

Habitat Selection of the Black-shouldered Kite (*Elanus caeruleus*) in Agroecosystems of Swaziland

KATHARINE A. HOWARD¹, THOMAS W. SCHWERTNER¹, JAMES P. MUIR^{1,2}, and ARA
MONADJEM^{3,4}

¹Department of Wildlife, Sustainability, and Ecosystem Sciences, Tarleton State University,
Stephenville, TX USA. ²Texas A&M AgriLife Research, Stephenville, TX USA. ³All Out Africa
Research Unit, Department of Biological Sciences, University of Swaziland, Private Bag 4, Kwaluseni,
Swaziland, ⁴Mammal Research Institute, Department of Zoology & Entomology, University of Pretoria,
Private Bag 20, Hatfield 0028, Pretoria, South Africa.

Abstract

A decline in many African raptor populations may contribute to loss of ecosystem services such as pest control. The Black-shouldered Kite (*Elanus caeruleus*) stands out as an exception, experiencing range expansions due to its proclivity for agricultural landscapes. We surveyed roads throughout Swaziland agroecosystems for Black-shouldered Kites and looked at habitat selection in a used versus available framework. We found that kites selected for savannahs and tall herbaceous vegetation but did not show preferences for or against land use types. We also found a significant relationship with kite presence and low tree densities in the immediate vicinity of a perch site. Singles, pairs and fledgling kites were observed in agricultural landscape, suggesting that although they are not selecting for these areas in proportion to its availability, they are foraging in it. This may be a boon for farmers who implement ecologically based rodent management programs.

KEYWORDS: Agroecosystems, Black-shouldered Kite, ecologically based rodent management, *Elanus caeruleus*, habitat selection, pest-control, Swaziland

Introduction

Loss of biodiversity impacts ecosystem function, and decreases in functional diversity can negatively affect ecosystem resilience (Allen et al. 2005). A system's resilience is a major factor in its ability to withstand changes such as climate, as well as continue to provide ecosystem services (Folke et al. 2002). Continued conversion from traditional subsistence farming to intense cash crop production in southern Africa may threaten many of the ecosystem services that are important to small and large scale farmers (Khumalo et al. 2012). One area of concern is a loss of pest control services due to simplification of both the pest and predator communities (Tschardt et al. 2005).

Many studies that address avian predator ability to control rodent pests agree that raptors alone cannot fully limit rodent populations (Sinclair et al. 1990, Van Gulck et al. 1998, Wolff et al. 1999, Vibe–Petersen et al. 2006, Paz et al. 2013). Raptors do, however, appear capable of dampening rodent population oscillations (Korpimaki and Norrdahl 1991). Increasing predation risk by raptors may also alter rodent foraging behaviour, potentially reducing crop depredation (Abramsky et al. 2002, Mohr et al. 2003, Preisser et al. 2005, Juliana et al. 2011). Factors that contribute to raptors' ability to dampen rodent populations include rapid numerical responses to increased rodent density and low propensity for raptor populations to be limited by territoriality (Korpimaki and Norrdahl

Studies assessing how agriculture affects functional groups across avian taxa in South Africa show that upper trophic level species' populations, including raptors, are declining, which could in turn impact pest control services (Child et al. 2009, Cumming and Child 2009). Throughout Africa, these declines in raptor populations (Sorley and Andersen 1994, Thiollay 2007, Ogada and Keesing 2010) have been attributed largely to human modification of landscapes that increasingly restrict raptors to protected areas (Buchanan et al. 2009, Coetzee et al. 2009). One species stands out as an exception: the nominate form of Black-shouldered Kite (*Elanus caeruleus* [hereafter, BSKI]; Brandl et al. 1985, Herremans and Herremans–Tonnoeyr 2001, Thiollay 2001, Monadjem and Rasmussen 2008).

The BSKI is a small, diurnal raptor and small mammal specialist (Tarboton 1978, Mendelsohn 1981). While this species by itself is not a substitute for raptors such as buzzards and eagles, it does exhibit those characteristics described by Korpimaki and Norrdahl (1989) that might enable it to impact rodent populations. BSKI are opportunistic specialists that respond to increases in available prey (Parejo and Aviles 2001). They are frequently nomadic and will congregate in communal

roosts in the non–breeding season, thus enabling them to appear in high concentrations during rodent outbreaks (Mendelsohn 1983, 1988). They also spend the bulk of their foraging efforts perch hunting (Mendelsohn and Jaksic 1989); land managers could take advantage of this behaviour by installing artificial perches to encourage BSKI presence (Kay et al. 1994, Wolff et al. 1999, Sheffield et al. 2001). These characteristics make BSKI uniquely qualified to provide agroecosystems with important pest control services.

In Africa, there is increasing interest in using ecologically based rodent management (EBRM) systems to deal with pests, particularly the Natal Multimammate Mouse (*Mastomys natalensis*) (Makundi and Massawe 2011). The Natal Multimammate Mouse is the top crop damage-causing rodent in Swaziland (Monadjem et al. 2011, Takele et al. 2011), where it is a significant pest in sugarcane fields and experiences population irruptions that can greatly increase economic losses (Stenseth et al. 2003). Sugarcane plantations tend to have over–simplified small mammal communities, and a recommendation to address this imbalance is that intensive agricultural systems integrate natural habitat features to counteract homogenization (Hurst et al. 2013, 2014). In conjunction with successful EBRM methods such as community trapping (Taylor et al. 2012), a mosaic of natural habitat within agroecosystems may also promote the presence of BSKI as a population control agent by providing roosting and breeding sites. BSKI are known predators of Natal Multimammate Mice in fallow agricultural fields (Van Gulck et al. 1998), as well as many other rodent species. Implementing EBRM with BSKI, however, presupposes a solid understanding of habitat utilization by this raptor. To address this issue our study examined habitat utilization by BSKI in Swaziland agroecosystems at multiple scales to identify those habitat characteristics positively associated with BSKI presence.

Methods

Study Area

Our study was conducted from 22 May 2014 to 3 August 2014 in agroecosystems throughout eastern Swaziland. Swaziland is a landlocked country in southeastern Africa, covering 17 565 km² with elevation ranging from 21–1 862 m above sea level (Goudie and Price Williams 1983). The country consists of four major geographical regions: the Highveld to the west, Middleveld in the center, the Lowveld to the east, and the Lubombo Mountains in the far east (Goudie and Price Williams 1983). Average annual temperatures range from 15° to 23° C, with a distinct rainy season from October to March (Goudie and Price Williams 1983). Natural vegetation types include grassland, open, acacia, and broadleaf savannah, as well as mixed bushveld. In 2013, 71% of the nation's land was devoted to agriculture, consisting largely of sugar cane, maize, and citrus (WB 2015).

Field Methods

We established a 440-km route along primary and secondary roads (Fig.1) and systematically surveyed one-third of the route with two observers for 4–5 hours between 08:00 and 17:00 at a cruising speed of approximately 80 km/h. When either observer sighted a BSKI, we took GPS coordinates of the location. We also selected a number of random points generated with ArcMap 10.1 (Environmental Systems Research Institute, Inc., Redlands, CA USA) along the available study route to compare use versus availability (Manly et al. 2002). We went to each of these points and identified the nearest available perch location (power line or tree) to measure the same variables as those taken at bird-sight locations. We considered a point independent at an 892-m radius, thus if at any time a bird was seen within that distance of one of the available sites, we

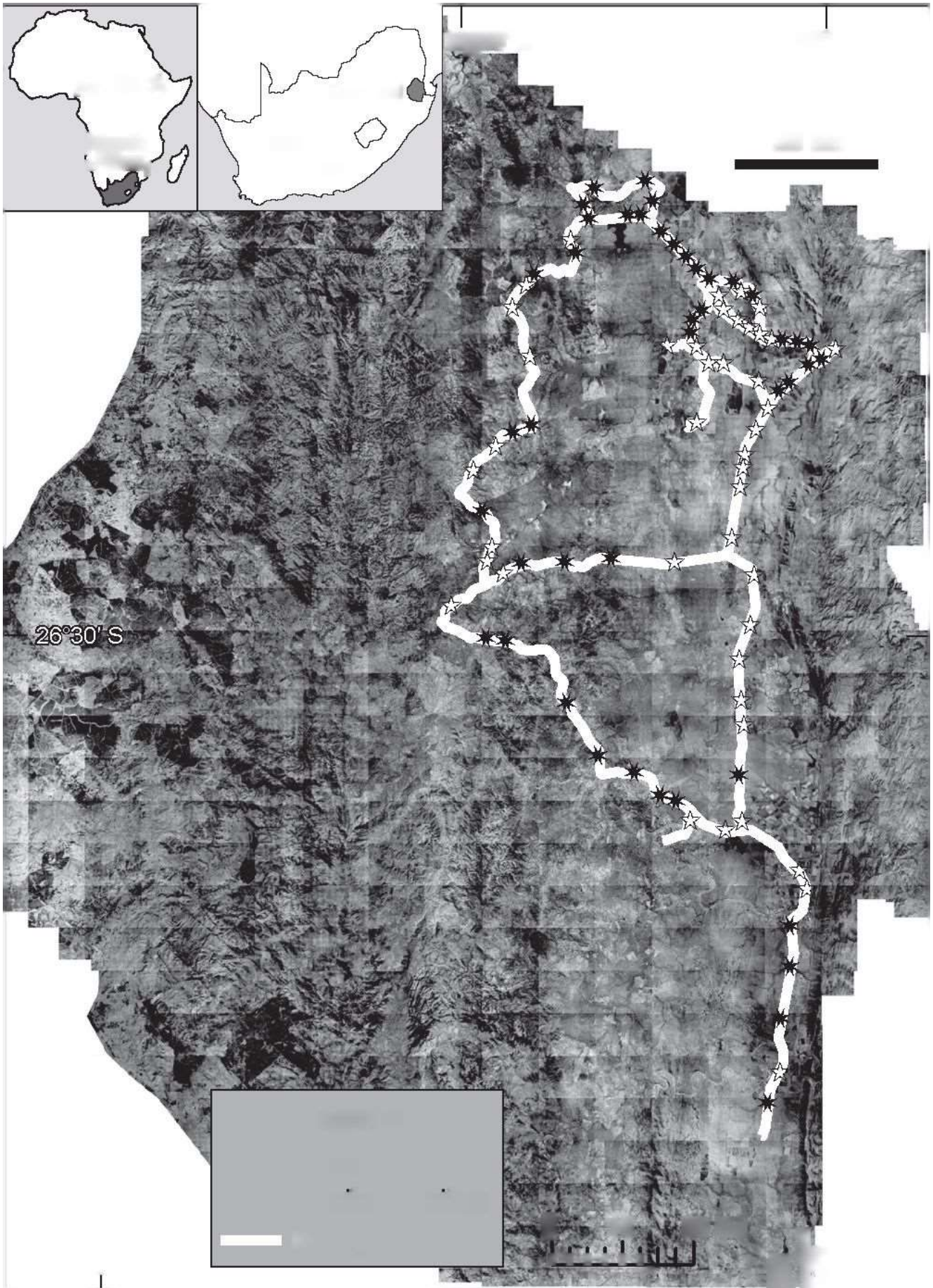


Figure 1. Map of Swaziland showing 440-km study route and used/available sites surveyed.

removed that site from the analysis. We chose this distance based on previously reported average home range size of 2.5 km² for BSKI in South Africa (Mendelsohn 1981).

We measured habitat variables for each of the sites including distance to nearest tree in each of the four cardinal directions up to 400 m using a rangefinder, whether the perch used was a power line or tree, and if the immediate area (30-m radius) was dominated by short (≤ 15 cm) or tall (> 15 cm) herbaceous vegetation. When we observed more than one bird at any given site or observed a bird at a known used site, we used the data from the earliest observation. We also categorized sites into habitat types using aerial imagery in ArcMap 10.1. No pre-existing land use classification was available, so we estimated proportions of the study area consisting of sugarcane, developed areas (urban areas, small villages, and settlements), and undeveloped areas (protected reserves and pastures) using ArcMap 10.1. A site was assigned to each category based on the centroid location. We additionally classified sites into land cover categories of bushveld, savannah, or cultivated lands according to the most dominant category.

In addition to recording bird sightings and measuring habitat variables, we made anecdotal observations of bird behaviour. We noted whether birds were seen in pairs, whether they were interacting with other bird species, and if a bird was a juvenile or adult based on the freshness of the reddish juvenile plumage on the chest and buffy tips to the flight feathers (Bustemante 1993).

Statistical Analysis

We analysed all categorical covariates using a two-way chi-square test as a design I with estimated proportions of available resource units (Manly et al. 2002). We analysed our only continuous covariate, average distance to nearest tree, with a Student's t-test (Zar 1984).

Results

We recorded a total of 88 bird sightings at 43 used sites; in some cases birds were seen within the 892-m radius of a previously recorded site, and occasionally multiple birds were seen at a single location. We compared these to 43 random sites that we considered available. Of the birds sited, 95% were observed perching on power lines. Black-shouldered Kites used savannah more often in proportion to its availability ($\chi^2 = 7.37, p = 0.025$; Table 1) and selected tall herbaceous ground cover over short cover ($\chi^2 = 10.47, p = 0.001$; Table 2). They did not select or avoid any other habitat type in relation to its availability. Average distance to nearest tree did not differ among used and available sites at the 400-m distance. However, when we removed distances greater than 100 m from the analysis, the mean distance to nearest tree was greater at used sites than at available sites ($t_{47} = -1.71, p = 0.047$; Fig. 2).

We observed seven pairs over the study period, including two adults with two fledglings. The fledglings were first observed on 15 July at a site that typically had a single hunting adult, and when the fledglings were recorded they were accompanied by two adults. These four birds were resighted together three more times, the last observation on 1 August. Another pair was seen actively defending their territory against a Lizard Buzzard (*Kaupifalco monogrammicus*) on 30 May, and a lone juvenile was seen adjacent to this site on 23 May.

Discussion

Black-shouldered Kites were conspicuous when perched on power lines along roads. Consequently, we may have had a greater detection probability for these birds as opposed to those in trees. That they selected for savannah confirms what we already know about this species' preferences from previous studies (Mendelsohn and Jaksic 1989, Thiollay 2001, Seavy and Apodaca 2002). Studies

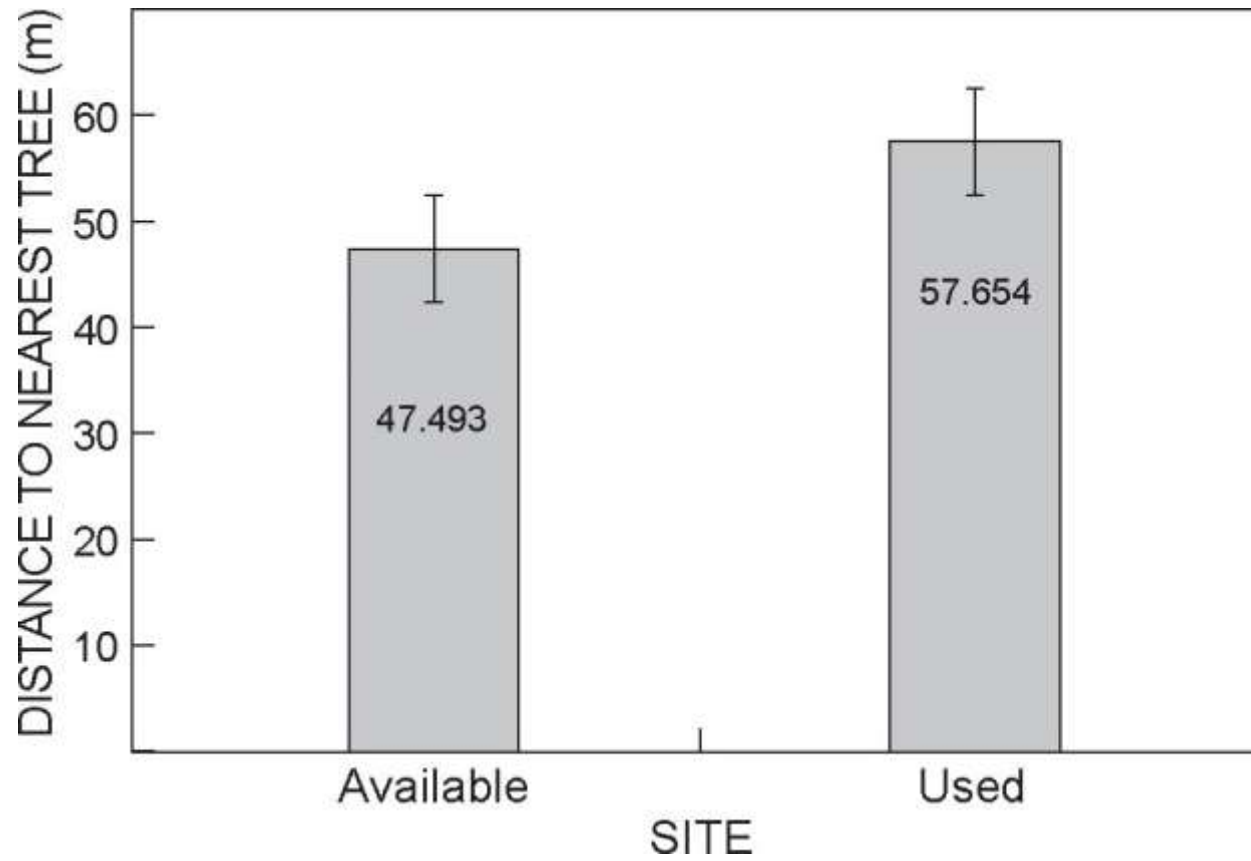


Figure 2. Mean distance to nearest tree up to 100 m (n = 50) with standard error bars at used and available sites in Swaziland.

Land cover	Available	$\hat{\pi}$	Used	$\hat{\delta}$	Selection ratio	Standardized ratio	Confidence limits		
							SE (selection ratio)	Lower	Upper
Bushveld	17	0.395	11	0.256	0.647	0.178	0.179	0.199	1.095
Savannah	9	0.209	21	0.488	2.333	0.643	0.708	0.562	4.104
Cultivated	17	0.395	11	0.256	0.647	0.178	0.179	0.199	1.095
Total	43	1	43	1	3.627	1			

Table 1. Design I with estimated proportions of available resource units (Manly et al. 2002) for land cover categories at used and available sites in Swaziland, where π represents a ratio of available to total units, and δ represents a ratio of used to total units.

Herb- aceous cover	Available	π	$\hat{\pi}$	U	\hat{U}	$\hat{\delta}$	Selection ratio	Standardized ratio	SE (selection ratio)	Confidence limits	
										Lower	Upper
Short (≤ 15 cm)	29	0.674	14	0.326	0.483	0.189	0.100	0.234	0.732		
Tall (> 15 cm)	14	0.326	29	0.674	2.071	0.811	0.459	0.924	3.219		
Total	43	1	43	1	3	1					

Table 2. Design I with estimated proportions of available resource units (Manly et al. 2002)

for land cover categories at used and available sites in Swaziland, where π represents a ratio of available to total units, and $\hat{\delta}$ represents a ratio of used to total units.

that examined foraging selection in other raptor species show that shorter vegetation is preferred over taller vegetation (Toland 1987, Garcia et al. 2006, Swolgaard et al. 2008); however, our results indicated kites select for tall herbaceous cover. In our study area, birds were sighted on road verges where there was a distinct contrast between areas that had been burned or mowed and areas that were not managed, hence our classification of short and tall herbaceous vegetation. In other systems, short vegetation presumably allows easier access to prey, while our study areas classified as short were typically nearly bare, and were likely avoided entirely by rodents (Monadjem 1999). Our decision to truncate the average distance to nearest tree revealed a significant relationship with lower tree densities, which corroborated previous findings (Balbontin et al. 2008). Canopy cover presumably decreases visibility for an aerial hunting predator, whereas trees more than 100 m from a perch site seem unlikely to have any effect.

We found no evidence that BSKI selected for or against land used for agriculture, although 22 of the 43 used sites contained agriculture activity and 11 birds were observed directly in agricultural areas. A recent study in northern Tanzania found BSKI nearly absent from cultivated areas adjacent to Serengeti National Park (Byrom et al. 2014); this contradicts our findings as well as other studies that have noted BSKI in agricultural areas (Brandl et al. 1985, Balbontin et al. 2008). The presence or absence of suitable perching sites may limit BSKI use of agricultural areas (Van Gulck et al. 1998). In areas where BSKI presence is desired, installation of perches may encourage occupation, particularly power line-like structures (Wolff et al. 1999).

Our behavioural observations indicated at least some birds remained in pairs on territories during the dry season. Black-shouldered Kites are thought to breed opportunistically in relation to prey availability, which typically fluctuates with precipitation (Mendelsohn 1984, Monadjem and Perrin 2003). The presence of adults indicated that these were recently fledged juveniles, and reports

of post-fledgling dependence period vary from an average of 34–81 days (Bustemante 1993, Mendelsohn 1981). Both of these sites, along with all but one of the other pairs observed, contained substantial amounts of sugarcane. Breeding pairs consume larger quantities of prey (Slotow and Perrin 1992) and the high observed incidence of pairs in agricultural settings indicate that birds may be keying in on the abundance of prey in and around sugarcane fields.

Our results confirm that BSKI do forage in agricultural settings in Swaziland, but our inability to individually mark animals raised many questions. Where they nest and how their home ranges vary in different habitat types are examples of questions that radio-marked animals may help answer. For managers attempting to encourage BSKI presence for rodent management purposes, power line-like perches, some moderate amount of herbaceous vegetation, and low tree densities are the only recommendations available without further investigation.

Acknowledgements

We thank Mbuluzi Game Reserve, All Out Africa, and the Savannah Research Station and their employees for providing accommodations and assistance during the field season. Thanks to Royal Swaziland Sugar Corporation for allowing access to sugarcane plantations. Much thanks to Lorena Mays for field assistance. Thanks to Christopher Briggs, Heather Mathewson, and Tal Fineberg for technical advice. Funding provided by United States Department of Agriculture National Institute for Food and Agriculture and Tarleton State University.

References

- Abramsky Z, Rosenzweig ML, Subach A. 2002. The costs of apprehensive foraging. *Ecology* 83: 1330–1340.
- Allen CR, Gunderson L, Johnson AR. 2005. The use of discontinuities and functional groups to assess relative resilience in complex systems. *Ecosystems* 8: 958–966.
- Balbontin, J, Negro JJ, Sarasola JH, Ferrero JJ, Rivera D. 2008. Land-use changes may explain the recent range expansion of the Black-shouldered kite *Elanus caeruleus* in southern Europe. *Ibis* 150: 707–716.
- Brandl R, Utschick H, Schmidtke K. 1985. Raptors and land-use systems in southern Africa. *African Journal of Ecology* 23: 11–20.
- Buchanan GM, Donald PF, Fishpool LDC, Arinaitwe JA, Balman M, Mayaux P. 2009. An assessment of land cover and threats in important bird areas of Africa. *Bird Conservation International* 19: 49–61.
- Bustemante J. 1993. The post-fledgling dependence period of the Black-shouldered Kite. *Journal of Raptor Research* 27: 185–190.
- Byrom AE, Craft ME, Durant SM, Nkwabi AJK, Metzger K, Hampson K, Mduma SAR, Forrester GJ, Ruscoe WA, Reed DN, Bukombe J, Mchetto J, Sinclair ARE. 2014. Episodic outbreaks of small mammals influence predator community dynamics in an east African savannah ecosystem. *Oikos* 123: 1014–1024.
- Child MF, Cumming GS, Amono T. 2009. Assessing the broad-scale impact of agriculturally transformed and protected area landscapes on avian taxonomic and functional richness. *Biological Conservation* 142: 2593–2601.
- Coetzee BWT, Robertson MP, Erasmus BFN, van Rensburg BJ, Thuiller W. 2009. Ensemble models predict important bird areas in southern Africa will become less effective for

conserving endemic birds under climate change. *Global Ecology and Biogeography* 18: 701–710.

Cumming GS, Child MF. 2009. Contrasting spatial patterns of taxonomic and functional richness offer insights into potential loss of ecosystem services. *Philosophical Transactions of the Royal Society B* 364: 1683–1692.

Folke C, Carpenter S, Elmqvist T, Gunderson L, Holling CS, Walker B. 2002. Resilience and sustainable development: building adaptive capacity in a world of transformations. *Ambio* 31: 437–440.

Garcia JT, Morales MB, Martinez J, Iglesias L, de la Morena EG, Suarez F, Vinuela J. 2006. Foraging activity and use of space by lesser kestrel *Falco naumanni* in relation to agrarian management in central Spain. *Bird Conservation International* 16: 83–95.

Goudie AS, Price Williams D. 1983. The atlas of Swaziland. The Swaziland National Trust Commission, Occasional Paper No. 4, Lobamba.

Herremans M, Herremans-Tonnoeyr D. 2001. Roadside abundance of raptors in the Western Cape Province, South Africa: a three-decade comparison. *Ostrich* 72: 96–100.

Hurst ZM, McCleery RA, Collier BA, Fletcher Jr RJ, Silvy NJ, Taylor PJ, Monadjem A. 2013. Dynamic edge effects in small mammal communities across a conservation–agricultural interface in Swaziland. *PLoS ONE* 8(9). <<http://www.plosone.org/article/info%3Adoi%2F10.1371%2Fjournal.pone.0074520;jsessionid=B8CF68CFF41BD7090A0273866005034D>>. Accessed 24 Sept 2013.

Hurst ZM, McCleery RA, Collier BA, Silvy NJ, Taylor PJ, Monadjem A. 2014. Linking changes in small mammal communities to ecosystem functions in an agricultural landscape. *Mammalian Biology* 79: 17–23.

- Juliana JR, Kotler BP, Brown JS, Mukherjee S, Bouskila A. 2011. The foraging response of gerbils to a gradient of owl numbers. *Evolutionary Ecology Research* 13: 869–878.
- Kay BJ, Twigg LE, Korn TJ, Nicol HI. 1994. The use of artificial perches to increase predation on house mice (*Mus domesticus*) by raptors. *Wildlife Research* 21: 95–106.
- Khumalo S, Chirwa PW, Moyo BH, Syampungani S. 2012. The status of agrobiodiversity management and conservation in major agroecosystems of southern Africa. *Agriculture, Ecosystems and Environment* 157: 17–23.
- Korpimäki E, Norrdahl K. 1989. Predation of Tengman's owls: numerical responses, functional responses and dampening impact on population fluctuations of microtines. *Oikos* 54: 154–164.
- Korpimäki E, Norrdahl K. 1991. Numerical and functional responses of kestrels, short-eared owls, and long-eared owls to vole densities. *Ecology* 72: 814–826.
- Makundi RH, Massawe AW. 2011. Ecologically based rodent management in Africa: potential and challenges. *Wildlife Research* 38: 588–595.
- Manly BFJ, McDonald LL, Thomas DL, McDonald TL, Erickson WP. 2002. Resource selection by animals: statistical design and analysis for field studies. Kluwer Academic Publishers, Dordrecht, the Netherlands.
- Mendelsohn JM. 1981. A study of the black-shouldered kite *Elanus caeruleus*. Dissertation, University of Natal, Pietermaritzburg, South Africa.
- Mendelsohn JM. 1983. Social behaviour and dispersion of the blackshouldered kite. *Ostrich* 54: 1–18.
- Mendelsohn JM. 1984. The timing of breeding in blackshouldered kites in southern Africa. Proceedings of the Fifth Pan-African Ornithological Congress 799–808.

- Mendelsohn JM, Jaksic FM. 1989. Hunting behaviour of black-shouldered kites in the Americas, Europe, Africa, and Australia. *Ostrich* 60: 1–12.
- Mohr K, Vibe-Petersen S, Jeppesen LL, Bildsoe M, Leirs H. 2003. Foraging of multimammate mice, *Mastomys natalensis*, under different predation pressure: cover, patch-dependant decisions and density-dependent GUDs. *Oikos* 100: 459–468.
- Monadjem A. 1999. Geographic distribution patterns of small mammals in Swaziland in relation to abiotic factors and human land-use activity. *Biodiversity and Conservation* 8: 223–237.
- Monadjem A, Perrin MR. 2003. Population fluctuations and community structure of small mammals in a Swaziland grassland over a three-year period. *African Zoology* 38: 127–138.
- Monadjem A, Mahlaba TA, Dlamini N, Eiseb SJ, Belmain SR, Mulungu LS, Massawe AW, Makundi RH, Mohr K, Taylor PJ. 2011. Impact of crop cycle on movement patterns of pest rodent species between fields and houses in Africa. *Wildlife Research* 38: 603–609.
- Monadjem A, Rasmussen MW. 2008. Nest distribution and conservation status of eagles, selected hawks and owls in Swaziland. *Gabar* 19: 1–22.
- Ogada DL, Keesing F. 2010. Decline of raptors over a three-year period in Laikipia, central Kenya. *Journal of Raptor Research* 44: 129–135.
- Parejo D, Aviles JM. 2001. Communal roosting and diet of black-shouldered kites (*Elanus caeruleus*) wintering in south-western Spain. *Journal of Raptor Research* 35: 162–164.
- Pavey CR, Eldridge SR, Heywood M. 2008. Population dynamics and prey selection of native and introduced predators during a rodent outbreak in arid Australia. *Journal of Mammalogy* 89: 674–683.

- Paz A, Jareno D, Arroyo L, Vinuela J, Arroyo B, Mougeot F, Luque-Larena JJ, Fargallo JA. 2013. Avian predators as a biological control system of common vole (*Microtus arvalis*) populations in northwestern Spain: experimental set-up and preliminary results. *Pest Management Science* 69: 444–450.
- Preisser EL, Bolnick DI, Benard MF. 2005. Scared to death? The effects of intimidation and consumption in predator–prey interactions. *Ecology* 86: 501–509.
- Seavy NE, Apodaca CK. 2002. Raptor abundance and habitat use in a highly-disturbed-forest landscape in western Uganda. *Journal of Raptor Research* 36: 51–57.
- Sheffield LM, Crait JR, Edge WD, Wang G. 2001. Response of American kestrels and grey-tailed voles to vegetation height and supplemental perches. *Canadian Journal of Zoology* 79: 380–385.
- Sinclair ARE, Olsen PD, Redhead TD. 1990. Can predators regulate small mammal populations? Evidence from house mouse outbreaks in Australia. *Oikos* 59: 382–392.
- Slotow R, Perrin MR. 1992. The importance of large prey for blackshouldered kite reproduction. *Ostrich* 63: 180–182.
- Sorley CS, Andersen DE. 1994. Raptor abundance in south–central Kenya in relation to land-use patterns. *African Journal of Ecology* 32: 30–38.
- Stenseth NC, Leirs H, Skonhøft A, Davis SA, Pech RP, Andreassen HP, Singleton GR, Lima M, Machang'u RS, Makundi RH, Zhang Z, Brown PR, Shi D, Wan X. 2003. Mice, rats, and people: the bio–economics of agricultural rodent pests. *Frontiers in Ecology and the Environment* 1: 367–375.
- Swolgaard CA, Reeves KA, Bell DA. 2008. Foraging by Swainson's hawks in a vineyard-dominated landscape. *Journal of Raptor Research* 42: 188–196.

- Takele S, Bekele A, Belay G, Balakrishnan M. 2011. A comparison of rodent and insectivore communities between sugarcane plantations and natural habitat in Ethiopia. *Tropical Ecology* 52: 61–68.
- Tarboton WR. 1978. Hunting and the energy budget of the black-shouldered kite. *Condor* 80: 88–91.
- Taylor PJ, Downs S, Monadjem A, Eiseb SJ, Mulungu LS, Massawe AW, Mahlaba TA, Kirsten F, Maltatz EV, Malebane P, Makundi RH, Lamb J, Belmain SR. 2012. Experimental treatment-control studies of ecologically based rodent management in Africa: balancing conservation and pest management. *Wildlife Research* 39: 51–61.
- Thiollay J. 2001. Long-term changes of raptor populations in northern Cameroon. *Journal of Raptor Research* 35: 173–186.
- Thiollay J. 2007. Raptor population decline in West Africa. *Ostrich* 78: 405–413.
- Toland BR. 1987. The effect of vegetative cover on foraging strategies, hunting success and nesting distribution of American kestrels in central Missouri. *Journal of Raptor Research* 21: 14–20.
- Tscharntke T, Klein AM, Kruess A, Steffan-Dewenter I, Thies C. 2005. Landscape perspectives on agricultural intensification and biodiversity—ecosystem service management. *Ecology Letters* 8: 857–874.
- Van Gulck T, Stoks R, Verhagen R, Sabuni CA, Mwanjabe P, Leirs H. 1998. Short-term effects of avian predation variation on population size and local survival of the multimammate rat, *Mastomys natalensis* (Rodentia, Muridae). *Mammalia* 62: 329–339.

- Vibe-Petersen S, Leirs H, DeBruyn L. 2006. Effects of predation and dispersal on *Mastomys natalensis* population dynamics in Tanzanian maize fields. *Journal of Animal Ecology* 75: 213–220.
- Wolff JO, Fox T, Skillen RR, Wang G. 1999. The effects of supplemental perch sites on avian predation and demography of vole populations. *Canadian Journal of Zoology* 77: 535–541.
- World Bank [WB]. 2015. WB homepage. <<http://data.worldbank.org/indicator/AG.LND.AGRI.ZS/countries/1W?display=default>>. Accessed 24 Feb 2015.
- Zar, JH 1984. Biostatistical analysis. Prentice-Hall International, London. 2nd Edition.