.0000	.9999
3	Sub-adult	Male	Blesbok	0.017	0.983	1.0000	0.0000	1.0000	.0000
4	Adult	Female	Hybrid	0.5988	0.4013	0.000	.9939	0.0000	.9971
5	Adult	Female	Blesbok	0.006	0.994	1.0000	0.0000	1.0000	.0000
6	Adult	Male	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
7	Adult	Male	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
8	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
9	Adult	Male	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
10	Adult	Female	Hybrid	0.562	0.438	0.000	.9997	0.0000	.9998
11	Sub-adult	Male	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
12	Adult	Male	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
13	Adult	Male	Hybrid	0.6273	0.3727	0.000	.8358	0.0000	.9742
14	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
15	Adult	Female	Hybrid	0.5218	0.4783	0.000	1.0000	0.0000	1.0000
16	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
17	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
18	Adult	Female	Hybrid	0.547	0.453	0.000	.7802	0.0000	.9721
19	Adult	Female	Blesbok	0.003	0.997	1.0000	0.0000	.9999	.0000
20	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
21	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
22	Adult	Female	Blesbok	0.003	0.997	1.0000	0.0000	.9999	.0000
23	Adult	Female	Blesbok	0.013	0.987	1.0000	0.0000	.9999	.0000
24	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
25	Adult	Female	Blesbok	0.002	0.998	1.0000	0.0000	1.0000	.0000
26	Juvenile	Female	Blesbok	0.01	0.99	1.0000	0.0000	.9999	.0000

Note: Identified hybrids are indicated in bold text.

FIGURE 3 Proportional memberships of each genetic cluster and hybrid class for simulated individuals, estimated in (a) STRUCTURE performed with $K=2$ and NEWHYBRIDS executed with (b) Jeffreys prior and (c) Uniform prior for 1 = Red hartebeest, 2 = Blesbok, 3 = F1 hybrids, 4 = F2 hybrids, 5 = backcross to red hartebeest and 6 = backcross to blesbok.



(*Mustela lutreola*) and the polecat (*M. putorius*; Cabria et al., 2011) and fur seals (*Arctocephalus gazelle*, *A. tropicalis*, *A. forsteri*; Lancaster et al., 2006).

The seven hybrid individuals identified were considered sterile, mainly due to the presence of underdeveloped testes in the three hybrid males. In addition, Bayesian clustering analysis identified only F1 hybrids in the group further confirming sterility of the hybrids. In cases where hybrids are fertile, multigeneration hybrids and backcrosses would be expected. This finding is in line with a previous report of sterility in a red hartebeest x blesbok hybrid individual (Robinson et al., 1991). The species have different chromosome numbers which could lead to chromosome pairing complications during meiosis producing sterile offspring (Robinson et al., 1991). Experimental crossing of sibling species in mosquitos from the *Anopheles gambiae* complex has been reported to result in hybrid individuals with underdeveloped testes or hybrid individuals with normal testes morphologies and non-motile spermatozoa depending on the direction of the crossings (Liang & Sharakhov, 2019). Anthropogenic hybridization with few or sterile offspring is characterised by wasted reproductive effort where there is no gene exchange between parental species (Allendorf et al., 2001). In these cases, consequences of concern are more ecological rather than genetic, such as slow population growth (Senanan et al., 2004). Hybridization resulting in wasted reproductive effort can lead to extinction if a threatened species is involved in hybrid mating with a more common species. An example is the bull trout (*Salvelinus confluentus*) that almost disappeared from a stream in Montana (USA) after the introduction of brook trout (*Salvelinus fontinalis*). The brook trout out-numbered the bull trout which led to initiation of hybridization between the two species (Leary et al., 1993). The number of bull trout thus declined over time as the number of sterile offspring increased.

Anthropogenic hybridization occurs when there is removal of reproductive barriers between species that includes translocation of species outside their former distribution ranges leading to a breakdown in geographical isolation (Allendorf et al., 2001) and in

addition the failure to ensure the presence of enough conspecific mates of the translocated animals (Robinson et al., 1991). The red hartebeest and blesbok is no different with both species having different habitat preferences that should prevent hybridization under normal conditions. The red hartebeest is more tolerant of woodland areas and high grass (IUCN SSC Antelope Specialist Group, 2017), while the blesbok prefers open grasslands (Dalton et al., 2017). Thus far, all recorded hybridization events between the two species are due to translocated animals not having conspecific mates (Robinson et al., 1991). Robinson et al. (1991) reported two different hybridization events, in both instances one red hartebeest male survived after translocation of a subpopulation which lead to hybridization with female blesbok. Thus, the long-term negative consequences of hybridization between these two species may be limited. However, it is recommended that management plans should include details with regards to the availability of sufficient conspecific mates, where both species are managed sympatrically in order to reduce the risk of hybridization.

AUTHOR CONTRIBUTIONS

Anna M. van Wyk: Formal analysis (equal); investigation (equal); methodology (equal); writing – original draft (equal). **Erika Schulze:** Investigation (equal); project administration (equal); writing – review and editing (equal). **Kim Labuschagne:** Methodology (equal); resources (equal); writing – review and editing (equal). **Seeng Thamae:** Methodology (equal); resources (equal); writing – review and editing (equal). **Antoinette Kotzé:** Writing – review and editing (equal). **Desiré Lee Dalton:** Conceptualization (equal); formal analysis (equal); methodology (equal); writing – review and editing (equal).

ACKNOWLEDGEMENTS

The Free State Department of Economic, Small Business Development, Tourism and Environmental Affairs (FS DESTEA) is acknowledged for providing the opportunity to sample the culled animals.

CONFLICT OF INTEREST STATEMENT

The authors declare no conflicts of interest.

DATA AVAILABILITY STATEMENT

All results have been deposited in Dryad (<https://doi.org/10.5061/dryad.69p8cz98h>).

ORCID

Desiré Lee Dalton  <https://orcid.org/0000-0001-5975-6425>

REFERENCES

- Abbott, R., Albach, D., Ansell, S., Arntzen, J., Baird, S., Bierne, N., Boughman, J., Brelsford, A., Buerkle, C., Buggs, R., Butlin, R., Dieckmann, U., Eroukhanoff, F., Grill, A., Cahan, S., Hermansen, J., Hewitt, G., Hudson, A., Jiggins, C., ... Zinner, D. (2013). Hybridization and speciation. *Journal of Evolutionary Biology*, *26*, 229–246.
- Allendorf, F. W., Leary, R. F., Spruell, P., & Wenburg, J. K. (2001). The problems with hybrids: Setting conservation guidelines. *Trends in Ecology & Evolution*, *16*, 613–622.
- Anderson, E., & Thompson, E. (2002). A model-based method for identifying species hybrids using multilocus data. *Genetics*, *160*, 1217–1229.
- Arnold, M. L. (1997). *Natural hybridization and evolution*. Oxford University Press.
- Avise, J. C., & Hamrick, J. L. (1996). In J. Avise & J. Hamrick (Eds.), *Conservation genetics: Case histories from nature*. Chapman & Hall.
- Barilani, M., Sfougari, A., Giannakopoulos, A., Mucci, N., Tabarroni, C., & Randi, E. (2007). Detecting introgressive hybridization in rock partridge populations (*Alectoris graeca*) in Greece through Bayesian admixture analyses of multilocus genotypes. *Conservation Genetics*, *8*, 343–354.
- Bhebe, E., Kogi, J., Holder, D. A., Arevalo, E., Derr, J. N., Linn, R. A., Ruvuna, F., Davis, S. K., & Taylor, J. F. (1994). Caprine microsatellite dinucleotide repeat polymorphisms at the SR-CRSP-6, SR-CRSP-7, SR-CRSP-8, SR-CRSP-9 and SR-CRSP-10 loci. *Animal Genetics*, *25*, 203.
- Bishop, M. D., Kappes, S. M., Keele, J. W., Stone, R. T., Sunden, S. L. F., Hawkins, G. A., Toldo, S. S., Fries, R., Grosz, M. D., Yoo, J., & Beattie, C. W. (1994). A genetic linkage map for cattle. *Genetics*, *136*, 619–639.
- Buchanan, F. C., Adams, L. J., Littlejohn, R. P., Maddox, J. F., & Crawford, A. M. (1994). Determination of evolutionary relationships among sheep breeds using microsatellites. *Genomics*, *22*, 397–403.
- Cabria, M. T., Michaux, J. R., Gómez-Moliner, B. J., Skumatov, D., Maran, T., Fournier, P., López de Luzuriaga, J., & Zardoya, R. (2011). Bayesian analysis of hybridization and introgression between the endangered European mink (*Mustela lutreola*) and the polecat (*Mustela putorius*). *Molecular Ecology*, *20*, 1176–1190.
- Cordingley, J. E., Sundaresan, S. R., Fischhoff, I. R., Shapiro, B., Ruskey, J., & Rubenstein, D. I. (2009). Is the endangered Grevy's zebra threatened by hybridization? *Animal Conservation*, *12*, 505–513. <https://doi.org/10.1111/j.1469-1795.2009.00294.x>
- Costa, M. C., Oliveira, P. R. R., Davanc, P. V., De Camargo, C., Laganaro, N. M., Azeredo, R. A., Simpson, J., Silveira, L. F., & Francisco, M. R. (2017). Recovering the genetic identity of an extinct-in-the-wild species: The puzzling case of the Alagoas curassow. *PLoS One*, *12*(1), e0169636.
- Coyne, J., & Orr, H. (2004). *Speciation*. Sinauer Associates.
- Dalton, D., Parrini, F., Viljoen, P., Gaylard, A., & Peinke, D. (2017). *Damaliscus pygargus* ssp. *phillipsi*. *The IUCN Red List of Threatened Species*, 2017, e.T30209A50197495. <https://doi.org/10.2305/IUCN.UK.2017-2.RLTS.T30209A50197495.en>
- Dalton, D. L., Tordiffe, A., Luther, I., Duran, A., van Wyk, A. M., Brettschneider, H., Oosthuizen, A., Modiba, C., & Kotzé, A. (2014). Interspecific hybridization between greater kudu and nyala. *Genetica*, *142*, 265–271.
- Dalton, D. L., van Wyk, A. M., & Kotzé, A. (2011). Isolation and characterization of microsatellite loci in blesbok (*Damaliscus pygargus phillipsi*). *Conservation Genetics Resources*, *3*, 677–679.
- Dobzhansky, T. (1937). *Genetics and the origin of species*. Columbia University Press.
- Du Plessis, P. C. (1997). *Benuttingsvorme in die wildbedryf: 'n bestuursrekeningkundige evaluering* (PhD). North-West University.
- Ede, A. J., Pierson, C. A., & Crawford, A. M. (1995). Ovine microsatellites at the OarCP9, OarCP16, OarCP20, OarCP21, OarCP23 and OarCP26 loci. *Animal Genetics*, *26*, 129–130.
- Els, J. (2017). *Determining the economic significance of the lion industry in the private wildlife tourism sector* (Master's Dissertation). North-West University.
- Excoffier, L., Laval, G., & Schneider, S. (2005). Arlequin (version 3.0): An integrated software package for population genetics data analysis. *Evolutionary Bioinformatics Online*, *1*, 47–50.
- Excoffier, L., & Lischer, H. E. L. (2010). Arlequin suite ver 3.5: A new series of programs to perform population genetics analyses under Linux and windows. *Molecular Ecology Resources*, *10*, 564–567. <https://doi.org/10.1111/j.1755-0998.2010.02847.x>
- Excoffier, L., Smouse, P., & Quattro, J. (1992). Analysis of molecular variance inferred from metric distances among DNA haplotypes: Application to human mitochondrial DNA restriction data. *Genetics*, *131*, 479–491.
- Falush, D., Stephens, M., & Pritchard, J. K. (2007). Inference of population structure using multilocus genotype data: Dominant markers and null alleles. *Molecular Ecology Notes*, *7*, 574–578.
- FAO. (2011). *Molecular genetic characterization of animal genetic resources* (No. 9). FAO.
- Gagnaire, P. A., Albert, V., Jónsson, B., & Bernatchez, L. (2009). Natural selection influences AFLP intraspecific genetic variability and introgression patterns in Atlantic eels. *Molecular Ecology*, *18*, 1678–1691.
- Green, W. C. H., & Rothstein, A. (1998). Translocation, hybridization, and the endangered black-faced impala. *Conservation Biology*, *12*, 475–480.
- Grobler, J. P., Rushworth, I., Brink, J. S., Bloomer, P., Kotze, A., Reilly, B., & Vrahimis, S. (2011). Management of hybridization in an endemic species: Decision making in the face of imperfect information in the case of the black wildebeest-*Connochaetes gnou*. *European Journal of Wildlife Research*, *57*, 997–1006.
- Grobler, P., van Wyk, A. M., Dalton, D. L., van Vuuren, B. J., & Kotzé, A. (2018). Assessing introgressive hybridization between blue wildebeest (*Connochaetes taurinus*) and black wildebeest (*Connochaetes gnou*) from South Africa. *Conservation Genetics*, *19*, 981–993. <https://doi.org/10.1007/s10592-018-1071-x>
- Gunnell, K., Tada, M. K., Hawthorne, F. A., Keeley, E. R., & Ptacek, M. B. (2008). Geographic patterns of introgressive hybridization between native Yellowstone cutthroat trout (*Oncorhynchus clarkii bouvieri*) and introduced rainbow trout (*O. mykiss*) in the south fork of the Snake River watershed, Idaho. *Conservation Genetics*, *9*, 49–64.
- Guo, S. W., & Thompson, E. A. (1992). Performing the exact test of Hardy-Weinberg proportion for multiple alleles. *Biometrics*, *48*, 361–372.
- Haas, R., & Payseur, B. (2011). Multi-locus inference of population structure: A comparison between single nucleotide polymorphisms and microsatellites. *Heredity*, *106*, 158–171.
- Herrero-Medrano, J., Megens, H., Groenen, M., Ramis, G., Bosse, M., Perez-Enciso, M., & Crooijmans, R. (2013). Conservation genomic

- analysis of domestic and wild pig populations from the Iberian Peninsula. *BMC Genetics*, 14, 1–13.
- Hsieh, H., Chiang, H., Tsai, L., Lai, S., Huang, N., Linacre, A., & Lee, J. C. (2001). Cytochrome b gene for species identification of the conservation animals. *Forensic Science International*, 122, 7–18.
- Hubisz, M. J., Falush, D., Stephens, M., & Pritchard, J. K. (2009). Inferring weak population structure with the assistance of sample group information. *Molecular Ecology Resources*, 9, 1322–1332.
- IUCN SSC Antelope Specialist Group. (2017). *Alcelaphus buselaphus ssp. caama* (on-line). <https://doi.org/10.2305/IUCN.UK.2017-2.RLTS.T814A50181496.en>
- Jakobsson, M., & Rosenberg, N. A. (2007). CLUMPP: A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. *Bioinformatics*, 23, 1801–1806.
- Kopelman, N. M., Mayzel, J., Jakobsson, M., Rosenberg, N. A., & Mayrose, I. (2015). Clumpak: A program for identifying clustering modes and packaging population structure inferences across K. *Molecular Ecology Resources*, 15, 1179–1191.
- Lancaster, M., Gemmell, N., Negro, S., Goldsworthy, S., & Sunnucks, P. (2006). Ménage à trois on Macquarie Island: Hybridization among three species of fur seal (*Arctocephalus* spp.) following historical population extinction. *Molecular Ecology*, 15, 3681–3692.
- Leary, R. F., Allendorf, F. W., & Forbes, S. H. (1993). Conservation genetics of bull in the Columbia trout and Klamath River drainages. *Conservation Biology*, 7, 856–865.
- Liang, J., & Sharakhov, I. V. (2019). Premiotic and meiotic failures lead to hybrid male sterility in the *Anopheles gambiae* complex. *bioRxiv*. <http://biorxiv.org/content/early/2019/01/10/507889.abstract>
- Lloyd, P., & David, J. (2008). *Damaliscus pygargus*. IUCN Red List of Threatened Species. Version 2011.1 (on-line). <https://doi.org/10.2305/IUCN.UK.2008.RLTS.T30208A9530977.en>
- Marie, A. D., Bernatchez, L., & Garant, D. (2011). Empirical assessment of software efficiency and accuracy to detect introgression under variable stocking scenarios in brook charr (*Salvelinus fontinalis*). *Conservation Genetics*, 12, 1215–1227.
- Massey, J. M., & Georges, M. (1992). Genmark's approach to marker-assisted selection. *Animal Biotechnology*, 3(1), 95–109.
- Mayr, E. (1963). *Animal species and evolution*. Harvard University Press.
- Moore, W. S., & Price, J. T. (1993). Nature of selection in the northern flicker hybrid zone and its implications for speciation theory. In R. G. Harrison (Ed.), *Hybrid zones and the evolutionary process* (pp. 196–225). Oxford University Press.
- Nei, M. (1987). *Molecular evolutionary genetics*. Columbia University Press.
- Nielsen, E. E., Bach, L. A., & Kotlicki, P. (2006). HYBRIDLAB (version 1.0): A program for generating simulated hybrids from population samples. *Molecular Ecology Notes*, 6, 971–973.
- Peakall, R., & Smouse, P. E. (2006). GenAlEx 6: Genetic analysis in excel. Population genetic software for teaching and research. *Molecular Ecology Notes*, 6, 288–295. <https://doi.org/10.1111/j.1471-8286.2005.01155.x>
- Peakall, R., & Smouse, P. E. (2012). GenAlEx 6.5: Genetic analysis in excel. Population genetic software for teaching and research—An update. *Bioinformatics (Oxford, England)*, 28, 2537–2539.
- Petit, E., Balloux, F., & Excoffier, L. (2002). Mammalian population genetics: Why not Y? *Trends in Ecology & Evolution*, 17, 28–33.
- Posada, D. (2008). jModelTest: Phylogenetic model averaging. *Molecular Biology and Evolution*, 25, 1253–1256.
- Power, R. (2014). *The distribution and status of mammals in the North West Province*. North West Provincial Government, Mahikeng: Department of Economic Development, Environment, Conservation & Tourism.
- Pritchard, J. K., Stephens, M., & Donnelly, P. (2000). Inference of population structure using multilocus genotype data. *Genetics*, 155, 945–959.
- Rhymer, J. M., & Simberloff, D. (1996). Extinction by hybridization and introgression. *Annual Review of Ecology and Systematics*, 27, 83–109. <https://doi.org/10.1146/annurev.ecolsys.27.1.83>
- Rhymer, J. M., Williams, M. J., & Braun, M. J. (1994). Mitochondrial analysis of gene flow between New Zealand mallards (*Anas platyrhynchos*) and Grey ducks (*A. superciliosa*). *The Auk*, 111, 970–978.
- Rice, W. R. (1989). Analyzing tables of statistical tests. *Evolution*, 43, 223–225.
- Robinson, T. J., Morris, D. J., & Fairall, N. (1991). Interspecific hybridization in the bovidae: Sterility of *Alcelaphus buselaphus* × *Damaliscus dorcas* F1 progeny. *Biological Conservation*, 58, 345–356.
- Rosenberg, N. A. (2004). DISTRUCT: A program for the graphical display of population structure. *Molecular Ecology Notes*, 4, 137–138.
- Senanan, W., Kapuscinski, A., Na-Nakorn, U., & Miller, L. (2004). Genetic impacts of hybrid catfish farming (*Clarias macrocephalus* × *C. gariepinus*) on native catfish populations in central Thailand. *Aquaculture*, 235, 167–184.
- Skinner, J. D., & Chimimba, C. T. (2005). *The mammals of the southern African subregion* (3rd ed.). Cambridge University Press.
- Smadja, C., & Butlin, R. (2009). On the scent of speciation: The chemosensory system and its role in premating isolation. *Heredity*, 102, 77–97.
- Sonstegard, T. S., de Leon, F. A. P., Beattie, C. W., & Kappes, S. M. (1997). A chromosome-specific microdissected library increases marker density on bovine chromosome 1 [1]. *Genome Research*, 7, 76–80.
- Spear, D., & Chown, S. L. (2009). Non-indigenous ungulates as a threat to biodiversity. *Journal of Zoology*, 279(1), 1–17.
- Stronen, A. V., & Paquet, P. C. (2013). Perspectives on the conservation of wild hybrids. *Biological Conservation*, 167, 390–395.
- Sunden, S. F. L., Stone, R. T., Bishop, M. D., Kappes, S. M., Keele, J. W., & Beattie, C. W. (1993). A highly polymorphic bovine microsatellite locus: BM2113. *Animal Genetics*, 24, 69.
- Tamura, K., Peterson, D., Peterson, N., Stecher, G., Nei, M., & Kumar, S. (2011). MEGA5: Molecular evolutionary genetics analysis using maximum likelihood, evolutionary distance and maximum parsimony methods. *Molecular Biology and Evolution*, 28, 2731–2739.
- Toldo, S. S., Fries, R., Steffen, P., Neiberger, H. L., Barendse, W., Womack, J. E., Hetzel, D. J. S., & Stranzinger, G. (1993). Physically mapped, cosmid-derived microsatellite markers as anchor loci on bovine chromosomes. *Mammalian Genome*, 4, 720–727.
- Uwimana, B., D'Andrea, L., Felber, F., Hooftman, D. A. P., den Nijs, H. C. M., Smulders, M. J. M., Visser, R. G. F., & van de Wiel, C. C. M. (2012). A Bayesian analysis of gene flow from crops to their wild relatives: Cultivated (*Lactuca sativa* L.) and prickly lettuce (*L. serriola* L.) and the recent expansion of *L. serriola* in Europe. *Molecular Ecology*, 21, 2640–2654.
- Vähä, J. P., & Primmer, C. R. (2006). Efficiency of model-based Bayesian methods for detecting hybrid individuals under different hybridization scenarios and with different numbers of loci. *Molecular Ecology*, 15, 63–72.
- Vaiman, D., Mercier, D., Moazami-Goudarzi, K., Eggen, A., Ciampolini, R., Lépingle, A., Velmalá, R., Kaukinen, J., Varvio, S. L., Martin, P., Levéziel, H., & Guérin, G. (1994). A set of 99 cattle microsatellites: Characterization, synteny mapping, and polymorphism. *Mammalian Genome*, 5, 288–297.
- Vaiman, D., Osta, R., Mercier, D., Grohs, C., & Levéziel, H. (1992). Characterization of five new bovine dinucleotide repeats. *Animal Genetics*, 23, 537–541.
- van Oosterhout, C., Hutchinson, W. F., Wills, D. P. M., & Shipley, P. (2004). MICRO-CHECKER: Software for identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology Notes*, 4, 535–538.
- van Wyk, A. M., Kotzé, A., Randi, E., & Dalton, D. L. (2013). A hybrid dilemma: A molecular investigation of south African bontebok (*Damaliscus pygargus pygargus*) and blesbok (*Damaliscus pygargus phillipsi*). *Conservation Genetics*, 14, 589–599.

Venter, J., & Child, M. (2016). A conservation assessment of *Alcelaphus buselaphus caama*. In M. Child, L. Roxburgh, E. Do Linh San, D. Raimondo, & H. Davies-Mostert (Eds.), *The red list of mammals of South Africa, Swaziland and Lesotho*. South African National Biodiversity Institute and Endangered Wildlife Trust.

SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

How to cite this article: van Wyk, A. M., Schulze, E., Labuschagne, K., Thamae, S., Kotzé, A., & Dalton, D. L. (2024). Hybridization in an isolated population of blesbok and red hartebeest. *Ecology and Evolution*, 14, e11194. <https://doi.org/10.1002/ece3.11194>