UNIVERSITEIT VAN PRETORIA UNIVERSITY OF PRETORIA YUNIBESITHI YA PRETORIA

REFERENCES

- 1. PETROVIC, M., Eljarret, E., Lopez de Alda, M.J., Barcelo, D., 2002. REVIEW: Recent advances in the mass spectrometric analysis related to endocrine disrupting compounds in aquatic environmental samples. *J.Chromatogr. A*, 974, 23-51.
- 2. LOPEZ de ALDA, M.J., Barcelo, D., 2001. Review of analytical methods for the determination of estrogens and progestogens in waste water. *Fresenius J. Anal. Chem.*, 371, 437-447.
- 3. SILVIA DIAZ-CRUZ, M., Lopez de Alda, M.J., Lopez, R., Barcelo, D., 2003. Determination of estrogens and progetogens by mass spectrometric techniques (GC/MS, LC/MS and LC/MS/MS). *J. Mass Spectrom.*, 38, 917-923.
- 4. VIDAEFF, A.C., Sever, L.E., 2005. REVIEW: In utero exposure to environmental estrogens and male reproductive health: a systematic review of biological and epidemiologic evidence. *Reproductive Toxicology*, 20, 5-20.
- GWRC: Global Water Research Coalition, London, U.K., Sept. 2003. Endocrine Disrupting Compounds - Priority list of EDCs. <u>http://www.globalwaterresearchcoalition.net</u>. Access 5 April 2006.
- 6. CADBURY, D., 1997. The feminization of nature our future at risk, *Penguin books*, London, England. ISBN 0-14-026205-9.
- ET AL, Schenectady International, 2 April 2003. Alkylphenols Category Section

 Development of categories and test plans / Chemical right-to-know initiative HPV challenge
 program. <u>http://www.epa.gov/hpv/pubs/summaries/alkylphn/c13007rt.pdf</u>. Access 5 Oct 2007.
- 8. COUSINS, I.T., Staples, C.A., Kle^{*}cka, G.M., Mackay, D., 2002. A Multimedia Assessment of the Environmental Fate of Bisphenol A. *Human and Ecological Risk Assesment*, 8, 1107-1135.
- 9. PORTER, A.J., Hayden, N. J., Nonylphenol in the Environment: A Critical Review, Department of Civil and Environmental Engineering University of Vermont Burlington, VT 05405.
- 10. de VOOGT, P., de Beer, K., van der Wielen, F., 1997. Determination of alkyphenol ethoxylates in industrial and environmental samples. *Trends in analytical chemistry*, 16, 584-595.
- 11. VOM SAAL, F., 16 February 2007. Oral presentation, American Association for the Advancement of Science (AAAS), San Fransico, California, U.S.A.
- 12. CORNWELL, T., Cohick, W., Raskin, I., 2004. REVIEW: Dietary phytoestrogens and health. *Phytochemistry*, 65, 995-1016.
- 13. HENLEY, D.V., Lipson, N., Korach, K.S., Bloch, C.A., 2007. Prepubertal Gynecomastia linked to lavendar and tea tree oils. *N. Engl. J. Med.*, 356, 479-485.
- 14. Et Al, Chemical, Toxicity, Safety and Environmental Analysis Information for Formaldehyde, <u>http://www.instantref.com/formald.htm</u>, access: 27 Aug 1999.
- 15. OSHA, Occupational Safety and Health Administration U.S. Department of Labor, OSHA Regulations Formaldehyde. 1910.1048. <u>http://www.osha-slc.gov</u>, access: 2 Feb 2004.



- 16. NIOSH Manual of Analytical Methods, Method 2016, 1998. Formaldehyde by HPLC-UV. 4th Edition.
- 17. CRUMP, D.R. in Volatile Organic Compounds in the Atmosphere, editors Hester, R.E., Harrison, R.M., 1995. Volatile organic Compounds in Indoor air. *Issues in Environmental Science and Technology*, The Royal Society of Chemistry, U.K, 4, 109-124.
- Occupational Safety and Health Administration U.S.Department of Labor, OSHA Fact Sheets -01/01/1995 - Occupational Exposure to formaldehyde. <u>http://www.osha-slc.gov</u>, access: 2 Feb 2004.
- 19. McNAUGHT, A.D., Wilkinson, A., 1997. *Compendium of chemical terminology IUPAC recommendations*, 2nd edition, Blackwell Science Ltd. Oxford. ISBN 0865426848.
- 20. WILLOUGHBY, R., Sheenan, E., Mitrovich, S., 1998. A global view of LC/MS how to solve your most challenging analytical problems, First Edition, Global view publishing, Pitsburgh, Pennsylvania. ISBN 09660813-0-7.
- 21. INGERSLEV, F., Halling-Sørensen, B., APS Danish Environmetal Protection Agency, 2003. "Evaluation of Analytical Chemical Methods for Detection of Estrogens in the Environment Working Report No. 44, 2003"
- 22. OOSTERKAMP, A.J., Hock, B., Seifert, M., Irth, H., 1997. Novel Monitoring strategies for xenoestrogens. *Trends in analytical chemistry*, 16, 544.
- 23. RICHARDSON, S.D., 2004. Environmental Mass Spectrometry: Emerging Contaminants and Current Issues. *Anal. Chem.*, 76, 3337-3364.
- 24. KUSTER, M., Lopez de Alda, M.J., Barcelo, D., 2004. Analysis and distribution of estrogens and progestogens in sewage sludge, soils and sediments. *Trends in analytical chemistry*, 23, 790.
- 25. SHIMADA, K., Mitamura, K., Higashi, T., 2001. Gas chromatography and high-performance liquid chromatography of natural steroids. *J.Chromatogr. A*, 935, 141-172.
- 26. SCHNEIDER, C., Schöler, H.F., Schneider, R.J., 2005. Direct sub-ppt detection of the endocrine disruptor ethinylestradiol in water with a chemiluminescence enzyme-linked immunosorbent assay. *Anal. Chim. Acta*, 551, 92-97.
- 27. LI, Z., Wang, S., Lee, N.A., Allan, R.D., Kennedy, I.R., 2004. Development of a solid-phase extraction—enzyme-linked immunosorbent assay method for the determination of estrone in water. *J.Chromatogr. A*, 503, 171-177.
- 28. GOMES, R.L.; Avcioglu, E.; Scrimshaw, M.D.; Lester, J.N., 2004. Steroid estrogen determination in sediment and sewage sludge: a critique of sample preparation and chromatographic/mass spectrometry considerations, incorporating a case study in method development. *Trends in analytical chemistry*, 23, 737-744.
- 29. SANTOS, F.J., Galceran, M.T., 2003. Review: Modern developments in gas chromatographymass spectrometry based environmental analysis. *J. Chromatogr. A*, 1000(27), 125-151.

- 30. NAKAMURA, S., Sian, T.H., Daishima, S., 2001. Determination of estrogens in river water by gas chromatography– negative-ion chemical-ionization mass spectrometry. *J.Chromatogr. A*, 919, 275-282.
- 31. BOESL, U., Heger, HJ., Zimmermann, R., Nagel, H., Püffel, P., 2000. Laser Mass Spectrometry in Trace Analysis, *Encyclopedia of Analytical Chemistry*, R.A. Meyers (Ed.), John Wiley & Sons Ltd, Chichester, 2087-2118.
- 32. MÜHLBERGER, F.; Zimmermann, R.; Kettrup, A., 2001. A mobile mass spectrometer for comprehensive on-line analysis of trace and bulk components of complex gas mixtures. *Anal. Chem.*, 73, 3590-3604.
- 33. HEGER, H. J.; Zimmermann, R.; Dorfner, R.; Beckmann, M.; Griebel, H.; Kettrup, A.; Boesl, U., 1999. On-Line Emission Analysis of Polycyclic Aromatic Hydrocarbons down to pptv Concentration Levels in the Flue Gas of an Incineration Pilot Plant with a Mobile Resonance-Enhanced Multiphoton Ionization Time-of-Flight Mass Spectrometer. *Anal. Chem.*, 71, 46-57.
- DORFNER, R.; Ferge, T.; Yeretzian, C.; Kettrup, A.; Zimmermann, R., 2004. Laser Mass Spectrometry as On-Line Sensor for Industrial Process Analysis: Process Control of Coffee Roasting. *Anal. Chem.*, 76 (5), 1386-1402.
- 35. DORFNER, R.; Ferge, T.; Kettrup, A.; Zimmermann, R.; Yeretzian, C., 2003. J. Agric.Food Chem., 51 (19), 5768-5773.
- 36. ZIMMERMANN, R.; Heger, H. J.; Kettrup, A., 1999. On-line Monitoring of Traces of Aromatic-, Phenolic- and Chlorinated Components in the Flue Gases of Industrial Incinerators, Mineral Oil Headspace and Cigarette Smoke by Direct-Inlet Laser Ionization-Mass Spectrometry (REMPI-TOFMS). *Fresenius' J. Anal. Chem.*, 363, 720-730.
- 37. HAULER, T. E.; Boesl, U.; Kaesdorf, S.; Zimmermann, R., 2004. Mobile resonance enhanced multiphoton ionisation–time-of-flight mass spectrometer with a novel hybrid laser desorption/molecular beam ion source for rapid detection of aromatic trace compounds from gas phase and solid samples. *J. Chromatogr. A*, 1058, 39-49.
- 38. OTSON, R., Fellin, P., 1988. Determination of airborne aldehydes literature Review. *Sceince of the Total Environment*, 77, 95.
- 39. KENNEDY, E.R., Teass, A.W., Gagnon, Y.T. in Formaldehyde Analytical chemistry and toxicology, Editor Turoski, V., 1985. Industrial Hygiene Sampling and analytical methods for formaldehyde Past and present. *Advances in chemistry and toxicology series*, 210, 1-11.
- 40. BEREZKIN, V. G.; Drugov, Y. S., 1991. Gas chromatography in air pollution analysis. *J. Chromatogr. Libr.*, 49, 190-194.
- 41. MUNSON, B., 1977. Chemical Ionization Mass Spectrometry. Anal. Chem., 49, 772-778A.
- 42. LUBMAN, D. M., Ed., 1990. Lasers and Mass Spectrometry; Oxford University Press: New York, ISBN 0195059298.
- 43. COOL, T. A.; Williams, B. A., 1992. Ultrasensitive Detection of Chlorinated Hydrocarbons by Resonance Ionization. *Combust. Sci. Technol.*, 82, 67.



- 44. BOESL, U.; Zimmermann, R.; Weickhardt, C.; Lenoir, D.; Schramm, K.-W.; Kettrup, A.; Schlag, E.W., 1994. Resonance Enhanced Laser Ionisation as a Compound-Selective Ion Source for Mass Spectrometry: Technique and Experimental Setup, *Chemosphere*, 29, 1429-1440.
- 45. MILLER, J. C.; Compton, R. N., Cooper, C.D., 1982. Vacuum Ultraviolet Spectroscopy of Molecules Using Third-Harmonic Generation in Rare Gases. J. Chem. Phys., 76 (8), 3967.
- 46. NOMAYO, M.; Thanner, R.; Pokorny, H.; Grotheer, H.-H.; Stützle, R., 2001. Measurements in the raw gas of a full scale municipal waste incinerator using a wavelength resolved REMPI mass spectrometer. *Chemosphere*, 43, 461-467.
- 47. BALTUSSEN, H.A., 2000. New Concepts in Sorption Based Sample Preparation for Chromatography. D.Phil. Thesis, Technical University Eindhoven, Eindhoven.
- 48. BEREZKIN, V.G., Drugov, Y.S, 1991. Gas chromatography in air pollution analysis. *Journal of Chromatography Library* -Elsevier, Germany, 49, 35-119.
- 49. PAWLISZYN, J., 1997. Solid Phase Microextraction Theory and Practice. Wiley-VCH, Canada, ISBN 0471190349.
- 50. BASHEER, C., Parthiban, A., Jayaraman, A., Kee Lee, H., Valiyaveettil, S., 2005. Determination of alkylphenols and bisphenol-A. A comparative investigation of functional polymer-coated membrane microextraction and solid-phase microextraction techniques. *J.Chromatogr. A*, 1087, 274-282.
- 51. KAWAGUCHI, M., Sakui, N., Okanouchi, N., Ito, R., Saito, K., Izumi, SI., Makino, T., Nakazawa, H., 2005. Stir bar sorptive extraction with in situ derivatization and thermal desorption-gas chromatography–mass spectrometry for measurement of phenolic xenoestrogens in human urine samples. *J.Chromatogr. B.*, 820, 49-57.
- 52. CHANG, CM., Chou, CC., Lee, MR., 2005. Determining leaching of bisphenol A from plastic containers by solid-phase microextraction and gas chromatography–mass spectrometry. *Anal. Chim. Acta*, 539, 41-47.
- 53. LATORREA, A., Lacorteb, S., Barceló, D., Monturya, M., 2005. Determination of nonylphenol and octylphenol in paper by microwave-assisted extraction coupled to headspace solid-phase microextraction and gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1065, 251-256.
- 54. NAKAMURA, S., Daishima, S., 2004. Simultaneous determination of alkylphenols and bisphenol A in river water by stir bar sorptive extraction with in situ acetylation and thermal desorption–gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1038, 291-294.
- 55. KAWAGUCHIA, M., Sakuia, N., Okanouchi, N., Ito, R., Saito, K., Nakazawa, H., 2005. Stir bar sorptive extraction and trace analysis of alkylphenols in water samples by thermal desorption with in tube silylation and gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1062, 23-29.
- 56. KAWAGUCHI, M.; Inoue, K.; Yoshimura, M.; Sakui, N.; Ito, R.; Izumi, S-I.; Makino, T.; Okanouchi, N.; Nakazawa, H., 2004. Stir bar sorptive extraction and thermal desorption-gas chromatography-mass spectrometry for the measurement of 4-nonylphenol and 4-tertoctylphenol in human biological samples. *J.Chromatogr. B.*, 799, 119-125.

- 57. BROWN, L., du Preez, J.L., Meintjies, E., 1997. The qualitative and quantitative evaluation of estrogen and estrogen-mimicking substances in the South African water environment: a 1996-1997 perspective. *Report on WRC project*.
- 58. PETROVICÈ, M., Eljarrat, E., Lopez de Alda, M.J., Barcelo, D., 2001. Analysis and environmental levels of endocrine disrupting compounds in freshwater sediments. *Trends in analytical chemistry*, 20, 637.
- 59. PAWLISZYN, J., 2006. Why move analysis from laboratory to on-site? *Trends in analytical chemistry*, 25, 633-634.
- 60. MARTOS, P.A., Pawliszyn, J., 1998. Sampling and determination of formaldehyde using solid phase microextraction with on-fibre derivitization. *Anal. Chem.*, 70, 2311-2320.
- 61. FERNANDES, M.J., 2001. Concentration and derivatisation in silicone rubber traps for gas chromatographic trace analysis of aldehydes. *MSc thesis*, University of Pretoria, Pretoria, South Africa.
- 62. LERCH, O., Zinn, P., 2003. Derivatisation and Gas chromatography-chemical ionisation mass spectrometry of selected synthetic and natural endocrine disruptive chemicals. *J.Chromatogr. A*, 991, 77-97.
- 63. ORTNER, E.K., 1999. Analysis of aqueous samples with the multichannel silicone rubber trap and capillary gas chromatography. *D.Phil. Thesis*, University of Pretoria, Pretoria, South Africa.
- 64. ORTNER, E.K., 1994. Alternative concentration techniques for the trace analysis of semivolatile organic air pollutants by capillary gas chromatography. *MSc thesis*, University of Pretoria, Pretoria, South Africa.
- 65. ORTNER, E.K., Rohwer, E.R., 1996. Trace analysis of semi-volatile organic air pollutants using thick film silicone rubber traps with capillary gas chromatography. *J. High Resolut. Chromatogr.*, 19, 339-344.
- 66. ORTNER, E.K., Rohwer, E.R., 1999. Trapping efficiency of aqueous pollutants using multichannel thick film silicone rubber traps with capillary gas chromatography. *J.Chromatogr. A*, 863, 57-68.
- 67. ORTNER, E.K.; Rohwer, E.R., 1999. Trace determination of organic compounds in water by direct enrichment in multichannel thick film silicone rubber traps with capillary gas chromatography. *J. High Resolut. Chromatogr.*, 22 (9), 521-526.
- 68. HASSET, A.J.; Rohwer, E.R., 1999. Analysis of odorous compounds in water by isolation by closed-loop stripping with a multichannel silicone rubber trap followed by GC-MS. *J.Chromatogr. A.*, 849, 521-528.
- 69. HARPER, M., 2000. Review: Sorbent trapping of volatile organic compounds from air. *J.Chromatogr. A.*, 885, 129-151.
- 70. NAMIESNIK, J., 1988. Preconcentration of gaseous organic pollutants in the atmosphere. *Talanta*, 35, 567-587.



- 71. GRIMALT, J.O., Flesca, N.G., Castellnou, A., 1997. Refridgerated multibed adsorption in sampling and analysis of atmospheric light hydrocarbons at ppb(v/v) and sub ppb(v/v) concentrations. *J.Chromatogr. A.*, 778, 269-277.
- 72. MASTROGIACOMO, A.R., Pierini, E., Sampaulo, L., 1995. A comparison of the critical parameters of some adsorbents employed in trapping and thermal desorption of organic pollutants. *Chromatographia*, 41(11), 599-604.
- 73. ET AL, 1999. SUPELCO Product catalogue, http://www.sigmaaldrich.com/Brands/Supelco_Home.html.
- 74. HUCK, C.W., Bonn, G.K., 2000. Review: Recent developments in polymer based sorbents for solid phase extraction. *J.Chromatogr. A.*, 885, 51-72.
- 75. SCHMIED, W., Przewosnik, M., Bachmann, K., 1989. Determination of traces of aldehydes and ketones in the troposphere via solid phase derivatisation with DNSH. *Fresenius J. Anal. Chem.*, 335, 464-468.
- 76. LEMPUHL, D.W., Birks, J.W., 1996. New gas chromatographic-electron-capture detection method for the determination of atmospheric aldehydes and ketones based on cartridge sampling and derivatisation with 2,4,6-trichlorophenylhydrazine. *J.Chromatogr. A.*, 740A, 71-81.
- 77. LAHANIATI, M., Calogirou, A., Duane, M., Larsen, B., Kotzias, D., 1998. Identification of biogenic carbonyls in air with O-(2,3,4,5,6-pentafluorobenzyl)hydroxylamine hydrochloride (PFBHA) coated C18-silica gel cartridges. *Fresenius.Environ.Bull.*, 7, 302-307.
- 78. STASHENKO, E.E., Ferreira, M.C., Sequeda, L.G., Martinez, J.R.; Wong, J.W., 1997. Comparison of extraction methods and detection systems in the gas chromatographic analysis of volatile carbonyl compounds. *J.Chromatogr. A*, 779, 360-369.
- DUNEMANN, L., Begerow, J., Bucholski, A. Executive editor Gerhartz W., 1988. Sample preparation for trace analysis Ullmann's Encyclopedia of Industrial Chemistry. VCH, vol B5, 84-89.
- 80. STILES, R., Yang, I., Lippincott, R.L., Murphy, E., Buckley, B., 2007. Identifying Potential Sources of Background Contaminants Resulting From Solid Phase Extraction and Solid Phase Microextraction. *J.Sep.Sci*, 30, 1029-1036.
- 81. POOLE, C.F., Poole, S.K., 1991. *Chromatography Today*, 5th Edition, Elsevier, Amsterdam, The Netherlands, 735-840. ISBN 0444891617.
- BALTUSSEN, E., David, F., Sandra, P., Janssen, H.G., Cramers, C.A., 1997. A new method for sorptive enrichment of gaseous samples: Application in air analysis and natural gas characterization. *J. High Resolut. Chromatogr.*, 20(7), 385-393.
- 83. BALTUSSEN, E., David, F., Sandra, P., Janssen, H.G., Cramers, C.A., 1998. Capillary GC determination of amines in aqueous samples using sorptive preconcentration on polydimethylsiloxane and polyacrylate phases. *J. High Resolut. Chromatogr.*, 21, 645-648.
- 84. BURGER, B.V., le Roux, M., 1992. Headspace gas analysis:Comparison of the efficiency of thick film, ultra thick film, and activated charcoal Open tubular capillary traps for the



concentration of volatile, airborne, organic compounds. J. High Resolut. Chromatogr., 15 (6), 373-376.

- 85. BLOMBERG, S., Roeraade, J., 1987. Preparative capillary gas chromatography II. Fraction collection on traps coated with a very thick film of immobilized stationary phase. *J. Chromatogr. A.*, 394, 443-453.
- 86. BLOMBERG, S., Roeraade, J., 1988. A technique for coating capillary columns with a very thick film cross-linked stationary phase. *J. High Resolut. Chromatogr.*, 11(6), 457-461.
- 87. BURGER, B.V., le Roux, M., Burger, W.J.G., 1990. Headspace analysis: A novel method for the production of capillary traps with ultra-thick stationary phase layers. *J. High Resolut. Chromatogr.*, 13(11), 777-779.
- 88. BALTUSSEN, E., David, F., Sandra, P., Janssen, H.G., Cramers, C.A., 1998. Sorption Tubes packed with polydimethylsiloxane: A new and promising technique for the preconcentration of volatiles and semi-volatiles from air and gaseous samples. *J. High Resolut. Chromatogr.*, 21(6), 332-340.
- 89. BALTUSSEN, E., David, F., Sandra, P., Janssen, H.G., Cramers, C.A., 1998. Retention model for sorptive extraction-thermal desorption of aqueous samples: application to the automated analysis of pesticides and polyaromatic hydrocarbons in water samples. *J.Chromatogr. A.*, 805, 237-247.
- BALTUSSEN, E., den Boer, A., Sandra, P., Janssen, H.G., Cramers, C.A., 1999. Monitoring of nicotine in air using sorptive enrichment on polydimethylsiloxane and TD-CGC-NPD. *Chromatographia*, 49, 520-524.
- 91. MOL, H.G.J., Janssen, H.G., Cramers, C.A., 1993. Use of open tubular trapping columns for on-line extraction-capillary gas chromatography of aqueous samples. *J. High Resolut. Chromatogr.*, 16, 413-418.
- 92. WELSCH, T., Teichmann, U., 1991. The Thermal imobilization of hydroxy-terminated Silicone phases in high-temperature-Silylated glass capillaries. A study of reaction mechanisms. *J. High Resolut. Chromatogr.*, 14(3), 153-159.
- 93. LORD, H., Pawliszyn, J., 2000. Review: Evolution of Solid Phase Microextraction technology. *J.Chromatogr. A*, 885, 153-193.
- 94. COLAS, A., Curtis, J., (DOW CORNING), 2005. Silicone Biomaterials: History and Chemistry & Medical applications of silicones, Reprinted from Biomaterials Science - An introduction to materials in medicine, 2nd Edition, Elsevier.
- 95. BALTUSSEN, E., Sandra, P., David, F., Cramers, C., 1999. Stir Bar Sorptive Extraction (SBSE), a novel extraction technique for aqueous samples: Theory and Principles. *J.Microcolumn Separations*, 11, 737-747.
- 96. LOVKVIST, P., Jonsson, J.A., 1987. Capacity of sampling and preconcentration columns with a low number of theoretical plates. *Anal. Chem.*, 59(3), 818-821.
- 97. GROB, K., Habich, A., 1985. Headspace gas analysis: the role and design of concentration traps specifically suitable for capillary GC. *J.Chromatogr. A*, 321, 45-58.



- 98. BURGER, B.V., Munroe, Z., 1986. Headspace gas analysis: Quantitative trapping and thermal desorption of volatiles using fused silica open tubular capillary traps. *J. Chromatogr. A*, 370, 449-464.
- 99. BURGER, B.V., le Roux, M., Munroe, Z.M., Wilken, M.E., 1991. Production and use of capillary traps for headspace gas chromatography of airborne volatile organic compounds. *J.Chromatogr. A*, 552, 137-151.
- 100. EISERT, R., Pawliszyn, J., 1997. Automated In-Tube Solid-Phase Microextraction Coupled to High-Performance Liquid Chromatography. *Anal. Chem.*, 69, 3140-3147.
- 101. MITANI, K., Kataoka, H., 2004. Online Concentration by In-Tube Solid Phase Microextraction and Its Application to Food Analysis. *Foods Food Ingredients J. Jpn.*, 209.
- 102. NARDI, L., 2002. Coupled in-tube SPME-HRGC: A complementary SPME technique to analyze aqueous samples by GC. *American Laboratory*, 30-37.
- 103. WU, J., Tragas, C., Lord, H., Pawliszyn, J., 2002. Analysis of polar pesticides in water and wine samples by automated in-tube solid-phase microextraction coupled with high-performance liquid chromatography-mass spectrometry. *J.Chromatogr. A*, 976, 357-367.
- 104. FERNANDES-WHALEY, M., Mühlberger, F., Whaley, A., Adam, T., Zimmermann, R., Rohwer, E., Walte, A., 2005. On-Line Derivatization for Resonance-Enhanced Multiphoton Ionization Time-of-Flight Mass Spectrometry: Detection of Aliphatic Aldehydes and Amines via Reactive Coupling of Aromatic Photo Ionization Labels. *Anal. Chem.*, 77, 1-10.
- 105. ET AL, 1998. Supelco Bulletin 923 Solid Phase Microextraction: Theory and Optimization of Conditions Sigma-Aldrich Co, 1-8.
- 106. MARSEL, J., Velikonja Bolta, S., Zupancic Kralji, L., 1998. Gas chromatographic determination of formaldehyde in air using SPME sampling. *Chromatographia*, 48, 95-99.
- 107. STASHENKO, E.E.; Martinez, J.R., 2004. Derivatization and solid-phase microextraction. *Trends in analytical chemistry*, 23, 553-561.
- 108. TIENPONT, B., David, F., Bicchi, C., Sandra, P., 2000. Headspace Sorptive extraction, Riva Del Garda Chromatography Conference Poster, 1-11.
- 109. BICCHI, C., Corderoa, C., Libertoa, E., Rubioloa, P., Sgorbinia, B., Sandra, P., 2005. Impact of phase ratio, polydimethylsiloxane volume and size, and sampling temperature and time on headspace sorptive extraction recovery of some volatile compounds in the essential oil field. *J.Chromatogr. A*, 1071, 111-118.
- 110. BURGER, B.V., Marx, B., le Roux, M., Burger, W.J.G., 2006. Simplified analysis of organic compounds in headspace and aqueous samples by high-capacity sample enrichment probe. *J.Chromatogr. A*, 1121, 259-267.
- 111. SANDRA, P., Baltussen, E., David, F., Hoffmann, A., 2000. A novel extraction technique for aqueous samples: Stir bar sorptive extraction, *Gerstel Application Note 1*, <u>www.gerstel.de/application.htm</u>.



- 112. TIENPONT, B.; David, F.; Desmet, K.; Sandra, P., 2002. Stir bar sorptive extraction-thermal desorption-capillary GC-MS applied to biological fluids. *Anal. Bioanal. Chem.*, 373, 46-55.
- 113. BICCHI, C., Cordero, C., Rubiolo, P., Sandra, P., 2003. SBSE in food matrices. *European Food Research and Technology*, 216, 449.
- 114. BALTUSSEN, E., Cramers, C.A., Sandra, P.J.F., 2002. Sorptive sample preparation a review. *Anal. Bioanal. Chem.*, 373, 3-22.
- 115. RAYMOND, A., Guiochon, G., 1975. The use of graphitized carbon black as a trapping material for organic compounds in light gases before a GC-analysis. *J. Chromatogr. Sci*, 13, 173-177.
- 116. POOLE, C.F., Schuette, S.A., 1983. Isolation and concentration techniques for Capillary column GC analysis. *J. High Resolut. Chromatogr*, 6 (10), 526-549.
- 117. FLUKA, 2002. "Derivatization reagents ANALYTIX Advances in Analytical Chemistry", Sigma-Aldrich Co. <u>http://www.sigmaaldrich.com</u>.
- 118. ROSENFELD, J.M., 2003. Derivatization in the current practice of analytical chemistry. *Trends in analytical chemistry*, 22, 785-798.
- 119. SUPELCO, 1997. Guide to Derivatization Reagents for GC. *Bulletin 909A*, Sigma-Aldrich Co. <u>http://www.sigmaaldrich.com</u>.
- 120. MARTINEZ, D., Borrull, F., Calull, M., Ruana, J., Colom, A., 1998. Application of solid phase extraction membrane discs in the determination of haloacetic acids in water by gas chromatography-mass spectrometry. *Chromatographia*, 48, 811-816.
- 121. KATAOKA, H., 1996. REVIEW: Derivatization reactions for the determination of amines by gas chromatography and their applications in environmental analysis. *J.Chromatogr. A*, 733, 19-34.
- 122. SEGURA, J., Ventura, R., Jurado, C., 1998. REVIEW: Derivatization procedures for gas chromatographic mass spectrometric determination of xenobiotics in biological samples, with special attention to drugs of abuse and doping agents. *J. Chromatogr. B*, 713, 61-90.
- 123. XIE, Y., 2001. Technical note: Analyzing haloacetic acids using gas chromatography/mass spectrometry. *Wat. Res.*, 35, 1599-1602.
- 124. SARRION, M.N., Santos, F.J., Galceran, M.T., 1999. Solid phase microextraction coupled with gas chromatography-ion trap mass spectrometry for the analysis of haloacetic acids in water. *J.Chromatogr. A*, 859, 159-171.
- 125. NIKOLAOU, A.D., Golfinopoulos, S.K., Kostopoulou, M.N., Lekkas, T.D., 2002. Technical note: Determination of haloacetic acids in water by acidic methanol esterification-GC-ECD method. *Wat. Res.*, 36, 1089-1094.
- 126. PIERCE®, 1998. Acylation Derivatization Reagents, Pierce Chemical Company. <u>www.piercenet.com</u>.



- 127. LITTLE, J.L., 1999. Artifacts in Trimethylsilyl Derivatization Reactions and Ways to Avoid Them. *J.Chromatogr. A*, 844, 1-22.
- 128. SCHLITT, H., 1997. Impinger sampling coupled to HPLC by a modified autoinjector interface. *J.Chromatogr. A*, 762, 187-192.
- 129. RUTTEN, G.A., Burtner, C.W.J., Rijks, J.A., Visser, H., 1988. The determination of aldehydes in exhaust gases of LPG fuelled engines. *Chromatographia*, 26, 274-280.
- 130. BEASLEY, R.K., Hoffmann, C.E., Rueppel, M.L., Worley, J.W., 1980. Sampling of formaldehyde in air with coated solid sorbent and determination by high Performance liquid chromatography. *Anal. Chem.*, 52, 1110-1114.
- 131. GUENIER, J.P., Simon, P., Delcourt, J., Didierjean, M.F., Lefevre, C., Muller, J., 1984. Air Sampling of aldehydes application to chromatographic determination of formaldehyde and acetaldehyde. *Chromatographia*, 18(3), 137-144.
- 132. SAKURAGAWA, A., Yoneno, T., Inoue, K., Okutani, T., 1999. Trace analysis of carbonyl compounds by liquid chromatography-mass spectrometry after collection as 2,4-dinitrophenylhydrazine derivatives. *J. Chromatogr. A*, 844, 403-408.
- 133. DALENE, M., Persson, P., Skarping, G., 1992. Determination of formaldehyde in air by chemisorption on glass filters impregnated with 2,4-dinitrophenylhydrazine using gas chromatography with thermionic specific detection. *J.Chromatogr. A*, 626, 284-288.
- 134. SESANA, G., Nano, G., Baj, A., Balestreri, S., 1991. New sampling tool for airborne volatile aldehydes. *Fresenius J. Anal. Chem.*, 339(7), 485-487.
- 135. OTSON, R., Fellin, P., Tran, Q., Stoyanoff, R., 1993. Examination of sampling methods for assessment of personal exposures to airborne aldehydes. *Analyst*, 118(10), 1253-1259.
- 136. NISHIKAWA, H., Sakai, T., 1995. REVIEW: Derivitization and chromatographic determination of aldehydes in gaseous and air samples. *J.Chromatogr. A*, 710(1), 159-165.
- 137. GROSJEAN, E., Green, P.G., Grosjean, D., 1999. Liquid chromatography analysis of carbonyl (2,4-Dinitrophenyl)hydrazones with detection by diode array Ultraviolet spectroscopy and by atmospheric pressure negative chemical ionization mass spectrometry. *Anal. Chem.*, 71, 1851-1861.
- 138. DYE, C., Oehme, M., 1992. Comments concerning the HPLC seperation of Acrolein from other C3 carbonyl compounds as 2,4-Dinitrophenylhydrazones: A proposal for improvement. *J. High Resolut. Chromatogr.*, 15(1), 5-8.
- 139. STASHENKO, E.E., Puertas, M.A., Salgar, W., Delgado, W., Martinez, J.R., 2000. Solid-Phase Microextraction with on-fibre derivitization applied to the analysis of volatile carbonyl compounds, Riva Del Garda Chromatography Conference Poster, 1-7.
- 140. CANCILLA, D.A., Chou, C., Barthel, R., Que Hee, S.S., 1992. Characterization of the O-(2,3,4,5,6-pentafluorobenzyl)-hydroxylamine hydrochloride (PFBOA) derivatives of some aliphatic mono- and dialdehydes and quantitative water analysis of these aldehydes. *J.AOAC International*, 75, 842-854.



- 141. NAWROCKI, J., Kalkowska, I., Dabrowska, A., 1996. Optimisation of solid phase extraction method for analysis of low ppb amounts of aldehydes-ozonation by-products. *J. Chromatogr. A*, 749A, 157-163.
- 142. OJALA, M., Kotiaho, T., Siirila, J., Sihvonen, M., 1994. Analysis of aldehydes and ketones from beer as O-(2,3,4,5,6-pentafluorobenzyl)hydroxylamine derivatives. *Talanta*, 41, 1297-1309.
- 143. VIDAL, J.P., Estreguil, S., Cantagrel, R., 1993. Quantitative analysis of Cognac carbonyl compounds at the ppb level by GC-MS of their O (pentafluorobenzyl amine) derivatives. *Chromatographia*, 36, 183-186.
- 144. STRASSNIG, S., Wenzl, T., Lankmayr, E.P., 2000. Microwave-assisted derivatization of volatile carbonyl compounds with O-(2,3,4,5,6-pentafluorobenzyl)hydroxylamine. *J.Chromatogr. A*, 891, 267-273.
- 145. WUE, LJ., Que Hee, S.S., 1995. A solid sorbent personal air sampling method for aldehydes. *Am. Ind. Hyg. Assoc. J.*, 56(4), 362-367.
- 146. TSAI, S.W., Que Hee, S.S., 1999. A new passive sampler for aldehydes. *Am. Ind. Hyg. Assoc. J.*, 60, 463-473.
- 147. MARTOS, P.A., Pawliszyn, J., 1999. Time weighted average sampling with solid-phase microextraction device: Implications for enhanced personal exposure monitoring to airborne pollutants. *Anal. Chem.*, 71, 1513-1520.
- 148. KENNEDY, E.R., O'Connor, P.F., Gagnon, Y.T., 1984. Determination of Acrolein in air as an oxazolidine derivative by gas chromatography. *Anal. Chem.*, 56, 2120-2123.
- 149. THOMAS, P., McGill, C.L., Towill, C.D., 1997. Determination of formaldehyde by conversion to hexahydroxazolo[3,4-a]pyridine in a denuder tube with recovery by thermal desorption and analysis by GC-MS. *Analyst*, 122, 1471-1476.
- 150. NIOSH Method 2541, 1994. Formaldehyde by GC. *NIOSH Manual of Analytical Methods*. <u>www.cdc.gov/niosh/nmam</u>.
- 151. NIOSH Method 2539, 1994. Aldehydes, screening. *NIOSH Manual of Analytical Methods*. <u>www.cdc.gov/niosh/mam</u>.
- 152. MIYAKE, T., Shibamoto, T., 1995. Quantitative analysis by GC of volatile carbonyl compounds in cigarette smoke. *J.Chromatogr. A*, 693, 376-381.
- 153. YASUHARA, A., Shibamoto, T., 1994. Gas chromatographic determination of trace amounts of aldehydes in automobile exhaust by cysteamine derivatisation methods. *J.Chromatogr. A*, 672, 261-266.
- 154. KARTSOVA, L.A., Makarova, Y.L., Stolyarov, B.V., 1997. Selective gas chromatographic determination of formaldehyde in air. *J.Anal.Chem*, 52(4), 337-340.
- 155. PELTONEN, K., Pfaffli, P., Itkonen, A., 1984. Determination of aldehydes in air as dimethone derivatives by gas chromatography with electron capture detection. *J.Chromatogr. A*, 315, 412-416.



- 156. RECHE, F., Garrigos, M.C., Sanchez, A., Jimenez, A., 2000. Simultaneous supercritical fluid derivatization and extraction of formaldehyde by the Hantzsch reaction. *J. Chromatogr. A*, 896, 51-59.
- 157. ZUREK, G., Karst, U., 1999. Liquid Chromatography mass Spectrometry method for the determination of aldehydes derivatized by the Hantzsch reaction. *J. Chromatogr. A*, 864, 191-197.
- 158. BLAU, K., King, G.S., 1977. Handbook of derivatives for Chromatography. Heyden & Son Ltd, London, U.K.
- 159. PAN, L., Chong, M., Pawliszyn, J., 1997. Determination of amines in air and water using derivatization combined with solid-phase microextraction. *J. Chromatogr. A*, 773, 249-260.
- 160. GYLLENHAAL, O.; Hoffmann, K-J.; Lamm, B.; Simonsson, R.; Vessman, J., 1986. Degradation of perfluoroacyl derivatives of tocainide and some of its analogues in the presence of an excess of anhydride reagent. *J.Chromatogr. A*, 355, 127-140.
- 161. MASUDA, Y.; Hoffmann, D., 1969. A method for the determination of primary amines of polynuclear aromatic hydrocarbons. *J. Chromatogr. Sci*, 7, 694-697.
- 162. CHIA, K-J.; Huang, S-D., 2005. Simultaneous derivatization and extraction of amphetamine like drugs in urine with headspace solid-phase microextraction followed by gas chromatography-mass spectrometry. *Anal. Chim. Acta*, 539, 49-54.
- 163. KOSTER, E.H.M.; Bruins, C.H.P.; de Jong, G.H., 2002. On-fiber derivatization for direct immersion solid-phase microextraction. Part II. Acylation of amphetamine with pentafluorobenzoylchloride for urine analysis. *Analyst*, 127, 598-602.
- 164. ZHAO, YY., Cai, LS., Jing, ZZ., Wang, H., Yu, JX., Zhang, HS., 2003. Determination of aliphatic amines using N-succinimidyl benzoate as a new derivatization reagent in gas chromatography combined with solid-phase microextraction. *J. Chromatogr. A*, 1021, 175-181.
- 165. SACHET, F., Lenz, S., Brauch, HJ., 1997. Analysis of primary and secondary aliphatic amines in waste water and surface water by gas chromatography-mass spectrometry after derivatization with 2,4-dinitrofluorobenzene or benzenesulfonyl chloride. *J.Chromatogr. A*, 764, 85-93.
- 166. LU, J., Cwik, M., Kanyok, T., 1997. Determination of paromomycin in human plasma and urine by reversed phase HPLC using 2,4-dinitrofluorobenzene derivatization. *J.Chromatogr. B*, 695, 329-335.
- 167. HEGEDU, H., Gergely, A., Veress, T., Horváth, P., 1999. Novel derivatization with Sanger's reagent (2,4-dinitrofluorobenzene [DNFB]) and related methodological developments for improved detection of amphetamine enantiomers by circular dichroism spectroscopy. *Analusis*, 27, 458-463.
- 168. BJORKLUND, J., Einarsson, S., Engstrom, A., Grzegorczyk, A., Becker, H.D., Josefsson, B., 1998. Automated amino acid determination by high-performance liquid chromatography with 2-(9-anthryl)ethyl chloroformate as precolumn reagent. *J.Chromatogr. A*, 798, 1-8.



- 169. HUSEK, P., 1998. REVIEW: Chloroformates in gas chromatography as general purpose derivatizing agents. *J.Chromatogr. B.*, 717, 57-91.
- 170. ROSENFELD, J.M., 1999. REVIEW: Solid-phase analytical derivatization: enhancement of sensitivity and selectivity of analysis. *J.Chromatogr. A*, 843, 19-27.
- 171. RODRIGUEZ, I., Turnes, M.I., Mejuto, C., Cela, R., 1996. Determination of chlorophenols at the sub-ppb level in tap water using derivatization, solid-phase extraction and gas chromatography with plasma atomic emission detection. *J. Chromatogr. A*, 721, 297-304.
- 172. MEYER, A., Kleiboumlhmer, W., 1995. Determination of pentachlorophenol in leather using supercritical fluid extraction with in situ derivatization. *J.Chromatogr. A*, 718, 131-139.
- 173. LLOMPART, M.P., Lorenzo, R.A., Cela, R., Jocelyn Par, J.R., Brlanger, J.M.R., Li, K., 1997. Phenol and methylphenol isomers determination in soils by in-situ microwave-assisted extraction and derivatization. *J. Chromatogr. A*, 757, 157-164.
- 174. KAWAGUCHI, M., Ito, R., Endo, N., Okanouchi, N., Sakui, N., Saito, K., Nakazawa, H., 2006. Liquid phase microextraction with in situ derivatization for measurement of bisphenol A in river water sample by gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1110, 1-5.
- 175. ALBERICI, R.M.; Sparrapan, R.; Jardim, W.F.; Eberlin, M.N., 2001. Selective trace level analysis of phenolic compounds in water by flow injection analysis-membrane introduction mass spectrometry. *Environ. Sci. Technol.*, 35, 2084-2088.
- 176. KAWAGUCHI, M., Ishii, Y., Sakui, N., Okanouchi, N., Ito, R., Inoue, K., Saito, K., Nakazawa, H., 2004. Stir bar sorptive extraction with in situ derivatization and thermal desorption–gas chromatography–mass spectrometry in the multi-shot mode for determination of estrogens in river water samples. *J.Chromatogr. A*, 1049, 1-8.
- 177. KAWAGUCHI, M., Inoue, K., Yoshimura, M., Sakui, N, Okanouchi, N., Ito, R., Yoshimura, Y., Nakazawa, H., 2004. Trace analysis of phenolic xenoestrogens in water samples by stir bar sorptive extraction with in situ derivatization and thermal desorption–gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1041, 19-26.
- 178. KAWAGUCHI, M.; Ishii, Y.; Sakui, N.; Okanouchi, N.; Ito, R.; Saito, K.; Nakazawa, H., 2005. Stir bar sorptive extraction with in-situ derivatization and thermal desorption-gas chromatography-mass spectrometry for determination of chlorophenols in water and body fluid samples. *Anal. Chim. Acta*, 533, 57-65.
- 179. ITOH, N.; Tao, H.; Ibusuki, T., 2005. Optimization of aqueous acetylation for determination of hydroxy polycyclic aromatic hydrocarbons in water by stirbar sorptive extraction and thermal desorption gas chromatography mass spectrometry. *Anal. Chim. Acta*, 535, 243-250.
- 180. MONTERO, L.; Conradi, S.; Weiss, H.; Popp, P., 2005. Determination of phenols in lake and ground water samples by stir bar sorptive extraction-thermal desorption -gas chromatographymass spectrometry. *J.Chromatogr. A*, 1071, 163.
- 181. KAWAGUCHI, M., Ito, R., Sakui, N., Okanouchi, N., Saito, K., Nakazawa, H., 2005. Dual derivatization–stir bar sorptive extraction–thermal desorption–gas chromatography–mass spectrometry for determination of 17β-estradiol in water sample. *J.Chromatogr. A*, 1105, 140-147.

- 182. STACH, J.; Zimmer, D.; Moder, M.; Herzschuh, R., 1989. Determination of hydroxyaromatic compounds in complex mixtures by tandem mass spectrometry. *Org. Mass Spectrom.*, 24, 946-952.
- 183. LEE, H.B., Peart, T.E., 1998. Determination of 17 beta-estradiol and its metabolites in sewage effluent by solid-phase extraction and gas chromatography mass spectrometry. *J. AOAC Int*, 81, 1209-1216.
- 184. PENG, X., Wang, Z., Yang, C., Chen, F., Mai, B., 2006. Simultaneous determination of endocrine-disrupting phenols and steroid estrogens in sediment by gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1116, 51-56.
- 185. WAHLBERG, C.; Renberg, L.; Wideqvist, U., 1990. Determination of nonylphenol and nonylphenol ethoxylates as their pentafluorobenzoates in water, sewage sludge and biota. *Chemosphere*, 20, 179-195.
- 186. McCALLUM, N.K.; Armstrong, R.J., 1973. The derivatisation of phenols for gas chromatography using electron capture detection. *J.Chromatogr. A*, 78, 303-307.
- 187. BUISSON, R.S.K.; Kirk, P.W.W.; Lester, J.N., 1984. Determination of chlorinated phenols in water, wastewater, and wastewater sludge by capillary GC/ECD. *J. Chromatogr. Sci*, 22, 339-342.
- 188. BIANCHI, F.; Careri, M.; Mucchino, C.; Musci, M., 2002. Improved determination of chlorophenols in water by solid-phase microextraction followed by benzoylation and gas chromatography with electron capture detection. *Chromatographia*, 55, 595-600.
- 189. MEIER, S., Klungsøyr, J., Boitsov,S., Eide, T., Svardal, A., 2005. Gas chromatography–mass spectrometry analysis of alkylphenols in cod (Gadus morhua) tissues as pentafluorobenzoate derivatives. *J.Chromatogr. A*, 1062, 255-268.
- 190. BOITSOV, S.; Meier, S.; Klungsoyr,J.; Svardal, A., 2004. Gas chromatography-mass spectrometry analysis of alkylphenols in produced water from offshore oil installations as pentafluorobenzoate derivatives. *J. Chromatogr. A*, 1059, 131-141.
- 191. AKRE, C.; Fedeniul, R.; MacNeil, J.D., 2004. Validation of a simple, sensitive method for the determination of β -estradiol in bovine urine using gas-chromatography negative-ion chemical ionization mass spectrometry. *Analyst*, 129, 145-149.
- 192. XIAO, XY., McCalley, D., McEvoy, J., 2001. Analysis of estrogens in river water and effluents using solid-phase extraction and gas chromatography-negative chemical ionisation mass spectrometry of the pentafluorobenzoyl derivatives. *J.Chromatogr. A*, 923, 195-204.
- 193. KUCH, H.M.; Ballschmiter, K., 2001. Determination of Endocrine-disrupting phenolic compounds and estrogens in surface and drinking water by HRGC-(NCI)-MS in the picogram per liter range. *Environ. Sci. Technol.*, 35, 3201-3206.
- 194. DASGUPTA, A.; Thompson, W.C.; Malik, S., 1994. Use of microwave irradiation for rapid synthesis of perfluorooctanoyl derivatives of fatty alcohols, a new derivative for gas chromatography mass spectrometric and fast atom bombardment mass spectrometric study. *J.Chromatogr. A*, 685, 279-285.
- 195. SHAREEF, A., Angove, M.J., Wells, J.D., 2006. Optimization of silylation using N-methyl-N-(trimethylsilyl)-trifluoroacetamide, N,O-bis-(trimethylsilyl)-trifluoroacetamide and N-(tert-



butyldimethylsilyl)-N-ethyltrifluoroacetamide for the determination of the estrogens estrone and ethinylestradiol. *J. Chromatogr. A*, 1108, 121-128.

- 196. DING, W.H., Chiang, CC., 2003. Derivatization procedures for the determination of estrogenic chemicals by GC/MS. *Rapid Commun. Mass Spectrom*, 17, 56-63.
- 197. MURAD, I.H., Helaleh, A., Fujii, S., Korenaga, T., 2001. Column silylation method for determining endocrine disruptors from environmental water samples by solid phase micro-extraction. *Talanta*, 54, 1039-1047.
- 198. CARPINTEIRO, J., Quintana, J.B., Rodriguez, I., Carro, A.M., Lorenzo, R.A., Cela, R., 2004. Applicability of solid-phase microextraction followed by on-fiber silylation for the determination of estrogens in water samples by gas chromatography–tandem mass spectrometry. *J.Chromatogr. A*, 1056, 179-185.
- 199. QUINTANA, J.B., Carpinteiro, J., Rodriguez, I., Lorenzo, R.A., Carro, A.M., Cela, R., 2004. Determination of natural and synthetic estrogens in water by gas chromatography with mass spectrometric detection. *J.Chromatogr. A*, 1024, 177-185.
- 200. LI, D., Park, J., Oh, J.R., 2001. Silyl Derivatization of Alkylphenols, Chlorophenols, and Bisphenol A for Simultaneous GC/MS Determination. *Anal. Chem.*, 73, 3089-3095.
- 201. OKEYO, P., Rentz, S.M., Snow, N.H., 1997. Analysis of steroids from human serum by SPME with headspace derivatization and GC/MS. *J. High Resolut. Chromatogr*, 20, 171-173.
- 202. YANG, L., Luan, T., Lan, C., 2006. Solid-phase microextraction with on-fiber silylation for simultaneous determinations of endocrine disrupting chemicals and steroid hormones by gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1104, 23-32.
- 203. HONGA, J.E., Pyoa, H., Park, SJ., Lee, W., 2005. Solid-phase microextraction with on-fiber derivatization for the determination of hydroxy-polychlorinated biphenyl compounds in urine. *Anal. Chim. Acta*, 539, 55-60.
- 204. WOO, K.L., Kim, J.I., 1999. New hydrolysis method for extremely small amounts of lipids and capillary gas chromatographic analysis as N(O)-tert-butyldimethylsilyl fatty acid derivatives compared with methyl ester derivatives. *J.Chromatogr. A*, 862, 199-208.
- 205. LE BIZEC, B., Monteau, F., Gaudin, I., Andre, F., 1999. Evidence for the presence of endogenous 19-norandosterone in humine urine. *J.Chromatogr. B.*, 723, 157-172.
- 206. HERBERER, T., Stan, H.J., 1997. Detection of more than 50 substituted phenols as their tbutyldimethylsilyl derivatives using gas chromatography-mass spectrometry. *Anal. Chim. Acta*, 341, 21-34.
- 207. KELLY, C., 2000. Analysis of steroids in environmental water samples using solid-phase extraction and ion-trap gas chromatography-mass spectrometry and gas chromatography-tandem mass spectrometry. *J.Chromatogr. A*, 872, 309-314.
- 208. MOL, H.G.J., Sunarto, S., Steijger, O.M., 2000. Determination of Endocrine Disruptors in water after derivatization with N-methyl-N-(tert-butyldimethylsilyltrifluoroacetamide) using gas chromatography with mass spectrometric detection. *J. Chromatogr. A*, 879, 97-112.

References



- 209. FINE, D.D., Breidenbach, G.B., Price, T.L., Hutchins, S.R., 2003. Quantitation of estrogens in ground water and swine lagoon samples using solid-phase extraction, pentafluorobenzyl/ trimethylsilyl derivatizations and gas chromatography–negative ion chemical ionization tandem mass spectrometry. *J.Chromatogr. A*, 1017, 167-185.
- 210. GALCERAN, M.T., Moyano, E., Poza, J.M., 1995. Pentafluorobenzyl derivatives for the gas chromatographic determination of hydroxy-polycyclic aromatic hydrocarbons in urban aerosols. *J.Chromatogr. A*, 710, 139-147.
- 211. SANDRA, P.J.F., 1994. Gas Chromatography in *Ullmann's Encyclopedia of Industrial Chemistry*, Hans-Jürgen Arpe (Editor), 5th Edition, VCH, vol B5, 181-236. *ISBN 3527201351*.
- 212. GROB, K., 2001. Split and Splitless Injection for Quantitative Gas Chromatography-Concepts, processes, Practical guidelines, Sources of Error, 4th Edition, Wiley-VCH Verlag GmbH & Co., Weinheim, Germany. ISBN 33527298797.
- 213. ET AL, ChromPack Thermal desorption Cryo-Trapping unit 4020 User Manual.
- 214. ET AL, ThermoDesorptionSystem (TDS) Global Analytical Solutions-Gerstel Application Note (Flyer), <u>www.gerstel.de/application.htm</u>.
- 215. ET AL, 2001. Dilution, Enrichment & Desorption Unit Handbook, Version 1.0, WMA Airsense Analysentechnik GmbH.
- 216. BURGER, B. V.; Snyman, T.; Burger, V. J. G.; van Rooyen, W. F., 2003. Thermal modulator array. *J.Sep.Sci*, 26, 123-128.
- 217. PHILLIPS, J. B; Beens, J., 1999. REVIEW: Comprehensive two-dimensional gas chromatography: a hyphenated method with strong coupling between the two dimensions. *J.Chromatogr. A*, 856, 331-347.
- 218. LIU, Z.; Phillips, J. B., 1994. Sensitivity and Detection Limit Enhancement of Gas Chromatographic Detection by Thermal Modulation. *J.Microcolumn Separations*, 6, 229-235.
- 219. PHILLIPS, J. B, Xu, J., 1995. REVIEW: Comprehensive multi-dimensional gas Chromatography. *J.Chromatogr. A*, 703, 327-334.
- 220. de GEUS, H. J.; de Boer, J.; Phillips, J. B.; Brinkman, U. Th., 1997. Development of a thermal desorption modulator for gas chromatography. *J. Chromatogr. A*, 767, 137-151.
- 221. PHILLIPS, J. B.; Ledford, E. B., 1996. Thermal Modulation: A Chemical Instrumentation Component of Potential Value in Improving Portability. *Field Anal. Chem. Technol.*, 23-29.
- 222. KINGHORN, R. M.; Marriot, P. J., 1997. Longitudinally Modulated Cryogenic System. A Generally Applicable Approach to Solute Trapping and Mobilization in Gas Chromatography. *Anal. Chem.*, 69, 2582-2588.
- 223. KINGHORN, R. M.; Marriot, P. J.; Dawes, P. A., 1998. Longitudinal Modulation Studies for Augmentation of Injection and Detection in Capillary Gas Chromatography. *J.Microcolumn Separations*, 10(7), 611-616.



- 224. GHOOS, Y., Hiele, M., Rutgeerts, P., Vantrappen, G., 1989. Porous -layer open-tubular gas chromatography in combination with an ion trap detector to assess volatile metabolites in human breath. *Biomedical and environmental MS*, 18, 613-616.
- 225. CANCILLA, D.A., Que-Hee, S.S., 1992. REVIEW: O-(2,3,4,5,6-Pentafluorophenyl)methylhydroxylamine hydrochloride: a versatile reagent for the determination of carbonyl-containing compounds. *J.Chromatogr. A*, 627, 1-16.
- 226. MUNCH, J.W., Munch, D.J., Winslow, S.D., 1998. A User's guide to aldehyde analysis using PFBHA derivatisation and GC-ECD detection : Avoiding the pitfalls. *Proc-Water Qual.Technol.Conf.[computer optical disk]*, 898-913.
- 227. VOGEL, A.I., 1989. Aldehyde and amine derivatization reactions with phenylhydrazine and benzaldehyde respectively. *Vogel's Textbook of practical organic chemistry*, 5th Ed revised, Editors Brian Ed Furniss [et al], Longman group, Essex, UK.
- 228. NAMIESNIK, J., 1984. Generation of standard gaseous mixtures. J.Chromatogr. A, 300, 79-108.
- 229. SCARANGELLI, F.P., O'Keefe, A.E., Rosenberg, E., Bell, J.P., 1970. Preparation of known concentrations of gases and vapours with permeation devices calibrated gravimetrically. *Anal. Chem.*, 42, 871-876.
- 230. ROHWER, E. R.; Pretorius, V.; Apps, P. J., 1983. Simple press-fit connectors for flexible fused silica tubing in gas-liquid chromatography. *J. High Resolut. Chromatogr*, 9, 295-297.
- 231. FRANK, H., Renschen, D., Klein, A., Scholl, H., 1995. Trace analysis of airborne haloacetates. J. High Resolut. Chromatogr, 18, 83-88.
- 232. TONG, H.Y., Karasek, F.W., 1984. Flame Ionisation Detector response factors for compound classes in quantitative analysis of complex organic mixtures. *Anal. Chem.*, 56, 2124-2128.
- 233. SCANLON, J.T., Willis, D.E., 1985. Calculation of flame ionization detector relative response factors using the effective carbon number concept. J. Chromatogr. Sci, 23, 333-340.
- 234. Walte, A., 2001. Airsense Analytics Dilution, Enrichment and Desorption Unit Handbook, WMA Airsense Analysentechnik, GmbH: Schwerin, Germany.
- 235. BOESL, U., Heger, H.J., Zimmermann, R., Nagel, H., Püffel, P., 2000. Laser Mass Spectrometry in Trace Analysis. *Encyclopedia of Analytical Chemistry*, R.A. Meyers (Ed.), John Wiley & Sons Ltd, Chichester, 2087-2118.
- 236. ENGELKE, F.; Hahn, J. H.; Henke, W.; Zare, R. N.,1987. Determination of phenylthiohydantoin-amino acids by two-step laser desorption/multiphoton ionization. *Anal. Chem.*, 59, 909-912.
- 237. CLEMETT, S. J.; Maechling, C. R.; Zare, R. N.; Swan, P. D.; Walker, R. M., 1993. Identification of complex aromatic molecules in individual interplanetary dust particles. *Science*, 262, 721-725.

- 238. DALE, M. J.; Jones, A. C.; Pollard, S. J. T.; Langridge-Smith, P. R. R.; Rowley, A. G., 1993. Trace analysis of polyaromatic hydrocarbons in water using multiphoton ionization-membrane introduction mass spectrometry. *Environ. Sci. Technol.*, 8, 1693-1695.
- 239. DALE, M. J.; Downs, O. H. J.; Costello, K. F.; Wright, S. J.; Langridge-Smith, P. R. R.; Cape, J. N., 1995. Direct analysis of polycyclic aromatic hydrocarbons in cloud-water aerosol filtrates using laser desorption mass spectrometry. *Environ. Pollut*, 89, 123-129.
- 240. HAEFLIGER, O. P.; Bacheli, T. D.; Zenobi, R., 1999. Analusius, 27, 337-340.
- 241. FURNISS, B. S.; Hannaford, A. J.; Smith, P. W. G.; Tatchell, A. R., 1989. Vogel's Textbook of Practical Organic Chemistry, 5th ed.; Longman Scientific and Technical: Essex, England., 1258.
- 242. WILLIAMS, B. A.; Tanada, T. N.; Cool, T. A., 1992. In Twenty-Fourth International Symposium on Combustion; The Combustion Institute: Pittsburgh, 1587-1596.
- 243. OOSTERKAMP, A.J., Hock, B., Seifert, M., Irth, H., 1997. Novel Monitoring strategies for xenoestrogens. *Trends in analytical chemistry*, 16, 544.
- 244. ZUO, Y., Zhang, K., 2005. Discussion Suitability of N,O-bis(trimethylsilyl)trifluoroacetamide as derivatization reagent for the determination of the estrogens estrone and 17β-ethinylestradiol by gas chromatography–mass spectrometry. *J.Chromatogr. A*, 1095, 201-202.
- 245. ZHANG, K., Zuo, Y., 2005. Pitfalls and solution for simultaneous determination of estrone and 17β-ethinylestradiol by gas chromatography–mass spectrometry after derivatization with N,O-bis(trimethylsilyl)trifluoroacetamide. *Anal. Chim. Acta*, 554, 190-196.
- 246. DEHENNIN, L., Reiffstock, A., Scholler, R., 1972. J. Chromatogr. Sci., 10, 224.
- 247. XIAO, XY., McCalley, D., 2000.Quantitative analysis of estrogens in human urine using gas chromaotgraphy/negative chemical ionisation mass spectrometry.*Rapid Commun. Mass Spectrom.*, 14, 1991-2001.
- 248. STOCKL, D., Thienpont, L.M., de Brabandere, V.I., de Leenheer, A.P., 1995.Novel Perfluoroacyl derviatives of corticosteroids.*J. Am. Soc. Mass Spectrom.*, 6, 264-276.
- 249. De BRABANDERE, V.I., Thienpont, L.M., Stockl, D., De Leenheerzy, A.P., 1997.13C-NMR and mass spectral data of steroids with a 17,17-dialkyl-l8-nor-l3(14)-ene substructure. *Journal of Lipid Research*, 38, 780.
- 250. SUPELCO®, 1997.Supelco Product Specification Perfluoroacyl anhydrides, Sigma-Aldrich Co., <u>http://www.sigmaaldrich.com/Brands/Supelco_Home.html</u>.
- 251. CROLEY, T.R., Lynn, Jr B.C., 1998. Molecular Ion Stabilization for Enhanced Tandem Mass Spectrometry Through Derivatization for Alkylphenols. *Rapid Commun. Mass Spectrom.*, 12, 171-175.
- 252. STEHMANN, A., Schröder, H.Fr., 2004. Derivatisation of 4-nonylphenol and bisphenol A with halogenated anhydrides. *Water Science and Technology*, 50, 115-118.



- 253. PETERS, D.G., Hayes, J.M., Hieftje. G.M., 1974. Chemical separations and measurements, Theory and practise of analytical chemistry, W.B.Saunders company, P.A., U.S.A.
- 254. MACKAY, D., Varhannickova, D., Ma, Kuo-Ching, Shiu, Wan-Ying, 1994. Chlorophenols and alkylphenols: A review and correlation of environmentally relevant properties and fate in an evaluative environment. *Chemosphere*, 29, 1155-1224.
- 255. JANSSEN, HG., 1998. GERSTEL Sample Introduction Techniques For Capillary Gas Chromatography, Global Analytical Solutions-Gerstel Technical Note Version 3, <u>www.gerstel.de/application.htm</u>.
- 256. SKOOG, D.A., West, D.M., Holler, F.J., 1994. Analytical chemistry An introduction, Saunders College Publishing, Sixth edition, ISBN 003097285X.
- 257. SHURMER, B., Pawliszyn, J., 2000. Determination of Distribution Constants between a Liquid Polymeric Coating and Water by a Solid-Phase Microextraction Technique with a Flow-Through Standard Water System. *Anal. Chem.*, 72, 3660-3664.
- 258. BALTUSSEN, E., Sandra, P., David, F., Janssen, HG., Cramers, C., 1999. Study into the Equilibrium Mechanism between Water and Poly(dimethylsiloxane) for Very Apolar Solutes: Adsorption or Sorption? *Anal. Chem.*, 71, 5213-5216.
- 259. HUNT, G., Pangaro, N., 1982. Potential contamination from the use of synthetic adsorbents in air sampling procedures. *Anal. Chem.*, 54, 369-372.
- 260. MORGAN, E.D., Bradley, N., 1989. Method for analysis of dilute vapours in flue gases and working atmospheres. *J. Chromatogr. A*, 40, 339-344.
- 261. KRUSCEL, B.D.; Bell, R.W.; Chapman, R.E.; Spencer, M.J.; Smith, K.V., 1994. Analysis of ambient polar and non-polar volatile organic compounds(VOC's) by thermal desorption, high resolution gas chromatography-mass spectrometry(TD/HRGC-MS). *J. High Resolut. Chromatogr.*, 17(3), 187-190.
- 262. HERNANDO, M.D.; Mezcua, M.; Gomez, M.J.; Malato, O.; Aguera, A.; Fernandez-Alba, A.R., 2004. Comparative study of analytical methods involving gas chromatography-mass spectrometry after derivatization and gas chromatography-tandem mass spectrometry for the determination of selected endocrine disrupting compounds in wastewaters. *J.Chromatogr. A*, 1047, 129-135.
- 263. BELFROID, A.C., Van Der Horst, A., Vethaak, A.D., Schafer, A.J., Rijs, G.B.J., Wegener, J., Cofino, W.P., 1999. Analysis and occurrence of estrogenic hormones and their glucuronides in surface water and waste water in The Netherlands. *Science of the Total Environment*, 225, 101-108.
- 264. DIAZ, A., Ventura, F., Galceran, M.T., 2002. Simultaneous Determination of Estrogenic Short Ethoxy Chain Nonylphenols and Their Acidic Metabolites in Water by an In-Sample Derivatization/Solid-Phase Microextraction Method. *Anal. Chem.*, 74, 3869-3876.
- 265. RODGERS-GRAY, T.P., Jobling, S., Morris, S., Kelly, C., Kirby, S., Janbakhsh, A., Harries, J.E., Waldock, M., Sumpter, J.P., Tyler, C.R., 2001. Long-Term Temporal Changes in the Estrogenic Composition of Treated Sewage Effluent and Its Biological Effects on Fish. *Environ. Sci. Technol.*, 34, 1521-1528.



- 266. MULLER, S., Moder, M., Schrader, S., Popp, P., 2003. Semi-automated hollow-fibre membrane extraction, a novel enrichment technique for the determination of biologically active compounds in water samples. *J.Chromatogr. A*, 985, 99-106.
- 267. BRAUN, P., Moeder, M., Schrader, S., Popp, P., Kuschk, P., Engewald, W., 2003. Trace analysis of technical nonylphenol, bisphenol A and 17α-ethinylestradiol in wastewater using solid-phase microextraction and gas chromatography-mass spectrometry. *J.Chromatogr. A*, 988, 41-51.
- 268. ROUTLEDGE, E.J., Brighty, G.C., Sumpter, J.P., Waldock, M., 1998. Identification of Estrogenic Chemicals in STW Effluent. 1. Chemical Fractionation and in Vitro Biological Screening. *Environ. Sci. Technol.*, 32, 1549-1557.
- 269. WATABE, Y., Kubo, T., Nishikawa, T., Fujita, T., Kaya, K., Hosoya, K., 2006. Fully automated liquid chromatography–mass spectrometry determination of 17β-estradiol in river water. *J. Chromatogr. A*, 1120, 252-259.
- 270. HU, J., Zhang, H., Chang, H., 2005. Improved method for analyzing estrogens in water by liquid chromatography–electrospray mass spectrometry. *J. Chromatogr. A*, 1070, 221-224.
- 271. PENALVER, A., Pocurull, E., Borrull, F., Marce, R.M., 2002. Method based on solid-phase microextraction–high-performance liquid chromatography with UV and electrochemical detection to determine estrogenic compounds in water samples. *J.Chromatogr. A*, 964, 153-160.
- 272. LOPEZ de ALDA, M.J., Barcelo, D., 2000. Determination of steroid sex hormones and related synthetic compounds considered as endocrine disrupters in water by liquid chromatography–diode array detection–mass spectrometry. *J. Chromatogr. A*, 892, 391-406.
- 273. LOPEZ de ALDA, M.J., Barcelo, D., 2001. Use of solid-phase extraction in various of its modalities for sample preparation in the determination of estrogens and progestogens in sediment and water. *J.Chromatogr. A*, 938, 145-153.
- 274. MITANI, K., Fujioka, M., Kataoka, H., 2005. Fully automated analysis of estrogens in environmental waters by in-tube solid-phase microextraction coupled with liquid chromatography-tandem mass spectrometry. *J.Chromatogr. A*, 1081, 218-224.
- 275. RODRIGUEZ-MOZAZ, S.; Lopez de Alda, M.J.; Barcelo, D., 2004. Picogram per liter level determination of estrogens in natural waters and waterworks by a fully automated on-line solid-phase extraction-liquid chromatography-electrospray tandem mass spectrometry. *Anal. Chem.*, 76, 6998-7006.
- 276. CHIARABARONTI, R. A., D'Ascenzo, G., Dicorcia, A., Gentili, A., Samperi, R., 2000. Monitoring Natural and Synthetic Estrogens at Activated Sludge Sewage Treatment Plants and in a Receiving River Water. *Environ. Sci. Technol.*, 34, 5059.
- 277. MILLER, J.N.; Miller, J.C., 2000. Statistics and chemometrics for analytical chemistry, Fourth edition, Prentice Hall. ISBN 0130228885.
- 278. LEE, W-Y.; Barud-Zubillaga, A., 2005. A Baseline Study on the Occurrence of Organic Wastewater Compounds in the Paso del Norte. Project number: W-04-08, University of Texas at El Paso.



			SURFACE	PORE			
	ADSORBENT	COMPOSITION	AREA	DIAMETER	APPLICATIONS	ADVANTAGES	DISDAVANTAGES
			(m ² /g)	(nm)			
	Activated	Coconut/	800-1000	2.0 / 1.8-2.2	Non-specific i.e. Most	Cheap, efficient,	Polar compounds
	carbon	petroleum based			organic and inorganic	permanent gases	irreversibly adsorbed.
	Anasorb 747	charcoal			compounds. Non-polar,	not adsorbed –	Incomplete desorption.
					polar, reactive and/or	H ₂ , N ₂ , O ₂ , CO,	H ₂ O reduces sorption of
					volatile. Mercury-	CH _{4.} Anasorb	other compounds.
					vapour.	absorbs less H ₂ O	Catalytic activity.
						and desorption	Reacts with oxygen or
ED						efficiencies for	sulphur derivatives.
BAS						polar compounds	
N N						are improved.	
BO	Graphitised	Pre-treated carbon			Non-specific, as above.	No irreversible	High desorption
CAR	carbon black	black under				adsorption sites.	temperatures (400°C)
	Carbotraps	vacuum and inert				No retention of	required.
		gas/ reductive				H ₂ O and low	Tiny particles of carbon
		atmosphere at				molecular mass	can enter desorption
		3000°C				compounds	unit.
						(CO_x, CH_4)	



			SURFACE	PORE			
	ADSORBENT	COMPOSITION	AREA	DIAMETER	APPLICATIONS	ADVANTAGES	DISDAVANTAGES
			(m ² /g)	(nm)			
	Carbon	Thermally			Adsorption of	High capacity	Inefficient retention of
	molecular sieves	decomposed			hydrocarbons and low-	for small volatile	polar compounds.
	Carbosieves	polymer e.g.			boiling C1-C4	molecules.	Solvent with high heat
SE		polyvinyl chloride			hydrocarbons, methyl	Suitable for	of adsorption required
- BA					formate and alkyl	thermal	for displacement of
NO					mercury compounds.	desorption.	adsorbates. H ₂ O can
RB							block cryotrap.
CA							
	Silica gel	Si-OH groups on	100-800	2-4	Polar compounds from	Cooling the	Hydrophilicity
		surface			air. Amines, halogens,	sorbent allows	decreases sorption
	Aluminium	Al ₂ O ₃			oxygen derivatives,	trapping of C1-	capacity.
NIC	oxide				organo-metallics,	C4 hydrocarbons	Thermal desorption
[GA]					MeOH, HCHO and		difficult.
VOR					DMF. Silica gel is often		Silica gel retains H ₂ O
H					used as a substrate for		and CO ₂
					coating with		
					derivatizing reagents.		
1	1				1		1



			SURFACE	PORE			
	ADSORBENT	COMPOSITION	AREA	DIAMETER	APPLICATIONS	ADVANTAGES	DISDAVANTAGES
			(m ² /g)	(nm)			
	Molecular	Zeolites	Varied	Varied	Toxic inorganic	Thermally	Organic compounds are
IC	sieves				compounds. Small	desorbed at	irreversibly adsorbed
AN					conc. of H_2S	240°C/extract	excl. HCHO, acrolein
ORG						with ice H ₂ O	and certain S-
INC							compounds. H ₂ O block
							cryotrap
	Tenax	Poly-2,6-diphenyl-	19	140	Organic bases, neutral	Tenax has a high	Not suited to solvent
		p-phenylene oxide			and high boiling	thermal limit	extraction due to low
RS					compounds.	350-400°C. Ideal	capacity for volatiles
ME					Chlorohydrocarbons.	for thermal	and is incompatible with
JLY					Support for derivatising	desorption.	many solvent systems.
S P(reagents. Broad		
tou					trapping range of		
POR					compounds of varied		
					molecular mass and		
					polarity.		



			SURFACE	PORE			
	ADSORBENT	COMPOSITION	AREA	DIAMETER	APPLICATIONS	ADVANTAGES	DISDAVANTAGES
			(m^2/g)	(nm)			
	XAD-2	Copolymer in	300-400	8.5	Nitroso-compounds and	XAD's,	
	(Amberlite,	which one moiety			polychlorinated	Porapaks and	
	Chromosorb	is styrene or			biphenyls, aromatic,	Chromosorbs	
IER	102)	ethylvinylbenzene			aliphatic nitro-	come in wide	
LYN		and the other			compounds.	ranges of	
ΓΟd	Porapak	monomer a polar	600-650	7.5	Depending on polarity.	polarity.	Polar Porapaks retain
SN		vinyl compound.			Non-polar to polar	Chromosorb 106	H ₂ O and require great
DRO	Chromosorb101,		50	300-400	compounds can be	greater capacity	amount of energy to
P(103, 104, 106,		varied	varied	adsorbed. Chromosorbs	than Tenax,	remove sorbates. Can't
	108.				adsorb inorganic	suited to thermal	withstand the high
					compounds	desorption.	temperature.



Reaction efficiency data

Additional experimental data obtained for results discussed in section 5.4.

Acrolein Reaction Efficiency



Crotonal Reaction Efficiency





Figure A2.1 Reaction efficiency graphs for the on-line derivatization of acrolein and crotonal with phenylhydrazine. The graph displays i) the amount of gas standard released over that time interval as determined by their permeation rate and ii) the amount of analyte gas trapped using *in-situ* derivatization on the SPME fibre as calculated using the internal standard and effective carbon number response for the signal obtained from the GC-FID for the derivative. A comparison of the gradients obtained from the standard and the actual amount of analyte trapped gives an approximation of the reaction/trapping efficiency for this reaction.



Figure A2.2 Reaction efficiency graphs for the on-line derivatization of propylamine and butylamine with benzaldehyde. The graph displays i) the amount of gas standard released over that time interval as determined by their permeation rate and ii) the amount of analyte gas trapped using *in-situ* derivatization on the SPME fibre as calculated using the internal standard and effective carbon number response for the signal obtained from the GC-FID for the derivative. A comparison of the gradients obtained from the standard and the actual amount of analyte trapped gives an approximation of the reaction/trapping efficiency for this reaction.



Confirmation of the alkylphenol-TFA derivatives



Figure A3.1

A) GC-TOFMS chromatogram obtained for the underivatized phenols, TOP $t_R = 17.47$ min, NP $t_R = 20.30$ min and BPA $t_R = 26.25$ min.

B) GC-TOFMS confirmation chromatogram for the trifluoroacetate derivatives prepared in a vial in acetone as described in section 6.2.6. TOP-TFA $t_R = 15.99$ min, NP-TFA $t_R = 19.00$ min and BPA-TFA $t_R = 20.35$ min. *Notice the absence of underivatized phenols*.

The TFA derivatives elute earlier than the underivatized phenols allowing for shorter chromatographic runs, while the mass spectra yields masses higher up in the mass range allowing for improved selectivity during analysis.



232.0677 230.0627 233.0533 00 220 240 260 280 320 340 420 440 460 480 500 520 540 560 580 600

Figure A3.2 GC-TOFMS mass spectrum obtained for the TOP-TFA derivative $t_R = 15.99$ min. M⁺ m/z 302, base peak m/z 231.



Figure A3.3 GC-TOFMS mass spectrum obtained for the NP-TFA derivative $t_R = 19.00$ min. M⁺ m/z 316, base peak m/z 203.



Figure A3.4 GC-TOFMS mass spectrum obtained for the BPA-TFA derivative $t_R = 20.35$ min. M⁺ m/z 420, base peak m/z 405.



Abundance





Figure A3.5 Reconstructed ion chromatgrams for m/z 231 and m/z 203 representing the TFA derivatives of TOP and NP respectively, along with m/z 135 and m/z 213 representing ions for the corresponding underivatized alkylphenols. The PDMS degradation peaks are indicated by the m/z 73 ion trace.

From figure A3.1, the unreacted phenols are expected to elute after the TFA derivatives. There is no clear evidence from the RICs that the underivatized phenols are present. Ions 213 and 135 that are present appear to originate from the PDMS thermal degradation peaks. See section 6.7.1.







Figure A3.6 Reaction efficiencies, determined by placing 1µl 42 ng/µl TOP, 44 ng/µl NP and 54 ng/µl BPA in acetone on the PDMS trap, 5 µl TFAA is added after the solvent has evaporated. The trap is then sealed with glass caps for the duration of the reaction. The reaction appears to be complete after 5 minutes. See section 6.7.2.



Significance test for comparing extraction efficiency from two different PDMS batches

 An F-test is used to compare the population standard deviations between the two batches of PDMS. These need to be the same in order to perform a t-test to compare the two mean results between the batches. The equation used to determine F is shown below [276]:

$$F = \frac{s_1^2}{s_2^2}$$
 A.4.1

Where s_1 and s_2 are the standard deviations for the measurement series and are arranged so that F > 1.

Critical values of F for a two-tailed test (P=0.05) are obtained from table A.4, page 256 [276].Where v1 and v2 are the degrees of freedom (n-1) for the number of measurements made (n) in the respective measurement series.

A t-test can now be used to compare the two mean results between the batches. The variance (s²) needs to be calculated as shown in equation A.4.2, in order to determine t from equation A.4.3 [276].

$$s^{2} = \frac{(n_{1} - 1)s_{1}^{2} + (n_{2} - 1)s_{2}^{2}}{n_{1} + n_{2} - 2}$$

$$k = \frac{(\bar{x}_{1} - \bar{x}_{2})}{s\sqrt{\frac{1}{n_{1}} + \frac{1}{n_{2}}}}$$
A.4.2

Where n is the number of measurements performed in each measurement series and x the average measurement result obtained.

Critical values of t for a two-tailed test (P=0.05) obtained from table A.2, page 254 [276], where t has $(n_1 + n_2 - 2)$ degrees of freedom (v).



Summary of results obtained using the equations as described above:

	ТОР	ТОР	NP	NP	BPA	BPA
	batch 1	batch 2	batch 1	batch 2	batch 1	batch 2
x	70	79	84	43	10	26
S	2.8	2.37	21.8	9.5	1.5	2.08
n	7	5	8	5	8	5
F _{crit}	9.197		9.074		9.074	
F	1.3	96	5.3	30	1.9	023
F < F _{crit}	Population standard		Population standard		Population standard	
	deviation of the two		deviation	of the two	deviation	of the two
	batches a	are equal	batches a	are equal	batches are equal	
t _{crit}	2.23		2.20		2.1	20
t	5.	83	3.9	92	16	.19
t > t _{crit}	Means of the two		Means of the two		Means of the two	
	batches differ		batches differ		batches differ	
	significantly		significantly		significantly	

Table A.4 Summary	of significance test results
-------------------	------------------------------



PDMS MCT trap drying investigation

A few drops of bromothymol blue indicator was added to a 5 ml Milli-Q water sample. The water was sampled at a flow rate of ~ 50 μ l/min through the PDMS MCT. The presence of the bromothymol blue gave a visual indication of water still trapped inside the PDMS channels. These drops of water were best removed by mechanical dropping of the trap, as opposed to purging with gas. Dropping the trap down a 1.5 m length of tube provided enough force to break the capillary action occurring between the water and PDMS walls.

After sampling the PDMS MCT was weighed on a 4 decimal place balance. The trap was weighed after each drying step. A summary of the results obtained for the drying steps performed in series are shown in the graph below. The trap appears to reach a constant mass after purging with hydrogen gas for 2 minutes at a flow rate of 500 ml/min. An equivalent result is obtained by purging for 1 min at 1L/min. The mass difference between the last drying step and thermal desorption of the trap is 0.2 mg. The mass balance performance was not monitored and this mass difference could easily fall within the uncertainty of the balance. However, this mass difference was later assumed to be residual water vapour, even though no water could be visually observed after the last drying step, degradation on the PDMS trap was still observed.

Weighing the trap after plugging with the silica gel caps was not performed as such a small mass cannot be determined accurately on the available mass balance.





Figure A.5 Summary of the drying steps performed in series with the resulting PDMS MCT mass loss achieved from each drying step.



Published article



Anal. Chem. 2005, 77, 1-10

Accelerated Articles

On-Line Derivatization for Resonance-Enhanced Multiphoton Ionization Time-of-Flight Mass Spectrometry: Detection of Aliphatic Aldehydes and Amines via Reactive Coupling of Aromatic Photo Ionization Labels

Maria Fernandes-Whaley,[†] Fabian Mühlberger,[‡] Alexander Whaley,[†] Thomas Adam,[‡] Ralf Zimmermann,^{*,‡,§,⊥} Egmont Rohwer,^{*,†} and Andreas Walte^{II}

Department of Chemistry, University of Pretoria, Pretoria, 0002, South Africa, Institute for Ecological Chemistry, GSF Research Centre, 85764 Oberschleissheim, Germany, Analytische Chemie, Institut für Physik, Universität Augsburg, D-86159 Augsburg, Germany, BlfA-Bayerisches Institut für Angewandte Umweltforschung und -technik GmbH, Abteilung für Umwelt- und Prozesschemie, D-86167 Augsburg, Germany, and Airsense Analytics, Hagenower Str. 73, 19061 Schwerin, Germany.

Resonance-enhanced multiphoton ionization time-of-flight mass spectrometry (REMPI-TOFMS) is a powerful technique for the on-line analysis of aromatic compounds with unique features regarding selectivity and sensitivity. Aliphatic compounds, however, are difficult to address by REMPI due to their unfavorable photo ionization properties. This paper describes the proof of concept for an online derivatization approach for converting nonaromatic target analytes into specific, photoionizable aromatic derivatives that are readily detectable by REMPI-TOFMS. A multichannel silicone trap or poly(dimethylsiloxane) (PDMS) open tubular capillary was used as a reaction medium for the derivatization of volatile alkyl aldehydes and alkylamines with aromatic "photoionization labels" and to concentrate the resulting aromatic derivatives. The aldehydes formaldehyde, acetaldehyde, acrolein, and crotonal, which when underivatized are poorly detectable by REMPI, were converted into their easily photoionizable phenylhydrazone derivatives by the on-line reaction with phenylhydrazine as reagent. Similarly, the methyl-, ethyl-, propyl-, and butylamines were converted into their REMPIionizable benzaldehyde alkylimine derivatives by the online reaction with benzaldehyde as reagent. The derivatives were thermally desorbed from the PDMS matrix and transferred into the REMPI-TOFMS for detection. The **REMPI-TOFMS** detection limits obtained for acetalde-

butylamine using this photo ionization labeling method were in the sub-parts-per-million range and, thus, readily below the permissible exposure limits set by OSHA.

hyde; acrolein; crotonal; and methyl-, ethyl-, propyl-, and

There is an increasing awareness of the harmful effects that volatile aldehydes and amines, particularly formaldehyde, can have on human health. Formaldehyde is classified as a probable human carcinogen by the EPA, OSHA, NIOSH, and the ACGIH.^{1–3} Low-molecular-mass aldehydes and amines are typically eye, nose, and throat irritants.^{3–5} As volatile polar compounds, they are notoriously difficult to analyze. Real time monitoring of these trace

* Direct correspondence to either author. E-mails: erohwer@ postino.up.ac.za, ralf.zimmermann@gsf.de.

[†] University of Pretoria.

[‡] GSF Research Centre.

[§] Universität Augsburg.

 $^{^\}perp$ BlfA-Bayerisches Institut für Angewandte Umweltforschung und -technik GmbH, Abteilung für Umwelt- und Prozesschemie.

Airsense Analytics.

Chemical, Toxicity, Safety and Environmental Analysis Information for Formaldehyde. http://www.instantref.com/formald.htm; Accessed Feb. 2, 2004.

⁽²⁾ Occupational Safety and Health Administration, U.S. Department of Labor. OSHA Permissible Exposure Limits; http://www.osha-slc.gov/pls/oshaweb/; Accessed Feb. 2, 2004.

⁽³⁾ Brabec, M. J. Aldehydes and Acetals. In *Patty's Industrial Hygiene and Toxicology*, 3rd rev. ed.; Clayton, G. D, Clayton, F. E., Eds.; John Wiley & Sons: New York, 1981; Vol 2A, pp 2692–2669.

organic compounds in air or process gases is not easily achieved. Measurement usually requires extended sample preconcentration, cleanup, and instrumental analysis, for example, by gas chromatography/mass spectrometry (GC/MS) in a well-equipped analytical laboratory.^{6–8} It involves a time-consuming and labor-intensive process that prevents the timely data generation required, for example, for effective pollution control measures.

Recently, several on-line monitoring methods based on direct inlet mass spectrometry (MS) with soft and selective ionization methods were established. This includes chemical ionization MS9 as well as photoionization MS techniques.^{10–19} One particularly powerful approach for real time monitoring of aromatic compounds is resonance-enhanced multiphoton ionization time-of-flight mass spectrometry (REMPI-TOFMS). The REMPI-TOFMS method, for example, has been used for the on-line monitoring of dioxin surrogates and other aromatic trace species in waste incinerator emissions,15,16 characterization of the formation of phenolic compounds during coffee roasting,^{17,18} and puff-resolved analysis of toxic aromatic compound release during the cigarette smoking process¹⁹ as well as the characterization of wood combustion.²⁰ In addition to the analysis of gaseous matrixes, solid matrixes can be handled as well in a two-step process using laser desorption followed by REMPI of the volatilized compounds.²¹⁻²⁵ The REMPI process is based on a two-UV-photon absorption/ionization utilizing excited intermediate states (i.e., UV absorption bands) for resonance enhancement. Most aromatic compounds exhibit strong

- (4) Occupational Safety and Health Administration, U.S. Department of Labor. OSHA Permissible Organic Method #36 Ethylamine; http://www. osha-slc.gov/dts/sltc/methods/organic/org036/org036.html; Accessed Feb. 2, 2004.
- (5) Occupational Safety and Health Administration, U.S. Department of Labor. OSHA Permissible Organic method #40 methylamine; http://www. osha-slc.gov/dts/sltc/methods/organic/org040/org040.html; Accessed Feb. 2, 2004.
- (6) Otson, R.; Fellin, P. Sci. Total Environ. 1988, 77, 95-131.
- (7) Kennedy, E. R.; Teass, A. W.; Gagnon, Y. T. Formaldehyde Analytical chemistry and toxicology, Turoski, V., Ed.; Advances in chemistry and toxicology series, 1985, 210, 1–11.
- (8) Berezkin, V. G.; Drugov, Y. S. J. Chromatogr. Libr. 1991, 49, 190-194.
- (9) Munson, B. Anal. Chem. 1977, 49, 772-778A.
- (10) Lubman, D. M., Ed. Lasers and Mass Spectrometry; Oxford University Press: New York, 1990.
- (11) Cool, T. A.; Williams, B. A. Combust. Sci. Technol. 1992, 82, 67.
- (12) Boesl, U.; Zimmermann, R.; Weickhardt, C.; Lenoir, D.; Schramm, K.-W.; Kettrup, A.; Schlag, E. W. *Chemosphere* **1994**, *29*, 1429.
- (13) Miller, J. C.; Compton, R. N. J. Chem. Phys. 1982, 76 (8), 3967.
- (14) Nomayo, M.; Thanner, R.; Pokorny, H.; Grotheer, H.-H.; Stützle, R. *Chemosphere* **2001**, *43*, 461–467.
- (15) Mühlberger, F.; Zimmermann, R.; Kettrup, A. Anal. Chem. 2001, 73, 3590– 3604.
- (16) Heger, H. J.; Zimmermann, R.; Dorfner, R.; Beckmann, M.; Griebel, H.; Kettrup, A.; Boesl, U. Anal. Chem. 1999, 71, 46–57.
- (17) Dorfner, R.; Ferge, T.; Yeretzian, C.; Kettrup, A.; Zimmermann, R. Anal. Chem. 2004, 76 (5), 1386–1402.
- (18) Dorfner, R.; Ferge, T.; Kettrup, A.; Zimmermann, R.; Yeretzian, C. J. Agric. Food Chem. 2003, 51 (19), 5768–5773.
- (19) Zimmermann, R.; Heger, H. J.; Kettrup, A. Fresenius' J. Anal. Chem. 1999, 363, 720–730.
- (20) Hauler, T. E.; Boesl, U.; Kaesdorf, S.; Zimmermann, R. J. Chromatogr., A 2004, in press.
- (21) Engelke, F.; Hahn, J. H.; Henke, W.; Zare, R. N. Anal. Chem. 1987, 59, 909–912.
- (22) Clemett, S. J.; Maechling, C. R.; Zare, R. N.; Swan, P. D.; Walker, R. M. Science 1993, 262, 721–725.
- (23) Dale, M. J.; Jones, A. C.; Pollard, S. J. T.; Langridge-Smith, P. R. R.; Rowley, A. G. Environ. Sci. Technol. 1993, 8, 1693–1695.
- (24) Dale, M. J.; Downs, O. H. J.; Costello, K. F.; Wright, S. J.; Langridge-Smith, P. R. R; Cape, J. N. *Environ. Pollut.* **1995**, *89*, 123–129.
- (25) Haefliger, O. P.; Bacheli, T. D.; Zenobi, R. Analusius 1999, 27, 337-340.

region is easily accessible by commercial laser systems. The combination of selectivity and immediate availability of mass spectral information eliminates the time-consuming separation step of gas chromatography. Unfortunately, many compounds not possessing an aromatic chromophore, such as aliphatic aldehydes and amines, cannot be easily detected by the rather simple one-color two-photon REMPI process. For example, many aldehydes require complicated REMPI schemes, which are based either on multilaser wavelength excitation or the inclusion of nonresonant multiphoton absorption steps. In other cases, as for many amines, the suitable REMPI wavelengths for the various compound homologues are different, preventing a simultaneous detection of the homologue profile.

absorption bands in the

A fast method for the on-line detection of aldehydes and amines, however, would have several potential applications in the field of process gas analysis, ambient air monitoring, or emission analysis. Furthermore, it would be desirable to also make use of the advantages of the REMPI-TOFMS method (i.e., selectivity, sensitivity, and measurement speed) for the detection of these aliphatic compounds. To make aldehydes and amines accessible to REMPI-TOFMS detection, a concept to convert the nonaromatic analytes into specific aromatic derivatives, which would then be detectable by the REMPI-TOFMS, was developed ("photoionization labeling"). Derivatization reactions which in principle can be used for "photoionization labeling" usually are performed in liquid solutions or, as recently demonstrated, in a poly(dimethylsiloxane) (PDMS) matrix as reaction medium. PDMS, for example, has been used for in situ derivatization of low-molecular-mass aldehydes for GC/MS analysis.^{26,27} The work presented here describes the development of a PDMS-based on-line "photoionization labeling" derivatization technique which can be directly hyphenated to the REMPI-TOFMS system. The PDMS devices are shown in Figure 1A. The principle of the "photoionization labeling" derivatization is as follows (depicted in Figure 1B).

The analytes from the sample gas current (i.e., containing traces of amines or aldehydes to be analyzed) as well as the derivatization reagent are coabsorbed in a PDMS trap. After a short enrichment phase, the trap is heated. The heating induces both the derivatization reaction itself and the thermal desorption of the formed derivatives. The desorbed derivatives are subsequently transferred to the REMPI-TOFMS spectrometer for analysis. This procedure can be repeated rapidly for a (quasi) on-line analysis.

At first, potential derivatization reactions were selected (derivatization of aldehydes with phenylhydrazine to form the respective phenylhydrazone derivatives and derivatization of amines with benzaldehyde to form the respective benzaldehyde alkylimine derivatives). The proof of principle (i.e., of efficient PDMS-mediated derivatization) was tested in a solid-phase microextraction (SPME) approach with GC/MS and GC-FID detection. Subsequently, an experimental on-line derivatization setup was built and coupled to the REMPI-TOFMS system. Two different setup variants were used for the derivatization procedure. In the first setup, a thermal modulator array²⁸ with a fused-silica capillary

- (26) Martos, P. A.; Pawliszyn, J. Anal. Chem. 1998, 70, 2311-2320.
- (27) Fernandes, M. J. M.Sc. thesis, University of Pretoria, Pretoria, 2001.
- (28) Burger, B. V.; Snyman, T.; Burger, V. J. G.; van Rooyen, W. F. J. Sep. Sci. 2003, 26, 123–128.



Figure 1. Two variations of silicone (PDMS) concentrators are shown in A, namely, the thick film capillary trap, used in the thermal modulator array (TMA), and the multichannel silicone rubber trap (MCSRT) used in the EDU. B Cross section of a capillary trap, which demonstrates the concentration and reaction within these concentrating devices.

column (3-µm silicone film, DB-1 equivalent) is used to absorb, derivatize, desorb, and refocus the analytes. The second setup consists of an enrichment desorption unit (EDU; Airsense Analytics, Schwerin, Germany)²⁹ with a multichannel silicone rubber trap^{27,30,31} as PDMS medium for derivatization followed by the above-mentioned arrangement with the thermal modulator array.²⁸ These experimental setups are shown in Figure 2A, B, and C, respectively.

EXPERIMENTAL SECTION

(A) Derivatization Reaction for "Photoionization Labeling" of Amines and Aldehydes. Phenylhydrazine^{32,33} and benzaldehyde³³ were selected as "photoionization labeling compounds" and were used to derivatize the aldehydes (formaldehyde, acetaldehyde, acrolein, and crotonal) and amines (methylamine, ethylamine, propylamine, butylamine), respectively. Methylamine, ethylamine, propylamine, benzaldehyde, formaldehyde (36.5% in water), and phenylhydrazine were purchased from Aldrich (Taufkirchen, Germany). Acetaldehyde, acrolein, and crotonal were obtained from Merck (Darmstadt, Germany). Butylamine was obtained from ChemService (Johannesburg, S. A.). *Caution: Because phenylhydrazine is highly poisonous and formaldehyde is a potential carcinogen, it is essential always to wear gloves and avoid inhalation when working with these reagents.*

- (29) Walte, A. Airsense Analytics Dilution, Enrichment and Desorption Unit Handbook, WMA Airsense Analysentechnik, GmbH: Schwerin, Germany, 2001.
- (30) Ortner, E. K.; Rohwer, E. R. J. High Resolut. Chromatogr. 1996, 19, 339– 344.
- (31) Ortner, E. K.; Rohwer, E. R. J. Chromatogr., A 1999, 863, 57-68.
- (32) Vogel, M.; Büldt, A.; Karst, U. Fresenius' J. Anal. Chem. 2000, 366, 781– 791.
- (33) Blau, K.; King, G. S. Handbook of derivatives for chromatography; Heyden & Son Ltd.: London, U.K., 1979.

Schemes for the deriva

UNIVERSITEIT VAN PRETORIA UNIVERSITY OF PRETORIA YUNIBESITHI YA PRETORIA

formed are shown in Figure 3A and B. These reagents were selected to introduce a REMPI-active chromophore to the analyte structure. Substituted rings, such as pentafluorinated benzaldehyde, were discarded because they pose the risk of reducing the REMPI efficiency. In addition, in order for the reaction to occur efficiently, both reagents had to possess a significant vapor pressure to ensure that the reagent would be present in excess in the gas phase. Stable gaseous concentrations of the analytes were obtained by preparing permeation and diffusion gas standards of the respective aldehydes and amines. Gas standard preparation and measurement has been described in the literature.^{34,35} Concentrations provided by the gas standards are listed in Table 1. Headspace from formaldehyde (stabilized with methanol in water) was used as the formaldehyde gas source. This concentration is rather high and could not be determined in the framework of the experiments presented here.

(B) Setup for SPME GC-FID-Based Testing of the PDMS-Mediated Derivatization Reactions. Simple reaction tests were performed to determine whether the selected derivatization reaction would take place in the PDMS and to estimate how efficiently the arrangement would trap the analyte. Figure 2A shows the on-line setup used to determine the approximate reaction efficiency for the various derivatization reactions. The gas standards were purged with nitrogen gas at a flow rate of 4 mL/min. The gas standards provide a known concentration of analyte gas into the glass Y press-fit connector³⁶ (obtained from Chromatography Research Supplies, Inc., Louisville, KY) via an uncoated length of fused-silica capillary. Similarly, the derivatizing reagent, also being purged with nitrogen gas at 4 mL/min, was introduced at the other end of the Y press-fit connector. A 1-mL portion of the derivatizing reagent was placed in a 2-mL vial and sealed with a crimp cap. Two holes were pierced into the septum of the vial. A length of uncoated fused-silica capillary was pushed through each hole in the septum. One capillary was connected to the nitrogen gas, the other to the Y press-fit connector. Leading from the combined exit of the Y press-fit connector was another length of uncoated fused-silica capillary. The measured flow rate at this point was 8 mL/min, similar to the flows obtained from the REMPI-TOFMS vacuum. The exiting capillary was sealed into another glass press-fit connector, the opposite end of which was modified to house the exposed SPME fiber.

Current concentration methods are mainly off-line.^{37–39} Solidphase microextraction (SPME)^{26,40} and the multichannel silicone rubber trap (MCSRT)²⁷ are two examples of a novel technique that uses poly(dimethylsiloxane) as the concentration and reaction medium, eliminating problems experienced with earlier concentration methods.^{27,37–39} In situ derivatization in PDMS has been used to trap low-molecular-mass aldehydes for GC-FID and GC/MS analysis.^{26,27} The PDMS concentrators used in this study are

- (35) Scarangelli, F. P.; O'Keefe, A. E.; Rosenberg, E.; Bell, J. P. Anal. Chem. 1970, 42, 871–876.
- (36) Rohwer, E. R.; Pretorius, V.; Apps, P. J. J. High Resolut. Chromatogr. 1983, 9, 295–297.
- (37) Berezkin, V. G.; Drugov, Y. S. J. Chromatogr. Libr. 1991, 49, 35-119.
- (38) Namiesnik, J. Talanta 1988, 35, 567-587.
- (39) Stashenko, E. E.; Ferreira, M. C.; Sequeda, L. G.; Martinez, J. R.; Wong, J. W. J. Chromatogr., A 1997, 779, 360–369.
- (40) Pawliszyn, J. Solid-Phase Microextraction Theory and Practice; Wiley-VCH: Canada, 1997.

⁽³⁴⁾ Namiesnik, J. J. Chromatogr., A 1984, 300, 79-108.



Figure 2. Experimental setup used for (A) determining the reaction efficiencies for the on-line derivatization reactions, (B) on-line concentration and derivatization for REMPI-TOFMS using the thermal modulator array (TMA) with a thick-film capillary as enrichment and reaction medium, and (C) on-line concentration and derivatization for REMPI-TOFMS using a MCSRT in an EDU as enrichment and reaction medium and the TMA with a thick film capillary for analyte modulation.

Table 1. Gas Standard Concentrations and Calculated Detection Limits for	or the Aldehydes and Amines Studied ^a
--	--

analytes (m/z)	gas std concn EDU-TMA (ppm v/v)	detection limit (S/N = 2 av 10) EDU-TMA (ppm v/v)	gas std concn TMA (ppm v/v)	detection limit (S/N = 2 av 10) TMA (ppm v/v)	PEL OSHA (ppm)
formaldehyde (120)					0.75
acetaldehyde (134)			79.4	2.04	200
acrolein (146)			37.4	0.101	0.1
crotonal (160)			199	1.52	2
methylamine (119)	34.3	0.257			10
ethylamine (133)	1.4	0.010	21.7	0.324	10
propylamine (147)	1.8	0.024	27.6	0.138	
butylamine (161)	2.9	0.100	44.7	0.501	5
^a Permissible exposure 1	limits (PEL) as set by	OSHA are also listed (see re	ef 2).		

depicted in Figure 1A. The SPME device consists of a 100-µm PDMS-coated fiber mounted on the tip of a syringe needle, which is housed within the syringe barrel when not exposed during sampling.40 A 100-um PDMS SPME fiber was exposed over increasing time intervals to a similar on-line arrangement used for the REMPI-TOFMS shown in Figure 2B. The SPME assembly and 100-µm PDMS fibers were obtained from Supelco (Bellefonte, PA). The fiber was desorbed in the heated inlet of a Varian 3300 GC at 150°C for 1 min. Quantitation was performed by flame ionization detection (FID) using undecane as internal standard and relative effective carbon number responses of the derivatives.^{27,41,42} Thermal desorption of the SPME fiber is performed simply and quickly in the heated inlet of the GC oven; however, desorption of the silicone trap requires a desorption unit with some form of cooling in order to focus the desorbed contents onto the GC column. This is usually a longer process.²⁷ When the above procedure is carried out in GC-FID or GC/MS, the low initial temperature of the GC oven also acts to focus or concentrate the derivatized analyte in a short band. For real-time on-line applications, in the absence of such a focusing mechanism in the direct coupling of the trap to the TOFMS, another concentration device is required to enhance detectability. The results for this experiment are shown in Figure 4.

(C) **REMPI-TOFMS.** The resonance-enhanced multiphoton ionization time-of-flight mass spectrometer used for this application is a home-built system containing a pulsed Nd:YAG laser (Quanta-Ray INDI 50; Spectra Physics, Stratford, CT). The initial 1064-nm laser beam (repetition rate 10 Hz, pulse duration 10 ns) is frequency tripled, and the resulting wavelength of 355 nm is used to pump a β -BBO crystal of a thermally stabilized type II OPOlaser system (GWU-Lasertechnik, Germany) to generate wavelength-tuneable laser pulses in the range of 220 nm to 2.5 μ m. The generated laser pulses (~10⁶ W cm⁻²) are directed into the ionization chamber of the TOF (Kaessdorf Instruments, Germany)

⁽⁴¹⁾ Scanlon, J. T.; Willis, D. E. J. Chromatogr. Sci. 1985, 23, 333–340.
(42) Tong, H. Y.; Karasek, F. W. Anal. Chem. 1984, 56, 2124–2128.



Figure 3. Reaction schemes for the derivatization of (A) the aldehydes with phenylhydrazine and (B) the alkylamines with ben-zaldehyde.

underneath the jet capillary inlet by optical elements. Molecular ions formed are accelerated and extracted into the flight tube of the reflectron TOFMS. Mass spectra are recorded via a transient recorder PC card (Aquiris, Switzerland, 250 MHz, 1 GS/ s, 128 k) whereby data processing is done by LabView (National Instruments, Austin, TX)-based home-written software. Wavelengths of 244 and 246 nm were selected for REMPI-TOFMS analysis of the formaldehyde– and acrolein–phenylhydrazone derivatives, respectively, and 240 nm for the benzaldehyde alkylimine derivatives. Spectroscopic investigations showed that for the REMPI-TOFMS setup used, these wavelengths are very efficient.

(D) On-Line Derivatization Setup for REMPI-TOFMS. Figure 2B and C shows the on-line derivatization REMPI-TOFMS setups. Unlike the arrangement for principal testing (SPME), the gas standards and reagents were not purged with nitrogen gas. In this case, the mass spectrometer vacuum provides the flow into the REMPI-TOFMS. On-line in situ derivatization was investigated using two different enrichment desorption devices:



trap (TMA) and (ii) an enrichment desorption unit with a multichannel PDMS rubber trap. Two setup variants were tested. In the first setup, only the thermal modulator array (i) with a PDMS thick-film capillary trap was used, whereas in the second setup, the enrichment desorption unit with a multichannel PDMS rubber trap (ii) was applied in combination with the thermal modulator array with a PDMS thick-film capillary trap (i).

(i) a thermal modulator a

In the following, the two experimental setups are described in more detail.

First Setup: Direct Supply of Analytes and Reagents through the Thermal Modulator Array (TMA-REMPI-TOFMS). The centerpiece of the derivatization setup is the segmented thermal modulator array.²⁸ The modulator houses a narrow bore capillary coated on the inside with a thick film of PDMS (capillary trap). This capillary represents the concentrating/derivatizing device. The amount of PDMS within the capillary is comparable to the amount of PDMS forming the SPME fiber. In detail, the modulator capillary consisted of a fused-silica capillary column (0.2-mm i.d.) coated with nonpolar phase PS-255 (3- μ m film, DB-1 equivalent). A capillary of 20-cm length was used with 5 cm of the stationary phase stripped off on either end, as described in reference 28.

A stainless steel capillary (105 mm \times 0.6 mm o.d. \times 0.35 mm i.d.) was converted to a modulator.²⁸ An electronic sequencer was used to provide current to the modulator in steps from 1 to 10 A at 5 V with a time duration of 10–2500 ms. To maintain reasonable flow rates and operate at atmospheric pressure, jet restrictors yielding a flow rate of between 0.6 and 1.0 mL/min were prepared according to the method described in reference 43 from an uncoated capillary (30 cm \times 0.32 mm i.d.). The restrictor was coupled to the modulator capillary with a suitable press-fit. All transfer capillaries and connection points were either directly heated to 150 °C, by a heating mantle or surrounded by a copper tube, which was then heated by a heating mantle.

Modulators have predominantly been developed for use as an interface between two columns in comprehensive two-dimensional gas chromatography.⁴⁴ Its function is to rapidly focus fractions of effluent from the first column onto the head of the second column. In this work, a modulator is used to transfer and focus the effluent from the capillary trap into the REMPI-TOFMS.

In principle, the sorption and desorption of effluent from the stationary phase in the modulator capillary can be controlled by careful manipulation of the capillary temperature. This was originally achieved by painting a segment of the modulator capillary with an electrically conductive paint, thus allowing the capillary to be resistively heated.^{45,46} This modulator was tedious to prepare and did not prove robust. Alternatively, a copper wire could be coiled around the modulator capillary.⁴⁷ A mechanically driven thermal sweeper was developed to eliminate the high thermal inertia experienced by the metal painted modulator.⁴⁸ A moveable slotted heating element was used to "sweep" periodically over the modulator capillary. This design demonstrates good

- (44) Phillips, J. B; Beens, J. J. Chromatogr., A 1999, 856, 331.
- (45) Liu, Z.; Phillips, J. B. J. Microcolumn Sep. 1994, 6, 229.
- (46) Phillips, J. B, Xu, J. J. Chromatogr., A 1995, 703, 327.
- (47) de Geus, H. J.; de Boer, J.; Phillips, J. B.; Brinkman, U. Th. J. Chromatogr., A 1997, 767,137.
- (48) Phillips, J. B.; Ledford, E. B. Field Anal. Chem. Technol. 1996, 1, 23.

⁽⁴³⁾ Hafner, K.; Zimmermann, R.; Rohwer, E. R.; Dorfner, R.; Kettrup, A. Anal. Chem. 2001, 73, 4171–4180.



Figure 4. Reaction efficiency results for the on-line derivatization of (A) formaldehyde and (B) acetaldehyde with phenylhydrazine. Both graphs display (i) the amount of gas standard released over that time interval, as determined by their permeation rate, and (ii) the amount of analyte gas trapped using in situ derivatization on the SPME fiber as calculated using the internal standard and effective carbon number response for the signal obtained from the GC-FID for the derivative. The graphs on the right-hand side represent an enlargement of the left-hand side graphs, where the initial accumulation on the SPME fiber appears linear. A comparison of the gradients obtained from the standard and the actual amount of analyte trapped gives an approximation of the reaction/trapping efficiency for this reaction.

temperature control and focusing, but is too bulky and complex, requiring the heating element to be at least 100°C higher in temperature than the capillary to effectively focus the effluent.⁴⁵ Thus, much attention has been given to cryogenic modulators. A longitudinal modulating cryogenic system^{49,50} consists of a moveable steel sleeve, which surrounds the capillary. Liquid CO_2 is supplied at timed intervals into the sleeve to cool the capillary. The GC oven provides heating to the capillary segments not being cooled by the moving sleeve. A similar approach in which the CO₂ is spraved directly onto the capillary⁵¹ was also used; however, contact of the moving modulator with the second column often causes column breakage. Therefore, a nonmoving dual jet cooling modulator was developed. Two different types exist: the first uses two nonmoveable CO_2 jets to cool the capillary trap while the GC oven is used for heating.52 The second, from the ZOEX Corporation, uses two cold and two warm nonmoveable nitrogen jets to cool and reinject the effluent from the capillary into the second column.⁵³ Although the cryomodulators provide excellent refocusing of effluent, they require expensive cryogens that require attention when in use.

The thermal modulator array²⁸ is an improved combination of the metal-painted and "sweep" modulators described above. Rapid resistive heating of consecutive segments of a stainless steel tube surrounding the capillary focuses the effluent inside the modulator capillary. This provides the "sweeping" heat motion without the disadvantageous cold spots or moveable parts. The segmented heating of the effluent in the capillary speeds up the chromatographic process in the capillary column, "compressing" zones from the rear and providing a focused chromatographic band that enters the REMPI-TOFMS. Although not providing the shortest injection pulse widths, the TMA is simple and compact; it does not require cryogenic cooling and can operate unattended, making it suitable for on-line analysis with the REMPI-TOFMS.

The outlet of the TMA device was directly coupled to the TOFMS . This setup was tested for detecting amines using benzaldehyde as photoionization labeling compound. Reagent and analytes (amine gas standard) were introduced simultaneously for 10 min into the cooled, PDMS, narrow bore, thick-film capillary trap (inside the modulator steel tube) where the reaction occurred. In this case, the MS vacuum provided a sampling flow rate of 0.7 mL/min. During modulation, the derivatives were desorbed into the REMPI-TOFMS. Similarly, the derivatization of the aldehydes

⁽⁴⁹⁾ Kinghorn, R. M.; Marriot, P. J. Anal. Chem. 1997, 69, 2582.

⁽⁵⁰⁾ Kinghorn, R. M.; Marriot, P. J.; Dawes, P. A. J. Microcolumn Sep. 1998,

 ^{(7), 611.} Beens, J.; Delluge, J.; Adahchour, M.; Vreuls, R. J. J.; Brinkman, U. Th. J. Microcolumn Sep. 2001, 13 (3) 134.

⁽⁵²⁾ Beens, J.; Adahchour, M.; Vreuls, R. J. J.; van Altena, K.; Brinkman, U. Th. J. Chromatogr., A 2001, 919, 127.

⁽⁵³⁾ Ledford, E. B., Jr.; Billesbach, C. J. High Resolut. Chromatogr. 2000, 23, 202.



Figure 5. Mass spectra obtained for the formaldehyde-phenylhydrazone derivative using two different ionization techniques. The EI mass spectrum was obtained from a prepared derivative on an accurate mass GC/TOFMS. The REMPI mass spectrum at 244 nm was obtained from the on-line concentration and derivatization experiment (using the TMA setup).

with phenylhydrazine was demonstrated using only the modulator trap, followed by REMPI-TOFMS detection. The results obtained with the TMA-REMPI-TOFMS setup are given in Table 1 and Figures 5 and 6.

Second Setup: Supply of Analytes and Reagents to an Enrichment Desorption Unit prior to the TMA (EDU-TMA-REMPI-TOFMS). The second setup used is as shown in Figure 2 C. Here, the multichannel silicone rubber trap in the enrichment desorption unit is used as concentration—reaction medium, and the TMA is used for subsequent temporal focusing. The multichannel silicone rubber trap consists of a glass tube containing several smaller silicone rubber tubes, each 10 cm long, arranged in parallel,^{27,30,31} as shown in Figure 1A. SIL-TEC medical grade silicone tubing for the silicone rubber trap was obtained from Technical Products Inc. (Georgia, U.S.A). It has been shown that the MCSRT can be used as an *inert* absorptive (off-line) concentrator^{27,30,31} having a very low pressure drop (or flow resistance) with properties similar to the packed PDMS trap,^{54–56} which has demonstrated better properties than other current off-line concentration methods.

The MCSRT is placed within the enrichment desorption unit that is connected via the TMA to the REMPI-TOFMS (EDU-TMA-REMPI-TOFMS). The EDU is an automated stand-alone sampling





Figure 6. REMPI-TOF mass spectra obtained for the on-line concentration and derivatization of (A) acetaldehyde, acrolein, and crotonal with phenylhydrazine at 246 nm (using the TMA setup) and (B) methylamine, ethylamine, propylamine, and butylamine with benzaldehyde at 240 nm (using the EDU-TMA setup).

and desorption device (Airsense Analytics, Schwerin, Germany). The principal difference between SPME (or the application of TMA solely) and MCSRT is the amount of PDMS available for concentration of analytes, with the MCSRT having a considerably larger amount of PDMS (approximate PDMS volumes are TMA trap 0.2 mm³ and the MCSRT 135 mm³). Thus, the MCSRT can concentrate and derivatize more analyte and, therefore, has the potential to provide lower detection limits.

The EDU system used in this work is a unique trap and thermal desorption system developed by Airsense Analytics for the Institute of Ecological Chemistry, GSF. Gaseous substances are trapped at sampling temperatures (ambient or less) on, for example, Tenax adsorption tubes and analyzed after thermal desorption. The enrichment factor is related to many different physical and sampling parameters. It can be calculated on the basis of breakthrough volumes known from common tables. Typically, the detection limit can be reduced by a factor of 20 with volatile compounds and up to 1000 with low volatiles. Temperatures of the adsorbent during sampling and desorption phases can be adjusted via settings within the related software EDU.

⁽⁵⁵⁾ Baltussen, E.; David, F.; Sandra, P.; Janssen, H. G.; Cramers, C. A. J. High Resolut. Chromatogr. 1998, 21, 332–340.

⁽⁵⁶⁾ Baltussen, E.; den Boer, A.; Sandra, P.; Janssen, H. G.; Cramers, C. A. Chromatographia 1999, 49, 520–524.

For increasing the speed of analysis, very small tubes with inner diameters of 1.5 mm filled with Tenax-TA can be used. With applications in very damp environments, this hydrophobic polymer is advantageous because it eliminates the negative influence of humidity on the analysis.

Peltier cooling is used in order to achieve sampling temperatures of 4 °C. After sampling, the tubes are desorbed by resistive heating. With this flash desorption, temperature increments of 200 °C are possible in just 4 s. The complete system is controlled by a microprocessor unit, which is programmed through a serial port.

By sucking air through a cold adsorption tube, the analytes are trapped. In the case of sampling hot gases, it is also possible to dilute the sampling gas to reduce the temperature of the gas. After sampling, a postsampling step is possible to sweep away noninteresting gases and vapors (e.g., humidity).

To extract analytes off the trap, thermal desorption is performed. For injection, the gas flow is reversed and leads into the detection system. Afterward, the tube is cleaned by heating it to a higher temperature than the desorption temperature and flushing the tube with cleaned air. After cooling to near ambient temperatures, the trap is ready for the next measurement. All analytical steps, sampling, postsampling, desorbing, injecting, cleaning, and cooling are performed automatically.

For in situ derivatization, illustrated in Figure 1B, the aromatic derivatizing reagent, in the gas phase, dissolves into the PDMS. Carbonyl compounds (aldehydes and ketones etc.), which pass through the trap, react selectively with the reagent and remain in the trap until they are thermally desorbed for analysis.^{26,27} In the case of the above-mentioned SPME-GC-FID approach, the desorption is performed for some time in the heated GC injector, ^{26,27,40} and the derivatized analytes are refocused in a short band due to low initial temperature of the GC oven. For on-line realtime analytical applications, however, analyte focusing can also be important, although not for the enhancement of the chromatographic resolution, but for time resolution and sensitivity. Analyte focusing can be achieved, as described in the first setup, by repetitive thermal modulation. Therefore, in this setup, the EDU is used in combination with a segmented thermal modulator array, as described above. Conditions for the EDU used in these experiments were as follows: sampling for 130 s at 6 °C with a sampling flow rate of 230 mL/min and thermal desorption for 60 s at 180 °C. Injection occurs under reversed flow conditions. During injection, the desorbed compounds are drawn into the REMPI-TOFMS at a flow rate of 15 mL/min, as restricted by the capillary jet leading into the ion source. Both the sampling line and the transfer line into the MS are heated at 150°C. Benzaldehyde was sampled for 60 s through the heated sampling line. After 10 s, the amine gas mixture was sampled through the sampling line for 60 s. The sampling flow rate was 230 mL/min. Benzaldehyde accumulates in the PDMS multichannel trap, cooled to 6 °C. The introduced amine gas subsequently reacts with the benzaldehyde in the trap. The reaction is further encouraged during desorption at 150 °C for 1 min. During the injection phase, the derivatives are transferred to the TMA, which submits timely focused concentrated pulses to the REMPI-TOFMS system. The results obtained with the EDU-TMA-REMPI-TOFMS setup are given in Table 1 and Figure 6.



Table 2. Approximation of the LUNBESTHEIT A PRETOR Reaction Efficiencies at Room Temperature without Catalyst, as Determined by SPME Setup (see Figure 2A)

compound	reagent	% reaction efficiency	R^2 (n)
formaldehyde	phenylhydrazine	41	$\begin{array}{c} 0.9579 \ (4) \\ 0.9404 \ (4) \\ 0.9990 \ (4) \\ 0.9251 \ (4) \\ 0.9570 \ (4) \\ 0.9205 \ (4) \end{array}$
acetaldehyde	phenylhydrazine	92	
acrolein	phenylhydrazine	61	
crotonal	phenylhydrazine	74	
propylamine	benzaldehyde	28	
butylamine	benzaldehyde	28	

RESULTS AND DISCUSSION

In the first experiments, the reaction efficiency of the selected derivatization reagents with the selected analytes was tested with the SPME GC-FID approach. The reaction efficiency graphs shown in Figure 4 for the on-line derivatization of formaldehyde and acetaldehyde with phenylhydrazine display the increasing mass accumulation of derivative on the SPME fiber over time. Both graphs display (i) the amount of gas standard released over that time interval, as determined by their permeation rate, and (ii) the amount of analyte gas trapped using in situ derivatization on the SPME fiber, as calculated using an internal standard and the effective carbon number response for the signals obtained from the GC-FID for the desorbed derivatives.^{27,41,42} The graphs on the right represent an enlargement of the graphs on the left, where the initial accumulation on the SPME fiber appears linear. A comparison of the initial gradients obtained from the analyte standard and the actual amount of analyte trapped gives an approximation of the reaction/trapping efficiency for this reaction.²⁷ The flattening off of the accumulation curves over time is the result of increased loss or "breakthrough" of the reaction product from the SPME fiber concentrator. The reaction efficiency data, shown in Table 2, were obtained at room temperature using the arrangement in Figure 2A. In Table 2, approximate reaction efficiencies of 28% for the reaction of propylamine and butylamine with benzaldehyde, 30% for the formaldehyde reaction with phenylhydrazine, and around 70% for the aldehydes with phenylhydrazine are indicated.

Incomplete reaction was confirmed by single photon ionization time-of-flight mass spectrometry (SPI-TOFMS)^{15,42} of the on-line, in situ derivatization of propylamine (59 m/z) and butylamine (m/z73) with benzaldehyde $(m/z \ 106)$. The presence of both derivatized (161 and 147 m/z) and underivatized analyte (59 and 73 m/z) was observed. Although these derivatization reactions are not 100% efficient at room temperature, they still occur readily without the aid of any catalysts. Thus, for quantitation, the use of internal or external standards is required. The results of the on-line tests with REMPI-TOFMS detection are given below. The experiments demonstrated that all investigated amines and aldehydes could be successfully derivatized, desorbed, and identified by REMPI-TOFMS using the on-line setups described above. Figure 5 displays the results obtained for formaldehyde. In the upper part (A), a conventional 70-eV EI mass spectrum for the formaldehydephenylhydrazone derivative is shown. This mass spectrum was obtained from a formaldehyde-phenylhydrazone derivative, prepared using the method described by Vogel et al.,⁵⁸ on an accurate mass TOFMS (Micromass, GCT, U.K.). The formaldehyde– phenylhydrazone derivative is detected at 120 m/z, together with its H loss of similar intensity (119 m/z). The base peak of the spectrum, however, is due to the C₆H₅NH⁺ fragment at 92 m/z. The peak at 93 m/z is probably due to C₆H₅NH₂⁺ formed in a rearrangement. Figure 5 also displays the REMPI mass spectrum (244 nm, averaged over 10 transients) obtained from the equivalent on-line derivatization reaction of formaldehyde using the TMA-REMPI-TOFMS setup described above. The soft ionization capability of REMPI provides simple mass spectra with nearly no fragmentation. The mass peak 94 m/z in the REMPI spectrum is suspected to be due to an impurity in the phenylhydrazine reagent (most likely phenol).

Figure 6 shows the REMPI-TOF mass spectra obtained for the TMA and EDU-TMA on-line derivatization of the aldehydes (A) and the amines (B), respectively. The REMPI mass spectrum of the aldehyde derivatives at 246 nm, Figure 6A, displays the [M -1] and [M - 2] mass peak for the acrolein-phenylhydrazone derivative (145 and 144 m/z). [M - 1] corresponds to the loss of a hydrogen atom and [M - 2] to the loss of two hydrogen atoms. The [M - 2] signal is off-scale. These peaks were also observed on the electron impact (EI) mass spectrum of the derivative (not shown here). Only the [M - 1] peak was observed for the acetaldehyde-phenylhydrazone derivative (133 m/z). The crotonal phenylhydrazone was detected as a [M + 1] peak (161 m/z). Additionally, only a very weak [M - 2] peak is visible (158 m/z). [M + 1] adduct peaks commonly are visible in chemical ionization mass spectra, also to a lesser extent in EI mass spectra obtained from ion trap mass spectrometers, when some unintentional chemical ionization can occur. However, [M + 1] peaks do not occur in photoionization TOF mass spectra under the chosen conditions (i.e., a pressure of 10^{-4} mbar in the ion source, avoiding protonation via ion-molecular reaction). The strong [M + 1] peak for crotonal phenylhydrazone, thus, is unexpected and indicates that most likely a side reaction has occurred during the derivatization. Because phenylhydrazine, like hydrazine, is a reducing agent, one possible explanation is the hydrogenation of the double bond of crotonal (either before or after the derivatization). The resulting derivative would be but anal phenylhydrazone (162 m/z), which may be detected as an [M - 1] peak (161 m/z), as found for the acetaldehyde and acrolein derivatives. However, it remains unexplained at the current level of research why the same hydrogenation does not take place for acrolein. If we summarize the result for the aldehydes, it can be stated that only formaldehyde can be detected at the unfragmented derivative mass [M] of 120 m/z. The other aldehyde derivatives, however, were identifiable at either the respective [M - 1] or [M - 2] peak ([M + 1] for crotonal). The molecular ion [M] for acetaldehvde, acrolein, and the crotonal phenylhydrazone were not observed at the applied REMPI wavelength of 246 nm. An EI mass spectrum of the acrolein phenylhydrazone product, however, clearly shows the



UNIVERSITEIT VAN PRETORIA UNIVERSITY OF PRETORIA YUNIBESITHI YA PRETORIA

146 in Figure 6A is due to the ¹³C isotope peak for the [M - 1] ion, not the molecular ion). This indicates that for higher aldehyde-phenylhydrazone derivatives, a photoinduced fragmentation is observable, which is, however, not a problem for the analytical application because the mass spectra are still very soft; i.e., only one (or two) peak(s) dominate the spectra. Phenylhydrazine itself was not observed at the used REMPI wavelength. Note, with other REMPI wavelengths or power densities, different relative sensitivities or photoinduced fragmentation activities for the different aldehydes may be observed.

The REMPI mass spectrum (240 nm) of the amine derivatives is shown in Figure 6B. Benzaldehyde–methylimine, –ethylimine, –propylimine, and –butylimine display two mass peaks of similar intensities, [M] and [M – 1], corresponding to the molecular ion and the hydrogen atom loss. This trend was also observed on the EI mass spectra. In addition, the derivatizing reagent, benzaldehyde, is also observed in the mass spectrum (Figure 6B). The signal [M] m/z 106 is off-scale. The [M + 1] m/z 107 peak is, therefore, the ¹³C isotope peak. The presence of m/z 106 confirms that the reagent is present in excess during the on-line reaction. A mass gate is required during on-line derivatization when an excessive quantity of reagent, such as benzaldehyde, is present to deflect these ions from the detector. The mass gate will prevent "blinding" of the detector to masses occurring after 106 mass units (the mass of benzaldehyde).

To summarize, the REMPI detectability of the amine derivatives is as successful as for the aldehydes: all analytes were detected as [M] and [M - 1] with no further fragments.

Detection limits were determined and are summarized in Table 1. They were calculated using the combined method of Heger et al.¹⁶ and Williams et al.,⁵⁹ using a S/N of 2 and an average of 10 mass spectra. These results demonstrate the potential of this technique in future applications. The calculated detection limits for the analytes are markedly below permissible exposure limits set by the Occupational Safety and Health Administration (OSHA).²

The EDU, constructed specifically for use with the on-line REMPI-TOFMS, allows for the use of a multichannel silicone rubber trap for preconcentration. Lower detection limits were achieved with this setup, since more PDMS is available for preconcentration. This is confirmed by the results obtained for the benzaldehyde-methylimine, -ethylimine, -propylimine, and -butylimine derivatives using the EDU-TMA and the TMA, respectively (see Table 1). In addition, off-line sampling together with a portable pump is also made possible, since the MCSRT trap is easily removed from the EDU.

CONCLUSIONS

The work presented here, on one hand, demonstrates that online derivatization concepts can be used to expand the unique online analytical properties of the resonance-enhanced multiphoton ionization time-of-flight mass spectrometer to aliphatic compound classes. In detail, a method for on-line in situ derivatization of alkylamines with benzaldehyde and alkyl aldehydes with phenylhydrazine followed by thermal desorption and detection by the REMPI-TOFMS was successfully tested. The detection limits obtained for all analytes, for which concentration standards were made, are below the permissible exposure limits set by OSHA. Formaldehyde, which is not easily detected by mass spectrometry,

⁽⁵⁷⁾ Pallix, J. B.; Schuhle, U.; Becker, C. H.; Huestis, D. L. Anal. Chem. 1989, 61, 805–811.

⁽⁵⁸⁾ Furniss, B. S.; Hannaford, A. J.; Smith, P. W. G.; Tatchell, A. R. Vogel's Textbook of Practical Organic Chemistry, 5th ed.; Longman Scientific and Technical: Essex, England, 1989; p 1258.

⁽⁵⁹⁾ Williams, B. A.; Tanada, T. N.; Cool, T. A. In *Twenty-Fourth Symposium (International) on Combustion*; The Combustion Institute: Pittsburgh, 1992, p 1587–1596.

can be detected as the phenylhydrazone derivative. In the future, formaldehyde gas standards of known concentration must be prepared to determine formaldehyde's detection limit for the online reaction.

The potential analytical impact of the concept presented here should not be underestimated. Through coupling of suitable photolabels to nonaromatic compounds, a larger variety of compound classes can now be considered for REMPI-TOFMS detection, including compounds such as sugars, sulfur compounds, organic acids, or alcohols. Fast screening methods, for example, for environmental samples, biological samples, or medical applications, may be developed on this basis.



We thank Professor Ben Burger for the supply of one of his thermal modulator systems and the useful discussions regarding its application. Financial support from the National Research Foundations in South Africa (NRF), the GSF Research Center, the Federal German Ministry of Education and Scientific Research (BMBF, WTZ program), and the German Environmental Foundation (DBU-Deutsche Bundesstiftung Umwelt) is acknowledged.

Received for review August 11, 2004. Accepted October 21, 2004. AC040151A