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Appendix I: Images that illustrate mining and rehabilitation of coastal dunes along the northeast coast of KwaZulu-Natal, South Africa.



Plate 1. Heavy machinery is used to clear vegetation from the coastal dunes prior to the extraction minerals from the sand.



Plate 2. The mine works as an open-cast dredging system whereby dune sand is taken up by a bucket wheel and separated from the heavy minerals (~4% of the sand) by means of a cyclonic system on the mining plant. This heavy mineral concentrate is taken to the smelter site for further processing where the rutile, zircon and ilmenite are further separated and prepared.



Plate 3. Once separated from the heavy minerals, the sand is stacked into shapes that mimic the pre-mining topographic profile.



Plate 4. Topsoil collected from cleared areas ahead of the mine is brought and spread over the newly stacked dunes. This is then sown with annuals and indigenous to stabilize the dune as soon as possible with a cover crop. Shade-netting is erected to prevent wind erosion, as well as shade and protect seedlings.

Seral stage 1



Plate 5. The cover crop grows up within months and between these annuals and grasses, *Acacia karroo* seedlings begin to germinate (insert picture).



Plate 6. Within three years an impenetrable *Acacia karroo* shrubland has formed (at three years there are 20724 ± 2143 trees/ha (van Dyk 1996)).

Seral stage 2



Plate 7. Within 11 years the *Acacia karroo* still dominates but has thinned to 737 ± 35 trees/ha by the age of 14 (van Dyk 1996). Forest canopy species are beginning to emerge, although the understory is not well-developed



Plate 8. After about 20 years of age the *Acacia karroo* trees begin to senesce and fall over or die standing (inset), forming canopy gaps of varying sizes. Although *A. karroo* remain dominant, the understory has become more developed and forest canopy species are more common.

Seral stage 3



Plate 9. After 30 years gaps of all sizes have formed as *Acacia karroo* continue to fall down. The forest is multi-layered and forest canopy trees are in excess of 8m tall.



Plate 10. In these oldest stands (35 years) *Acacia karroo* has thinned to 141 ± 11 trees/ha (van Dyk 1998), larger gaps comprise grassy patches and clumps of forest tree species. Importantly, these gaps are not recolonized by *A. karroo* (Grainger 2012).

Appendix II: List of species from three taxa recorded in the study area

Table A-1. List of woody plant species identified in the regenerating and unmined forests, third column indicates species associated with forest habitats (●).

Species	Family	Forest-associated species
<i>Acalypha glabrata</i>	Euphorbiaceae	
<i>Acacia karroo</i>	Mimosaceae	
<i>Acacia kraussiana</i>	Mimosaceae	
<i>Acokanthera oppositifolia</i>	Apocynaceae	
<i>Albizia adianthifolia</i>	Mimosaceae	
<i>Allophylus africanus</i>	Sapindaceae	
<i>Allophylus natalensis</i>	Sapindaceae	
<i>Annona senegalensis</i>	Annonaceae	
<i>Antidesma venosum</i>	Euphorbiaceae	
<i>Apodytes dimidiata</i>	Icacinaceae	
<i>Artabotrys monteiroae</i>	Annonaceae	
<i>Barringtonia racemosa</i>	Lecythidaceae	
<i>Bauhinia tomentosa</i>	Caesalpiniaceae	
<i>Bersama lucens</i>	Melianthaceae	
<i>Brachylaena discolor</i>	Asteraceae	
<i>Bridelia cathartica</i>	Euphorbiaceae	
<i>Bridelia micrantha</i>	Euphorbiaceae	
<i>Canthium inerme</i>	Rubiaceae	
<i>Capparis sepiaria</i>	Capparaceae	
<i>Capparis tomentosa</i>	Capparaceae	
<i>Carissa bispinosa</i>	Apocynaceae	●
<i>Carissa macrocarpa</i>	Apocynaceae	
<i>Casuarina equisetifolia</i>	Casuarinaceae	
<i>Cassine eucleiformis</i>	Celastraceae	
<i>Cassipourea gummiflua</i>	Rhizophoraceae	
<i>Cassipourea malosana</i>	Rhizophoraceae	
<i>Cassine tetragona</i>	Celastraceae	
<i>Cassinopsis tinifolia</i>	Icacinaceae	
<i>Catunaregam spinosa</i>	Rubiaceae	
<i>Celtis africana</i>	Ulmaceae	
<i>Cestrum laevigatum</i>	Solanaceae	
<i>Chaetacme aristata</i>	Ulmaceae	●
<i>Chionanthus battiscombei</i>	Oleaceae	
<i>Chionanthus foveolatus</i>	Oleaceae	
<i>Chionanthus peglerae</i>	Oleaceae	●
<i>Chrysanthemoides monilifera</i>	Asteraceae	
<i>Citrus lemon</i>	Rutaceae	
<i>Clausena anisata</i>	Rutaceae	



<i>Clerodendrum glabrum</i>	Verbenaceae	
<i>Clerodendrum myricoides</i>	Verbenaceae	
<i>Cola natalensis</i>	Sterculiaceae	
<i>Commiphora neglecta</i>	Burseraceae	
<i>Cordia caffra</i>	Boraginaceae	●
<i>Croton sylvaticus</i>	Euphorbiaceae	
<i>Cussonia sphaerocephala</i>	Araliaceae	
<i>Dalbergia armata</i>	Fabaceae	
<i>Deinbollia oblongifolia</i>	Sapindaceae	
<i>Dichrostachys cinerea</i>	Mimosaceae	
<i>Diospyros inhacaensis</i>	Ebenaceae	
<i>Diospyros lycioides</i>	Ebenaceae	
<i>Diospyros natalensis</i>	Ebenaceae	●
<i>Dodonaea angustifolia</i>	Sapindaceae	
<i>Dovyalis longispina</i>	Flacourtiaceae	●
<i>Dovyalis rhamnoides</i>	Flacourtiaceae	
<i>Dracaena aletriiformis</i>	Dracaenaceae	
<i>Drypetes natalensis</i>	Euphorbiaceae	
<i>Drypetes reticulata</i>	Euphorbiaceae	
<i>Ekebergia capensis</i>	Meliaceae	●
<i>Elaeodendron croceum</i>	Celastraceae	
<i>Englerophytum natalense</i>	Sapotaceae	
<i>Ephippiocarpa orientalis</i>	Apocynaceae	
<i>Erythrococca berberidea</i>	Euphorbiaceae	●
<i>Erythroxyllum emarginatum</i>	Erythroxyllaceae	
<i>Erythrina lysistemon</i>	Fabaceae	
<i>Euclea natalensis</i>	Ebenaceae	
<i>Euclea racemosa subsp. sinuata</i>	Ebenaceae	●
<i>Eugenia capensis</i>	Myrtaceae	
<i>Eugenia natalitia</i>	Myrtaceae	●
<i>Ficus burtt-davyi</i>	Moraceae	
<i>Ficus craterostoma</i>	Moraceae	
<i>Ficus lutea</i>	Moraceae	
<i>Ficus natalensis</i>	Moraceae	
<i>Ficus polita</i>	Moraceae	
<i>Ficus sur</i>	Moraceae	
<i>Ficus sycomorus</i>	Moraceae	
<i>Ficus trichopoda</i>	Moraceae	
<i>Garcinia livingstonei</i>	Clusiaceae	
<i>Gardenia thunbergia</i>	Rubiaceae	
<i>Grewia caffra</i>	Tiliaceae	
<i>Grewia occidentalis</i>	Tiliaceae	
<i>Halleria lucida</i>	Scrophulariaceae	
<i>Harpephyllum caffrum</i>	Anacardiaceae	
<i>Hibiscus tiliaceus</i>	Malvaceae	



<i>Hymenocardia ulmoides</i>	Euphorbiaceae	
<i>Inhambanella henriquesii</i>	Sapotaceae	
<i>Keetia gueinzii</i>	Rubiaceae	
<i>Kiggelaria africana</i>	Flacourtiaceae	
<i>Kraussia floribunda</i>	Rubiaceae	
<i>Lagynias lasiantha</i>	Rubiaceae	
<i>Lantana camara</i>	Verbenaceae	
<i>Macaranga capensis</i>	Euphorbiaceae	
<i>Maesa lanceolata</i>	Myrsinaceae	
<i>Maerua nervosa</i>	Capparaceae	
<i>Manilkara concolor</i>	Sapotaceae	
<i>Manilkara discolor</i>	Sapotaceae	
<i>Maytenus cordata</i>	Celastraceae	
<i>Maytenus heterophylla</i>	Celastraceae	
<i>Gymnosporia mossambicensis</i>	Celastraceae	
<i>Gymnosporia nemorosa</i>	Celastraceae	
<i>Maytenus procumbens</i>	Celastraceae	
<i>Gymnosporia senegalensis</i>	Celastraceae	
<i>Maytenus undata</i>	Celastraceae	
<i>Melia azedarach</i>	Meliaceae	
<i>Mimusops caffra</i>	Sapotaceae	
<i>Mimusops obovata</i>	Sapotaceae	
<i>Monanthes caffra</i>	Annonaceae	
<i>Myrica serrata</i>	Myricaceae	
<i>Mystroxydon aethiopicum</i>	Celastraceae	
<i>Ochna arborea</i>	Ochnaceae	
<i>Ochna natalitia</i>	Ochnaceae	
<i>Olea capensis</i>	Oleaceae	●
<i>Olea woodiana</i>	Oleaceae	
<i>Osyris compressa</i>	Santalaceae	
<i>Oxyanthus speciosus</i>	Rubiaceae	
<i>Ozoroa obovata</i>	Anacardiaceae	
<i>Pancovia golungensis</i>	Sapindaceae	●
<i>Parinari capensis subsp. incohata</i>	Chrysobalanaceae	
<i>Passerina rigida</i>	Thymelaeaceae	
<i>Pavetta lanceolata</i>	Rubiaceae	
<i>Pavetta natalensis</i>	Rubiaceae	
<i>Pavetta revoluta</i>	Rubiaceae	
<i>Pavetta Sp01</i>	Rubiaceae	
<i>Peddiea africana</i>	Thymelaeaceae	●
<i>Persea americana</i>	Lauraceae	
<i>Phoenix reclinata</i>	Arecaceae	
<i>Pinus elliotti</i>	Pinaceae	
<i>Pisonia aculeata</i>	Nyctaginaceae	
<i>Protorhus longifolia</i>	Anacardiaceae	



<i>Psidium guajava</i>	Myrtaceae	
<i>Psychotria capensis</i>	Rubiaceae	●
<i>Psydrax obovata</i>	Rubiaceae	
<i>Rapanea melanophloeos</i>	Myrsinaceae	
<i>Rauvolfia caffra</i>	Apocynaceae	
<i>Rhoicissus digitata</i>	Vitaceae	
<i>Rhoicissus revoilii</i>	Vitaceae	
<i>Rhoicissus rhomboidea</i>	Vitaceae	●
<i>Rhoicissus tomentosa</i>	Vitaceae	
<i>Rhoicissus tridentata</i>	Vitaceae	
<i>Rhus natalensis</i>	Anacardiaceae	
<i>Rhus nebulosa</i>	Anacardiaceae	
<i>Ricinus communis</i>	Euphorbiaceae	
<i>Rothmannia globosa</i>	Rubiaceae	
<i>Salacia gerrardii</i>	Celastraceae	
<i>Sapium integerrimum</i>	Euphorbiaceae	
<i>Schinus terebinthifolius</i>	Anacardiaceae	
<i>Schefflera umbellifera</i>	Araliaceae	
<i>Sclerocarya birrea</i>	Anacardiaceae	
<i>Scolopia zeyheri</i>	Flacourtiaceae	
<i>Scutia myrtina</i>	Rhamnaceae	
<i>Senna pendula</i>	Caesalpiniaceae	
<i>Sideroxylon inerme</i>	Sapotaceae	
<i>Solanum mauritianum</i>	Solanaceae	
<i>Strychnos gerrardii</i>	Loganiaceae	●
<i>Strychnos henningsii</i>	Loganiaceae	
<i>Strychnos madagascariensis</i>	Loganiaceae	
<i>Strychnos mitis</i>	Loganiaceae	
<i>Strelitzia nicolai</i>	Strelitziaceae	
<i>Strychnos spinosa</i>	Loganiaceae	
<i>Strychnos usambarensis</i>	Loganiaceae	
<i>Syzygium cordatum</i>	Myrtaceae	
<i>Syzygium cumini</i>	Myrtaceae	
<i>Tarenna junodii</i>	Rubiaceae	
<i>Tarenna littoralis</i>	Rubiaceae	
<i>Tarenna pavettoides</i>	Rubiaceae	
<i>Tecomaria capensis</i>	Bignoniaceae	
<i>Teclea gerrardii</i>	Rutaceae	●
<i>Thespesia acutiloba</i>	Malvaceae	
<i>Trema orientalis</i>	Ulmaceae	
<i>Tricalysia delagoensis</i>	Rubiaceae	
<i>Trichilia dregeana</i>	Meliaceae	
<i>Trichilia emetica</i>	Meliaceae	
<i>Tricalysia lanceolata</i>	Rubiaceae	
<i>Tricalysia sonderiana</i>	Rubiaceae	



<i>Turraea floribunda</i>	Meliaceae	●
<i>Turraea obtusifolia</i>	Meliaceae	
<i>Uvaria caffra</i>	Annonaceae	
<i>Vangueria cyanescens</i>	Rubiaceae	
<i>Vangueria infausta</i>	Rubiaceae	
<i>Vangueria randii</i>	Rubiaceae	
<i>Vepris lanceolata</i>	Rutaceae	
<i>Voacanga thouarsii</i>	Apocynaceae	
<i>Xylothea kraussiana</i>	Flacourtiaceae	
<i>Zanthoxylum capense</i>	Rutaceae	
<i>Ziziphus mucronata</i>	Rhamnaceae	

Table A-2. List of millipede species identified in the regenerating and unmined forests, third column indicates species associated with forest habitats (●).

Species	Family	Forest-associated species
<i>Doratogonus sp.</i>	Spirostreptidae	
<i>Centrobolus fulgidus</i>	Spirobolidae	
<i>Centrobolus richardii</i>	Spirobolidae	
<i>Centrobolus rugulosus</i>	Spirobolidae	●
<i>Gnomeskelus tuberosus</i>	Dalodesmidae	●
<i>Orthoporoides sp.*</i>	Spirostreptidae	●
<i>Orthoporoides pyrocephalus</i>	Spirostreptidae	●
<i>Sphaerotherium giganteum</i>	Sphaerotheridae	●
<i>Sphaerotherium punctulatum</i>	Sphaerotheridae	●
<i>Sphaerotherium rotundatum</i>	Sphaerotheridae	●
<i>Sphaerotherium sp. B</i>	Sphaerotheridae	●
<i>Sphaerotherium sp. C</i>	Sphaerotheridae	●
<i>Sphaerotherium sp. D</i>	Sphaerotheridae	●
<i>Sphaerotherium sp. E</i>	Sphaerotheridae	●
<i>Sphaerotherium sp. F</i>	Sphaerotheridae	●
<i>Spinotarsus anguliferus</i>	Odontopygidae	
<i>Spirostreptidae sp. Imm.</i>	Spirostreptidae	●
<i>Spirostreptidae sp. Imm. 2</i>	Spirostreptidae	●
<i>Ulodesmus micramma zuluensis</i>	Dalodesmidae	

Table A-3. List of dung beetle species identified in the regenerating and unmined forests, third column indicates species associated with forest habitats (●).

Species	Forest-associated species
<i>Allogymnopleurus thalassinus</i>	●
<i>Anachalcos convexus</i>	
<i>Caccobius nigrutilus</i>	
<i>Caccobius obtusus</i>	
<i>Caccobius sp. 1</i>	●
<i>Caccobius sp. 2</i>	
<i>Caccobius sp. 3</i>	●
<i>Caccobius sp. 4</i>	
<i>Caccobius sp. 5 = Caccobius cavatus</i>	
<i>Catharsius sp. 1 (endemic)</i>	
<i>Catharsius mossambicanus</i>	
<i>Catharsius tricornutus</i>	●
<i>Cleptocaccobius viridicollis</i>	
<i>Copris inhalatus ssp santaluciae</i>	
<i>Copris puncticollis</i>	
<i>Copris urus</i>	
<i>Digitonthophagus gazella</i>	
<i>Drepanocerus impressicollis (now Afrodrepanus impressicollis)</i>	
<i>Drepanocerus kirbyi</i>	
<i>Euoniticellus intermedius</i>	
<i>Garreta azureus</i>	●
<i>Garreta unicolor</i>	●
<i>Gyronotus carinatus</i>	
<i>Heliocopris hamadryas</i>	
<i>Hyalonthophagus alcyonides</i>	
<i>Kheper lamarcki</i>	●
<i>Liatongus militaris</i>	●
<i>Metacatharsius sp. 1 (=zuluanus)</i>	●
<i>Milichus sp. 1</i>	
<i>Neosisyphus confrater</i>	
<i>Neosisyphus mirabilis</i>	●
<i>Neosisyphus spinipes</i>	
<i>Odontoloma sp.</i>	
<i>Oniticellus formosus</i>	●
<i>Oniticellus planatus</i>	
<i>Onthophagus aeruginosus</i>	●
<i>Onthophagus ambiguus (now Mimonthophagus ambiguus)</i>	
<i>Onthophagus bicavifrons</i>	
<i>Onthophagus depressus</i>	
<i>Onthophagus fimetarius (coastal var.) possibly new</i>	
<i>Onthophagus flavolimbatus</i>	
<i>Onthophagus lacustris</i>	
<i>Onthophagus nanus</i>	
<i>Onthophagus obtusicornis</i>	
<i>Onthophagus pugionatus</i>	
<i>Onthophagus quadrinodosus</i>	●
<i>Onthophagus signatus</i>	●
<i>Onthophagus sp 1 (=horned pullus)</i>	
<i>Onthophagus sp. 2 (v. small endemic)</i>	



<i>Onthophagus sp 3 (=sp. e George)</i>	
<i>Onthophagus sp 4</i>	
<i>Onthophagus sp nr bicavifrons</i>	●
<i>Onthophagus sp. nr sugillatus (coastal var.) possibly new</i>	
<i>Onthophagus ursinus</i>	●
<i>Onthophagus vinctus</i>	●
<i>Onthophagus stellio or variegatus gp??</i>	
<i>Onthophagus sp - mottled tail</i>	
<i>Onthophagus sp A</i>	
<i>Pachylomerus femoralis</i>	
<i>Pedaria sp. IV</i>	●
<i>Pedaria sp. III</i>	
<i>Proagoderus aciculatus</i>	
<i>Proagoderus aureiceps</i>	
<i>Proagoderus brucei (now P. chalcostolus)</i>	
<i>Scarabaeus bornemisszai</i>	
<i>Scarabaeus goryi</i>	●
<i>Sisyphus natalensis (cited as the syn. S. bornemisszanus)</i>	●
<i>Sisyphus seminulum</i>	
<i>Sisyphus sordidus</i>	●
<i>Sisyphus sp nr gazanus</i>	●
<i>Sisyphus sp y</i>	
<i>Stiptopodius sp. 1</i>	

Table A-4. List of bird species identified in the regenerating and unmined forests, third column indicates species associated with forest habitats (●).

<i>Species</i>	Common Name	Forest-associated species
<i>Acrocephalus palustris</i>	Eurasian Marsh Warbler	
<i>Alcedo cristata</i>	Malachite Kingfisher	
<i>Amblyospiza albifrons</i>	Thick-billed Weaver	
<i>Andropadus importunus</i>	Sombre Greenbul	
<i>Anthus cinnamomeus</i>	African Pipit	
<i>Apalis flavida</i>	Yellow-breasted Apalis	
<i>Apalis ruddi</i>	Rudd's Apalis	
<i>Apalis thoracica</i>	Bar-throated Apalis	
<i>Apaloderma narina</i>	Narina Trogon	●
<i>Aplopelia larvata</i>	Lemon Dove	●
<i>Ardea melanocephala</i>	Black-headed Heron	
<i>Batis capensis</i>	Cape Batis	
<i>Batis fratrum</i>	Woodwards' Batis	●
<i>Bostrychia hagedash</i>	Hadedda Ibis	
<i>Bradornis pallidus</i>	Pale Flycatcher	
<i>Bycanistes bucinator</i>	Trumpeter Hornbill	●
<i>Calendulauda sabota</i>	Sabota Lark	
<i>Camaroptera brachyura</i>	Green-backed Camaroptera	
<i>Campephaga flava</i>	Black Cuckooshrike	
<i>Campethera abingoni</i>	Golden-tailed Woodpecker	
<i>Caprimulgus europaeus</i>	European Nightjar	
<i>Centropus burchellii</i>	Burchell's Coucal	
<i>Cercotrichas leucophrys</i>	White-browed Scrub-Robin	
<i>Cercotrichas quadrivirgata</i>	Bearded Scrub-Robin	
<i>Cercotrichas signata</i>	Brown Scrub-Robin	
<i>Ceuthmochares aereus</i>	Green Malkoha	●
<i>Chalcomitra amethystina</i>	Amethyst Sunbird	
<i>Chalcomitra senegalensis</i>	Scarlet-chested Sunbird	
<i>Chlorocichla falviventris</i>	Yellow-bellied Greenbul	
<i>Chrysococcyx caprius</i>	Diederik Cuckoo	
<i>Chrysococcyx cupreus</i>	African Emerald Cuckoo	●
<i>Chrysococcyx klaas</i>	Klaas's Cuckoo	
<i>Cinnyris bifasciata</i>	Purple-banded Sunbird	
<i>Cisticola chinianus</i>	Rattling Cisticola	
<i>Cisticola cinnamomeus</i>	Pale-crowned Cisticola	
<i>Cisticola fulvicapilla</i>	Neddicky	
<i>Cisticola juncidis</i>	Zitting Cisticola	
<i>Cisticola natelensis</i>	Croaking Cisticola	
<i>Clamator jacobinus</i>	Jacobin Cuckoo	
<i>Coccygia melanotis</i>	Swee Waxbill	



<i>Colius striatus</i>	Speckled Mousebird	
<i>Columba delegorguei</i>	Eastern Bronze-naped Pigeon	●
<i>Coracias caudata</i>	Lilac-breasted Roller	
<i>Coracias garrulus</i>	Eurasian Roller	
<i>Coracina caesia</i>	Grey Cuckooshrike	
<i>Corvus albus</i>	Pied Crow	
<i>Cossypha caffra</i>	Cape Robin-Chat	
<i>Cossypha dichroa</i>	Chorister Robin-Chat	
<i>Cossypha natalensis</i>	Red-capped Robin-Chat	
<i>Coturnix coturnix</i>	Common Quail	
<i>Cuculus canorus</i>	Common Cuckoo	
<i>Cuculus gularis</i>	African Cuckoo	
<i>Cuculus solitarius</i>	Red-chested Cuckoo	
<i>Cyanomitra olivacea</i>	Eastern Olive Sunbird	●
<i>Cyanomitra veroxii</i>	Grey Sunbird	●
<i>Dendropicos fuscescens</i>	Cardinal Woodpecker	
<i>Dendropicos griseocephalus</i>	Olive Woodpecker	
<i>Dicrurus adsimilis</i>	Fork-tailed Drongo	
<i>Dicrurus ludwigii</i>	Square-tailed Drongo	●
<i>Dryoscopus cubla</i>	Black-backed Puffback	
<i>Estrilda astrild</i>	Common Waxbill	
<i>Estrilda perreini</i>	Grey Waxbill	
<i>Euplectes axillaris</i>	Fan-tailed Widowbird	
<i>Euplectes orix</i>	Southern Red Bishop	
<i>Eurystomus glaucurus</i>	Broad-billed Roller	
<i>Guttera edouardi</i>	Crested Guineafowl	
<i>Halcyon albiventris</i>	Brown-hooded Kingfisher	
<i>Hedydipna collaris</i>	Collared Sunbird	
<i>Hippolais icterina</i>	Icterine Warbler	
<i>Indicator minor</i>	Lesser Honeyguide	
<i>Indicator variegatus</i>	Scaly-throated Honeyguide	
<i>Ispidina picta</i>	African Pygmy-Kingfisher	
<i>Lagonosticta rubricata</i>	African Firefinch	
<i>Lamprotornis corruscus</i>	Black-bellied Starling	●
<i>Laniarius ferrugineus</i>	Southern Boubou	
<i>Lanius collaris</i>	Common Fiscal	
<i>Lanius collurio</i>	Red-backed Shrike	
<i>Lanius minor</i>	Lesser Grey Shrike	
<i>Lonchura cucllata</i>	Bronze Mannikin	
<i>Lonchura nigriceps</i>	Red-backed Mannikin	
<i>Lybius torquatus</i>	Black-collared Barbet	
<i>Macronyx croceus</i>	Yellow-throated Longclaw	
<i>Malaconotus blanchoti</i>	Grey-headed Bush-Shrike	
<i>Mandingoa nitidula</i>	Green Twinspot	



<i>Megaceryle maxima</i>	Giant Kingfisher	
<i>Melaenornis pammelaina</i>	Southern Black Flycatcher	
<i>Merops pusillus</i>	Little Bee-eater	
<i>Mirafra africana</i>	Rufous-naped Lark	
<i>Monticola rupestris</i>	Cape Rock-Thrush	
<i>Motacilla aguimp</i>	African Pied Wagtail	
<i>Motacilla capensis</i>	Cape Wagtail	
<i>Muscicapa adusta</i>	African Dusky Flycatcher	●
<i>Muscicapa caerulescens</i>	Ashy Flycatcher	●
<i>Muscicapa striata</i>	Spotted Flycatcher	
<i>Musophaga porphyreolopha</i>	Purple-crested Turaco	●
<i>Myioparus plumbeus</i>	Grey Tit-Flycatcher	
<i>Nicator gularis</i>	Eastern Nicator	
<i>Oriolus larvatus</i>	Black-headed Oriole	
<i>Oriolus oriolus</i>	Eurasian Golden Oriole	
<i>Passer domesticus</i>	House Sparrow	
<i>Phyllastrephus terrestris</i>	Terrestrial Brownbul	
<i>Phylloscopus trochilus</i>	Willow Warbler	
<i>Platysteira peltata</i>	Black-throated Wattle-eye	●
<i>Plectropterus gambensis</i>	Spur-winged Goose	
<i>Ploceus</i>	Weavers	
<i>Ploceus bicolor</i>	Dark-backed Weaver	●
<i>Ploceus cucullatus</i>	Village Weaver	
<i>Ploceus intermedius</i>	Lesser Masked-Weaver	
<i>Ploceus ocularis</i>	Spectacled Weaver	
<i>Ploceus subaureus</i>	Yellow Weaver	
<i>Pogoniulus bilineatus</i>	Yellow-rumped Tinkerbird	●
<i>Pogoniulus pusillus</i>	Red-fronted Tinkerbird	
<i>Pogonocichla stellata</i>	White-starred Robin	
<i>Prinia subflava</i>	Tawny-flanked Prinia	
<i>Pycnonotus tricolor</i>	Dark-capped Bulbul	
<i>Rhinopomastus cyanomelas</i>	Common Scimitarbill	
<i>Sarothrura elegans</i>	Buff-Spotted Flufftail	
<i>Saxicola torquata</i>	African Stonechat	
<i>Serinus canicollis</i>	Cape Canary	
<i>Serinus mozambicus</i>	Yellow-fronted Canary	
<i>Serinus sulphuratus</i>	Brimstone Canary	
<i>Sigelus silens</i>	Fiscal Flycatcher	
<i>Smithornis capensis</i>	African Broadbill	
<i>Stactolaema leucotis</i>	White-eared Barbet	●
<i>Streptopelia capicola</i>	Cape Turtle Dove	
<i>Streptopelia semitorquata</i>	Red-eyed Dove	
<i>Sylvia borin</i>	Garden Warbler	
<i>Sylvietta rufescens</i>	Long-billed Crombec	



<i>Tauraco corythis</i>	Knysna Turaco	●
<i>Tauraco livingstonii</i>	Livingstone's Turaco	●
<i>Tchagra australis</i>	Brown-crowned Tchagra	
<i>Tchagra senegala</i>	Black-crowned Tchagra	
<i>Telophorus olivaceus</i>	Olive Bush-Shrike	
<i>Telophorus quadricolor</i>	Gorgeous Bush-Shrike	
<i>Telophorus sulfureopectus</i>	Orange-breasted Bush-Shrike	
<i>Terpsiphone viridis</i>	African Paradise-Flycatcher	
<i>Tockus alboterminatus</i>	Crowned Hornbill	●
<i>Trachyphonus vallantii</i>	Crested Barbet	
<i>Treron calva</i>	African Green-Pigeon	
<i>Trochocercus cyanomelas</i>	Blue-mantled Crested Flycatcher	●
<i>Turdus libonyanus</i>	Kurrichane Thrush	
<i>Turtur chalcospilos</i>	Emerald-spotted Wood-Dove	
<i>Turtur tympanistria</i>	Tambourine Dove	●
<i>Uraeginthus angolensis</i>	Blue Waxbill	
<i>Urocolius indicus</i>	Red-faced Mousebird	
<i>Vidua macroura</i>	Pin-tailed Whydah	
<i>Zoothera guttata</i>	Spotted Ground-Thrush	
<i>Zosterops virens</i>	Cape White-eye	



1 Appendix III: Manuscript accepted for publication at *Landscape and Ecological*
2 *Engineering* (DOI: 10.1007/s11355-013-0211-1).

3 **Journal:** Landscape and Ecological Engineering

4 **Manuscript type:** Original paper

5

6 **Title:** Coastal dune topography as a determinant of abiotic conditions and biological
7 community restoration in northern Kwazulu-Natal, South Africa

8

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14

15 **Keywords:** aspect, dune morphology, elevation, gradient, microclimate, soil

16 **Word count:** 4,813 (including main text and references)

17

18 **Abstract**

19 Topography is rarely considered as an independent goal of restoration. Yet, topography
20 determines micro-environmental conditions and hence living conditions for species.
21 Restoring topography may therefore be an important first step in ecological restoration. We
22 aimed at establishing the relative importance of topography where coastal dunes destroyed by
23 mining are rebuilt as part of a rehabilitation programme.

24 We assessed the response of 1) microclimatic and soil conditions, and 2) woody plant
25 and millipede species richness and density, to location-specific topographic profiles. We
26 enumerated the topographic profile using variables of dune morphology (aspect, elevation
27 and gradient) as well as relative position on a dune (crest, slope, valley).

28 Temperature, relative humidity and light intensity varied with aspect, elevation,
29 gradient and position. However, regeneration age was a better predictor of soil nutrient
30 availability than these topographic variables. Age also interacted with topographic variables
31 to explain tree canopy density and species richness, as well as millipede species richness. The
32 density of keeled millipedes (forest specialists) was best explained by topographic variables
33 alone. The transient nature of these new-growth coastal dune forests likely masks
34 topography-related effects on communities because age-related succession (increasing
35 structural complexity) drives the establishment and persistence of biological communities,
36 not habitat conditions modulated by topography. However, our study has shown that the
37 microhabitats associated with topographic variability influence specialist species more than
38 generalists.

39

40

41 **Introduction**

42 Ecological restoration is widely recognised as a conservation tool and aims to re-instate
43 natural processes that sustain biological diversity (Dobson et al. 1997; MacMahon & Holl
44 2001; Rands 2012). Such diversity is determined by both regional and local forces, the latter
45 often as a function of topography due to cascading effects on microclimatic conditions, water
46 retention, and nutrient availability (Larkin et al. 2006). These relationships are especially
47 well-documented in mountainous regions (Burnett et al. 1998; Nichols et al. 1998; Tateno &
48 Takeda 2003; da Silva et al. 2008), but less often for coastal sand dune ecosystems (e.g.
49 Martínez et al. 2001; Acosta et al. 2007). The restoration of topography may be a priority
50 (Weiss & Murphy 1990; Palik et al. 2000; Larkin et al. 2006), but difficult or costly to
51 achieve. However, an approximation of the original topography may be sufficient to maintain
52 desired ecological processes. This may well be the case in our study areas where succession
53 drives forest regeneration, but where the full complement of species has not yet been
54 regained (van Aarde et al. 1996b; Grainger 2012). This may be due to the micro-
55 environmental needs of specialist species not being met due to constraints imposed by
56 topography. Justification to restore terrain requires an assessment of the relevance of
57 topography for species and ecological processes. In this study, we assess the influence of
58 dune topography on abiotic and biotic conditions (Table) in coastal dune forests regenerating
59 in response to an ecological restoration program.

60 The aspect, elevation, and gradient of slopes are collectively referred to as dune
61 morphology, while the relative position is described as the crest, slope, or valley. These
62 variables of dune topography can modulate habitat conditions in various ways (Larkin et al.
63 2006). For example, nutrients leaching from dune crests into valleys where plant-
64 communities are light-limited results in nutrient-limited communities on crests, but greater
65 nutrient availability in valleys (Tateno & Takeda 2003). Canopy structure changes with

66 gradients in soil fertility and light (Nichols et al. 1998; Tateno & Takeda 2003), even with
67 limited altitudinal variation (da Silva et al. 2008). This may explain patterns in plant species
68 composition, abundance, and distribution (Chen et al. 1997; Oliviera-Filho et al. 1998). The
69 aspect and gradient of dune slopes may amplify these differences as they also influence light
70 availability (Oliviera-Filho et al. 1998; Bennie et al. 2008) and wind exposure (Chen et al.
71 1997; Acosta et al. 2007). Wind sculpts tree canopies (Kubota et al. 2004), hastens canopy
72 gap formation (Ritter et al. 2005), and contributes to seed dispersal (Furley & Newey 1979).
73 The windward slopes of coastal dunes have higher evaporation rates than leeward slopes and
74 are more exposed to salt spray that increases salt concentrations in the soil, in turn
75 influencing soil pH and the availability of nutrients (Furley & Newey 1979; Chen et al. 1997;
76 Acosta et al. 2007). We therefore hypothesized that dune morphology and position would 1)
77 modulate microclimatic conditions (temperature, relative humidity, and light intensity) and 2)
78 influence soil nutrient availability (C:N ratio) and soil pH (see Table). Disturbed or
79 destroyed topographic profiles could therefore hinder the ecological restoration of plant and
80 animal communities of new-growth forests, or simply alter heterogeneity and rearrange the
81 distribution of resources. Thus the structure and composition of biotic communities at
82 locations with different dune morphologies should be assessed to determine the importance of
83 restoring the topographic profile.

84 Topography influences plant growth and species richness in old-growth forests
85 (Tateno & Takeda 2003; da Silva et al. 2008), which has cascading effects on biota through the
86 responses of microclimatic conditions to topography (Larkin et al. 2006). Physiological trade-
87 offs associated with the small size and ectothermy of invertebrates, such as millipedes,
88 renders them sensitive to microclimatic conditions that dictate habitat preferences (Ashwini
89 & Sridhar 2008; Loranger-Merciris et al. 2008; David & Gillon 2009). We therefore assessed
90 the importance of the topographic profile in structuring millipede assemblages. We

91 hypothesized that within a seral stage, dune morphology and position would 3) influence
92 plant community structure and composition, and 4) millipede community structure and
93 composition in regenerating stands of new-growth coastal dune forest (Table). If millipedes
94 respond to topography, changes in the topographic profile should result in changes in
95 millipede diversity. If this is not the case, topography has a limited role to play, if any, in
96 explaining millipede community structure. Although this study is based upon coastal dune
97 forests, it may have implications for any disturbed dune system under restoration.

98 **Methods**

99 *Study area*

100 The study area was located north of Richards Bay town (between 28°46' and 28°34' south) on
101 the sub-tropical north coast of Kwazulu-Natal, South Africa (Fig. 1). The climate is humid
102 with a mean annual rainfall of 1458 ± 493.5 mm (mean \pm SD, $n = 34$ years between 1976 and
103 2009), peaking in February. The mean annual temperature was $23.79 \pm 3.40^\circ\text{C}$ ($n = 3$ years
104 between 2006 and 2009). Winds of between 10 and 40 $\text{km}\cdot\text{h}^{-1}$ blew from the north-east for
105 about 20% of the time, as did those from south-south west and south-west combined (data
106 courtesy of Richards Bay Minerals).

107 The establishment of forests on the coastal dunes here occurred with the return of
108 warm interglacial conditions between 6,500 and 4,000 years ago, making them among the
109 highest vegetated dunes in the world (Weisser & Marques 1979; Lawes 1990). These forests
110 are therefore relatively young and harbour few endemic species (Lawes 1990; van Wyk &
111 Smith 2001). Coastal dune forests are sensitive to disturbance but previous work has shown
112 that they are relatively resilient and are thus able to recover (e.g. Wassenaar et al. 2005;
113 Grainger et al. 2011).

114 Richards Bay Minerals (RBM) has leased this area since 1976 for the extraction of
115 heavy metals from the coastal sands. Ahead of the dredging pond, all vegetation was cleared
116 and the dunes were collapsed for mining. After mining, sand tailings were stacked to
117 resemble pre-mining topography and were covered with topsoil (van Aarde et al. 1996c). A
118 third of the mined area was set aside for the restoration of indigenous coastal dune forest and
119 this area comprised known-aged stands that at the time of the study ranged in age from 1 year
120 (in the northeast) to 33 years (in the southwest) (see Fig. 1). This age-range represented three
121 seral stages based on those defined by Grainger (2012): seral stage one = 1-10 years, two =
122 11-25 years, and three >25 years. Adjoined by a coastal strip of unmined vegetation about
123 200 m wide, the stands were themselves no wider than 2 km, set in a mosaic of active mining
124 areas, plantations, degraded woodland, and rural villages (Wassenaar et al. 2005).

125 *Microclimatic data*

126 Fifteen HOBO[®] 4-channel data loggers (Onset Computer Corporation, 470 MacArthur Blvd.,
127 Bourne, MA 02532, U.S.A.) were deployed in the 22-year old stand (see Fig. 1) on custom-
128 made platforms placed 10 cm above the ground (five on the crest, five on a slope and five in
129 the valley). We programmed these loggers to record ground-level temperature, relative
130 humidity, and light intensity (see Table 2 for definitions) every 10 minutes between 08:00, 28
131 January and 05:00, 4 February 2011, yielding 14,850 records.

132 *Soil surveys*

133 An auger was used to collect soil samples to 20 cm depth at the corners and centre of each of
134 the millipede survey transects (see below). These five samples were mixed into a single bag
135 and consequently 65 bags were analyzed at the Department of Plant Production and Soil
136 Science at the University of Pretoria using procedures described in van Aarde et al. (1998;
137 see supplementary information for detailed chemical profile). We used Nitrogen and Carbon

138 concentrations to calculate the carbon-to-nitrogen ratio (C:N, Table 2) and included the pH
139 values of each sample in our analysis.

140 *Woody plant surveys*

141 All woody plants taller than 0.2 m in 106 randomly placed quadrats (16×16-m, at least 100 m
142 apart) in six stands of known regeneration age (10, 14, 18, 22, 26, and 33 years) were
143 sampled between July and November 2010. Each plant was identified against reference
144 material. We calculated six variables of woody plant community structure for each quadrat
145 (see Table 2).

146 *Millipede surveys*

147 Millipede species occurring on the ground up to 3 m on plants were counted between 13
148 January and 4 February 2011 in 65 randomly placed transects within a 10, 14, 18, 22, 26, and
149 33 year-old stand (see Fig. 2). Each transect was 32 × 6-m wide and comprised 48 2 × 2-m
150 cells. Surveys were conducted by three observers per transect, each responsible for a column
151 of 16 cells. All millipedes found in a cell during five minutes were identified based on
152 reference images and descriptions (Porter et al. 2007), counted, and removed to avoid
153 recounting. We calculated the number of millipede species and the density of cylindrical,
154 keeled, and pill millipedes (see Table 2) within each location-specific transect.

155 *Topographic data*

156 We used classified topographic data based on eight cardinal directions (aspect), seven
157 elevation categories, and five gradient categories that had been extracted from a topographic
158 map (see Fig.1.) based on a Light Detection and Ranging (LIDAR) mission conducted in
159 2010 (post-mining). We used GIS overlay procedures to relate all of the sampling points and
160 quadrat locations recorded in the field to location-specific variables of dune morphology
161 based on the topographic maps.



162 *Statistical analyses*

163 We used stratified random sampling to extract one microclimate record (including the
164 temperature, relative humidity, and light intensity readings) per hour, per logger for each
165 sampling day (29 January – 3 February 2011), rendering 2,475 records to be included in
166 analyses. We \log_{10} -transformed the light intensity data to meet assumptions for analyses of
167 variance (ANOVA). To determine whether microclimatic conditions were modulated by
168 topography, we conducted repeated measures ANOVA with hour and day as repeated
169 measures factors, and categorized variables of dune morphology as between-groups factors.

170 We assessed the influence of dune morphological variables on soil C:N ratios and pH,
171 as well as woody plant and millipede community variables in each of the three seral stages.
172 We assessed these using generalized linear models with age as a covariate (Analyses of
173 Covariance (ANCOVA) for all seral stages for woody plants and seral stages 2 and 3 for soil
174 and millipedes. Millipede and soil data for seral stage 1 comprised too few cases and was
175 therefore not assessed separately. All statistical analyses were conducted using STATISTICA
176 10 (Statsoft Inc., Tulsa, Oklahoma).

177 Woody plant and millipede species abundance data were \log_{10} -transformed and
178 calculated the similarity between quadrats, with different dune morphological characteristics
179 using the Bray-Curtis index. Cluster analyses and non-metric multi-dimensional scaling
180 (NMDS) were used to detect community clusters based on the four characteristics of dune
181 morphology. Analyses of similarity (ANOSIM) allowed us to assess the significance of
182 community groupings based on dune morphology within each successional stage. To identify
183 the distinguishing species, we conducted similarity percentage (SIMPER) analyses
184 (SIMPER) for those community groupings that differed significantly based on dune
185 morphological characteristics. All multivariate techniques were conducted using PRIMER 6
186 software (Clarke 1993).

187 **Results**188 *Dune topography and abiotic variables*

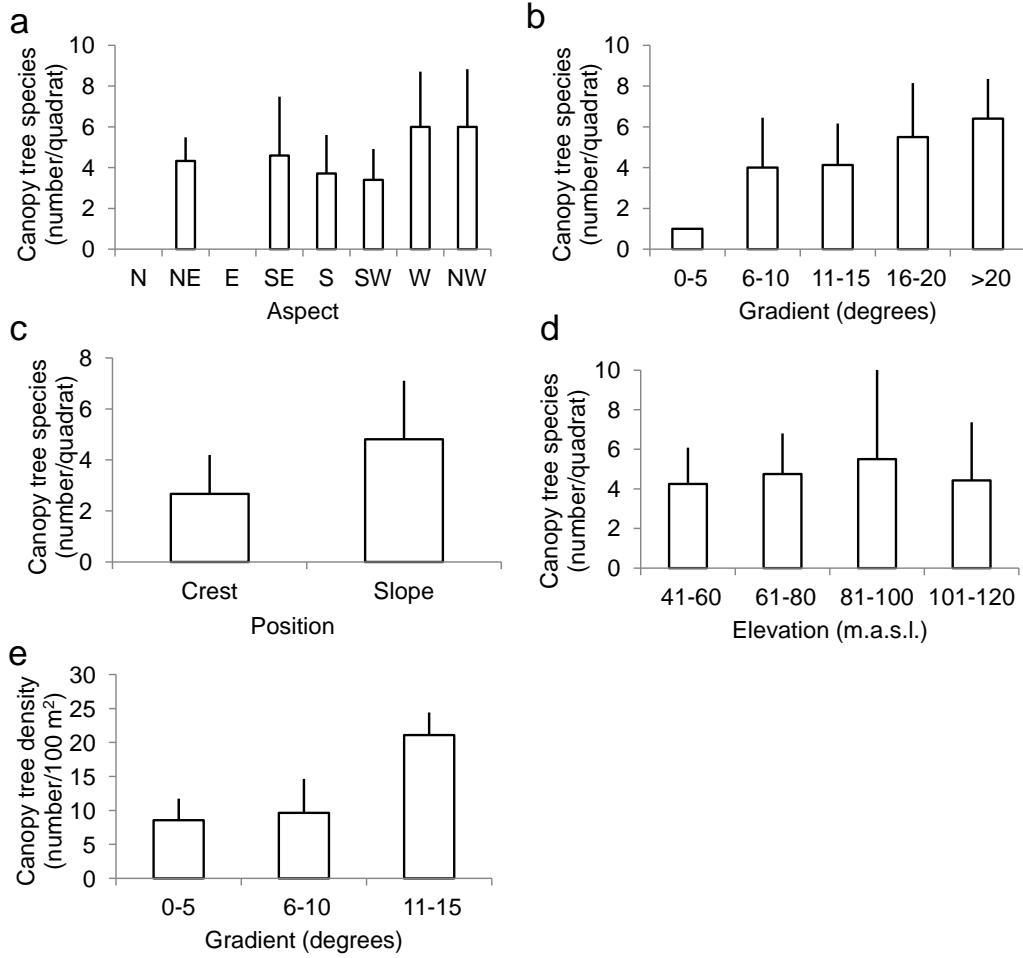
189 Temperature was significantly modulated by aspect and gradient when sampling day and time
190 of day were taken into account (repeated measures ANOVA: $F_{(575, 1035)} = 1.33, p < 0.001$ and
191 $F_{(230, 1380)} = 1.27, p = 0.007$, respectively). Similarly, relative humidity was significantly
192 modulated by elevation ($F_{(345, 1265)} = 1.7632, p < 0.001$), gradient ($F_{(230, 1380)} = 1.69, p < 0.001$)
193 and position ($F_{(230, 1380)} = 1.65, p < 0.001$), while light intensity was influenced by aspect ($F_{(575,$
194 $1035)} = 1.93, p < 0.001$) and position ($F_{(230, 1380)} = 1.38, p < 0.001$). Northern slopes were hotter
195 and lighter than other slopes, although south-facing slopes were also relatively warm. Low-
196 lying areas were relatively humid compared to higher dunes. Slopes with mid-range steepness
197 were generally more humid, but cooler than comparatively gentle and steep slopes. Valleys
198 were generally more humid and darker than crests and slopes. For illustrative purposes, we
199 presented one day's data for these significant cases (see Fig. 2).

200 Variability in soil pH was best explained by age in seral stage 2, and a model
201 including aspect, elevation, and position in addition to age in seral stage 3 (ANCOVA and
202 AIC; Table 3). However, none of the models significantly explained variability in soil C:N
203 ratios (Table 3).

204 *Dune topography and biotic variables*

205 The 8,833 woody plants sampled in 106 quadrats comprised 7,122 canopy and 1,736
206 understory plants among 88 species. Variability in all woody plant variables was best
207 explained by models that included age as a covariate within pooled seral stages, as was the

208 case when seral stage 2 was treated separately (ANCOVA and AIC;



209

210 Figure 5-3. Mean \pm one standard deviation of the mean of woody plant response variables
211 presented as a function of those variables that best-explained their variability significantly
212 despite stand age (see Table 2).

213 Table 5-3). The number of tree canopy species in seral stage 1 was best explained by a model
214 including aspect, elevation, gradient, and position, but not age. There were more species on
215 west- and northwest-facing slopes compared to south- and southwest-facing slopes (Fig. 3a),
216 while relatively flat slopes had fewer species than other gradients (Figure 5-3b), as did crests
217 relative to slopes (Fig. 3c). However, canopy tree species richness varied little with elevation
218 (Fig. 3d). Tree density in seral stage 3 increased significantly with gradient (ANCOVA and
219 AIC; Fig. 3e).

220 Only 11% of the variability in tree species abundances was explained by gradient in
221 seral stage 2, although the NMDS plot was unconvincing of this separation (ANOSIM, $p <$
222 0.05, Fig.). Nevertheless, SIMPER analysis revealed consistent dominance by *Acacia karroo*
223 Hayne and *Celtis africana* Burm.f. (contributing more than 80% of the community) across all
224 gradients (Table 4). However, the number of species increased with gradient so that in
225 addition to these two species, *Allophylus natalensis* Sond. (Dune False Currant) characterized
226 slopes ranging from 0 to 15° and *Brachylaena discolor* DC. (Coast Silver-oak) those of 11 to
227 15°. Slopes of more than 15° were characterised by the addition of *Grewia occidentalis* L.
228 (Cross-berry), *Chaetachme aristata* Planch. (Giant Pock Ironwood) and *Teclea gerrardii*
229 I. Verd. (Zulu Cherry-orange), though all with less than a 5% contribution to tree communities
230 on these slopes (Table 4).

231 Elevation explained 32% of the variability in understory species abundances in seral stage 3
232 (ANOSIM, $p < 0.05$, Fig.). However, this was the result of most cases representing mid-
233 elevations of 41–60 m.a.s.l, with very few cases for other elevation categories. Nevertheless,
234 these mid-elevations were dominated (61% contribution) by *Rhoicissus revoilii* Planch.
235 (Bushveld grape), followed by *Scutia myrtina* Burm.F (Cat-thorn) that contributed 28%, and
236 the invasive alien species, *Chromolaena odorata* L. (Triffid Weed), contributing 11% (Table
237 5). Elevations of 61–80 m.a.s.l. were dominated by *S. myrtina* alone (Table 5).

238 *Millipede assemblages*

239 We recorded 28,987 millipedes (28,351 cylindrical, 513 keeled, and 123 pill millipedes) from
240 16 species in 65 quadrats. The number of millipede species in the transects of seral stage 2
241 covaried with dune position (Table 3), whereby valleys had the most species, though that of
242 slopes and crests did not differ from one another (Fig. 5). Models including age as a covariate
243 in addition to variables of dune morphology best explained the density of cylindrical
244 millipedes for pooled and separated seral stages. Pill millipede density was very low and also
245 driven by rehabilitating stand age in combination with dune morphological variables for
246 pooled as well as separate seral stages. The density of keeled millipedes for pooled seral
247 stages was best explained by a model including aspect, elevation, gradient, and position, but
248 not age (Table 3). These millipedes were most prolific in valleys (Fig. 5b), as well as east-
249 facing slopes (Fig. 5c) with gradients steeper than 10° (Fig. 5d). However, we found little
250 correlation between millipede communities and elevation (Fig. 5e), and when seral stages
251 were separated age was included in the best-fit model (Table 3). Based on our ANOSIM
252 analyses none of the variables of dune morphology significantly influenced species-specific
253 millipede abundances.

254 **Discussion**

255 In line with our hypotheses, dune morphology modulated microclimatic conditions in a
256 similar manner as reported for other studies (Tateno & Takeda 2003; Bennie et al. 2008). We
257 acknowledge though, that the conditions on each dune face are likely the product of
258 conditions ameliorated or exacerbated by surrounding dunes that have consequences for wind
259 channelling and shading, thus cumulatively influencing microclimatic conditions. Contrary to
260 our hypotheses, variability in soil nutrient concentrations was not explained by dune
261 morphology, but rather by regeneration age. The processing of sand as part of the mining

262 operation probably reshuffled soil nutrients and minerals that accumulate through natural
263 processes. With only a few years of post-mining regeneration of biotic activity and
264 mechanical processes (e.g. leeching) it is not surprising that soil fertility (C:N ratios) and pH
265 levels are not yet conforming to expected spatially structured patterns induced by dune
266 topography. Given the weak associations between topographic and abiotic variables, it is also
267 not surprising that spatial variability in woody plant and millipede community structure could
268 not be explained by topographic variables.

269 Species richness and density, as well as species-specific abundances of canopy trees
270 and the understory varied with topography, as did millipede species richness, all in support of
271 our formulated hypotheses, though with the caveat of an overriding influence of regeneration
272 age. Keeled millipedes, a group of invertebrates associated with forests, also responded to
273 topography, although cylindrical and pill millipedes did not. This suggests that forest
274 specialists may be more sensitive to microhabitats induced by topography, but this requires
275 further investigation.

276 Increasing slope steepness resulted in more dense woody plant canopies in stands
277 older than 25 years, a finding similar to that of van Dyk (1996) for earlier stages of
278 regeneration in the study area. Laurance et al. (1999) also described a decrease in the number
279 of large trees with increased tree density on steep slopes. Although woody plant communities
280 of different gradients in stands of 11-25 years were generally dominated by similar sets of
281 forest tree species, species composition varied with the gradient of slopes. Incidentally, the
282 majority of these dominant species were identified by Grainger (2012) as species that could
283 colonize newly formed gaps in the woodland. This was likely due to their wide tolerance to
284 irradiance, temperatures, and moisture that change along dune slopes with elevation and
285 gradient (Ritter et al. 2005). Species abundances of canopy and understory communities
286 responded to different gradients in stands of 11-25 years, and elevation in stands of >25

287 years, respectively. The number of canopy species, though not their abundances, was best
288 explained by aspect, elevation, gradient, and position in stands younger than 11 years,
289 suggesting that dune morphology may provide habitat conditions that support different
290 species in the early stages of succession when conditions are likely to be most harsh.

291 Millipede variables also responded to age and dune morphology. Explanatory models
292 for cylindrical and pill millipede density included age as a covariate. These relationships are
293 likely the result of age-related increases in woodland complexity (Kritzing & van Aarde
294 1998), moisture-retention and nutrient accumulation associated with litter accumulation (van
295 Aarde et al. 1998) and the modulation of microclimate by topography as discussed above. As
296 in Greyling et al. (2001), two closely related cylindrical millipedes (Centrobolidae)
297 dominated these new-growth forests. This may have obscured patterns in species-specific
298 abundances related to topography. However, the number of millipede species covaried with
299 position on the dune face in stands of 11-25 years, whereby valleys supported more millipede
300 species than slopes and crests. When considering the microclimatic data, this likely relates to
301 the moderate temperature and light intensities but relatively humid conditions that existed in
302 the valleys in comparison with ambient conditions such as wind and high temperatures.
303 Keeled millipedes responded to topographic variables independent of age and this likely
304 relates to the provision of microhabitats for this relatively small, forest-associated species and
305 justifies further study.

306 Despite the idiosyncratic responses by woody plants and millipedes, position on the
307 dune, as well as aspect, elevation and gradient of the dune face contributed to age-related
308 changes in community structure. Our study also suggests that due to its modulation of
309 microclimatic conditions, dune topography provides habitats conducive to forest-associated
310 species that have narrow climatic habitat tolerances. This suggests that even though these
311 new-growth forests are in transition, topography may influence the structure and composition



312 of biological communities of new-growth forests, especially when acting in concert with
313 other site-level factors. Such factors are likely to include those previously identified as
314 determinants of community structure and composition, such as landscape composition
315 (Grainger et al. 2011), and age (Wassenaar et al. 2005; Grainger & van Aarde 2012a).

316 The role of dune morphology seems more obvious in well-established ecosystems (Chen et
317 al. 1997; Oliviera-Filho et al. 1998; Tateno & Takeda 2003; Larkin et al. 2006), than the new-
318 growth forests that we studied, where age explained changes in assemblages better than
319 topography. Dune topography shaped as part of the rehabilitation procedure provides for the
320 topography that influences local conditions and therefore possibly for ecosystem patterns and
321 processes in a set manner according to prevailing climatic conditions. Topographically, these
322 dunes may differ from those shaped by natural forces (wind, water) which will probably
323 affect patterns and processes. However, these differences may be negligible and therefore not
324 be reflected in biological patterns, especially during the early stages of succession-driven
325 forest regeneration where most community variables vary with regeneration age.

326 For instance, age-related increases in habitat complexity provide an increasing variety
327 of conditions that accommodate more animal species and associated ecological processes
328 (Kritzinger & van Aarde 1998; Wassenaar et al. 2005). For example, increased plant
329 diversity, tree senescence and the associated development of a litter layer, increased soil
330 water retention, and nutrient accumulation would presumably benefit millipede communities
331 (e.g. Scheu & Schaefer 1998; Greyling et al. 2001; Berg & Hemerik 2004). In conclusion,
332 topography matters, more so for specialists than generalists. Response to topographic
333 variability is clearly species-specific and not necessarily reflected at the community level.

334

335



336 **Acknowledgements**

337 The authors declare that they have no conflict of interest. The study forms part of a larger
338 program conducted by the Conservation Ecology Research Unit (CERU), University of
339 Pretoria and financed by the Department of Trade and Industry and Richards Bay Minerals.
340 The authors also benefited from National Research Foundation grants. We thank members of
341 CERU that assisted with fieldwork and provided helpful comment on earlier versions of this
342 document. The authors declare that the research conducted as part of this study complied with
343 the requirements of South African legislation.

344

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460 bay and the Mfolozi River, 1937 - 1974. *Bothalia* 12:711-721.
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463 **Tables**

464 **Table 1** Key questions and hypotheses of this study

Key question	General assumptions	Hypotheses	Examples from the literature
1. Does dune topography influence abiotic conditions?	Dune topography modulates microclimatic conditions Dune topography influences soil nutrient availability	Irradiation, temperature and humidity may increase or decrease, depending exposure to wind and sun that is facilitated or eased by dune aspect, elevation, and position Soil carbon-to-nitrogen ratio and soil pH will be greater in valleys and at low elevations	(Tateno & Takeda 2003; Bennie et al. 2008) (Chen et al. 1997; Tateno & Takeda 2003)
2. Does dune topography influence biotic conditions?	Dune topography influences woody plant community structure and distribution Dune topography influences millipede community structure and distribution	<ul style="list-style-type: none"> Woody plant richness will depend on aspect, elevation and position depending on their exposure to wind Woody plant canopy structure will depend on gradient and position Species-specific woody plant abundances will differ based on dune morphology and position Millipede richness, as well as taxon-specific density may be influenced by aspect, elevation, and position depending on their exposure to wind and sunlight Species-specific millipede abundances will differ based on dune morphology and position 	(van Dyk 1996; Oliviera-Filho et al. 1998; da Silva et al. 2008; Laurance et al. 2010) (Weiss & Murphy 1990; Moir et al. 2009)

465

466 **Table 2** Definitions of response variables

Variable	Definition and units		
Microclimate	Temperature	Ambient temperature measured in degrees Celsius (°C)	
	Relative humidity	The partial pressure of water vapor measured as a percentage (%) of the saturated vapor pressure	
	Light intensity	Incident sunlight, measured as luminous power per area (illuminance) in lumens (lux)	
Soil	Soil pH	Soil acidity	
	Soil C:N	Carbon and nitrogen percentage content in soil samples presented as a ratio of carbon-to-nitrogen	
Response variables	Woody plants	Canopy tree species	Total number of species forming the canopy (height class 2-5, referred to as trees) per quadrat
		TH	Mean tree height (TH) class (2 [>2–4 m], 3 [>4–6 m], 4 [>6–8 m], and 5[>8 m]) of each quadrat
		CBH	Per-quadrat mean circumference at breast height (CBH), measurement carried out on all trees (height class 2-5) at ~1.4 m above ground
	Canopy tree density	Number of trees per 100 m ² , calculated for each quadrat	
	Understory species	Total number of species making up the understory (height class 1 [0-2m], referred to as understory plants) per quadrat	
	Understory density	Number of understory plants per 100 m ² calculated for each quadrat	
	Millipedes	Millipede species	Total number of millipede species per quadrat
Cylindrical density		Number of <i>Centrobolus</i> spp., <i>Doratagonus</i> sp., <i>Spinotarsus anguiliferus</i> , and <i>Spirostreptidae</i> spp. per 100 m ² calculated for each quadrat	
Keeled density		Number of <i>Gnomeskelus tuberosus</i> individuals per 100 m ² calculated for each quadrat	
Pill density		Number of <i>Sphaerotheridae</i> spp. individuals per 100 m ² calculated for each quadrat	

468 **Table 3** Dune morphological variables included in the most parsimonious models (based on Akaike
469 Information Criteria (AIC) scores) explaining variance in abiotic and biotic variables for each of
470 three seral stages and pooled stages, as well as the significance of the model ($p < 0.05$). Those
471 response variables that were explained by dune morphological variables in the absence of age are
472 highlighted in boldface text.

Response variables	Seral stage	Explanatory variables					ANCOVA results		
		Dune morphology					df	AIC	P
		Aspect	Elevation	Gradient	Position	Age			
Soil	1	Insufficient cases							
	2					X	1	54.35	0.0005
	3	X	X		X	X	12	13.73	< 0.0001
	Pooled					X	1	157.54	< 0.0001
	1	Insufficient cases							
	2			X			4	284.46	0.119
	3					X	1	186.59	0.745
Pooled					X	1	542.26	0.778	
Woody plants	1	X	X			X	9	23.03	< 0.0001
	2		X			X	5	23.77	< 0.001
	3	X		X	X	X	12	46.27	< 0.001
	Pooled		X			X	5	236.79	0.024
	1	X	X		X	X	10	158.23	< 0.0001
	2					X	1	280.81	< 0.001
	3	X		X		X	10	229.17	< 0.001
Pooled	X	X	X		X	16	787.49	< 0.001	
Mean canopy tree CBH	1	X	X	X	X		13	125.49	0.002
	2		X			X	5	195.20	< 0.001
	3			X		X	3	169.77	0.015
	Pooled		X		X	X	7	528.67	< 0.001
Number of species in canopy	2		X			X	5	195.20	< 0.001
	3			X		X	3	169.77	0.015
	Pooled		X		X	X	7	528.67	< 0.001



	1				X	1	-10.80	0.009	
Mean canopy tree density	2	X	X	X	X	15	-120.87	< 0.001	
	3			X		2	-120.89	0.0004	
	Pooled	X	X		X X	14	-155.73	< 0.001	
	1	X	X	X	X X	14	97.78	0.009	
Number of species in understory	2			X	X	5	105.12	< 0.001	
	3			X	X X	5	57.26	0.0006	
	Pooled	X		X	X	12	289.13	0.0008	
	1				X	1	74.66	0.679	
Mean understory density	2	X	X	X	X	15	-171.69	< 0.001	
	3				X	1	-113.38	0.0005	
	Pooled				X	1	-305.24	0.003	
<hr/>									
	1	Insufficient cases							
Number of species	2				X	2	126.00	0.016	
	3	X	X	X	X X	14	83.35	< 0.001	
	Pooled				X X	3	271.23	< 0.001	
<hr/>									
Millipedes	1	Insufficient cases							
	Cylindrical millipede density	2		X		X	6	92.18	< 0.0001
		3	X	X		X X	13	-18.05	< 0.001
		Pooled	X	X	X	X X	20	183.63	< 0.001
	<hr/>								
	1	Insufficient cases							
	Keeled millipede density	2	X	X	X	X X	15	-114.18	< 0.0001
		3	X	X		X X	12	-96.33	0.0004
		Pooled	X	X	X	X	9	-235.50	< 0.001
	<hr/>								
1	Insufficient cases								
Pill millipede density	2	X	X	X		X	14	-221.88	0.004
	3	X		X	X X	12	-110.06	0.0001	
	Pooled	X	X	X		X	18	-400.40	< 0.001

474 **Table 4** Characteristic tree species (taller than 2 m) forming the canopies on slopes of different
 475 gradients in seral stage two based on similarity percentage analysis (SIMPER).

Species	Family	Average abundance	Average similarity	Similarity SD	Percentage contribution	Cumulative percentage
<i>0-5 degree slope</i>						
Average similarity: 54.91						
<i>Acacia karroo</i> Hayne	Mimosaceae	3.26	39.44	6.11	71.83	71.83
<i>Celtis africana</i> Burm.f.	Celtidaceae	0.92	7.70	1.12	14.02	85.85
<i>Allophylus natalensis</i> Sond.	Sapindaceae	0.76	4.90	0.88	8.93	94.78
<i>6-10 degree slope</i>						
Average similarity: 48.96						
<i>Acacia karroo</i>	Mimosaceae	3.39	35.83	2.44	73.18	73.18
<i>Allophylus natalensis</i> Sond.	Sapindaceae	0.70	4.41	0.98	9.01	82.19
<i>Celtis Africana</i> Burm.f.	Celtidaceae	0.99	3.32	0.78	6.78	88.97
<i>Cestrum laevigatum</i> Schlttdl.	Solanaceae	0.52	1.19	0.41	2.42	91.39
<i>11-15 degree slope</i>						
Average similarity: 52.42						
<i>Acacia karroo</i> Hayne	Mimosaceae	3.44	40.69	4.33	77.62	77.62
<i>Celtis Africana</i> Burm.f.	Celtidaceae	0.70	3.71	0.72	7.07	84.70
<i>Brachylaena discolor</i> (DC.)	Asteraceae	0.35	1.67	0.45	3.19	87.89
<i>Allophylus natalensis</i> Sond.	Sapindaceae	0.47	1.54	0.37	2.94	90.82
<i>16-20 degree slope</i>						



Average similarity: 50.52

Acacia karroo Hayne Mimosaceae 3.49 26.86 4.48 53.17 53.17

Celtis Africana Burm.f. Celtidaceae 2.08 14.93 5.77 29.55 82.72

Grewia occidentalis L. Tiliaceae 0.87 2.47 0.56 4.89 87.61

Chaetachme aristata Ulmaceae 0.55 1.03 0.37 2.04 89.66

Planch.

Teclea gerrardii Rutaceae 0.30 0.87 0.39 1.72 91.38

I. Verd.

>20 degree slope

Less than two samples in a group

476

477

478 **Table 5** Characteristic species occurring in the understory of each elevation category within seral
479 stage three based on similarity percentage analysis (SIMPER).

Species	Family	Average abundance	Average similarity	Similarity SD	Percentage contribution	Cumulative percentage
21-40 m.a.s.l.						
Less than 2 samples in group						
41-60 m.a.s.l.						
Average similarity: 44.51						
<i>Rhoicissus revoilii</i>	Vitaceae	1.49	27.25	1.10	61.23	61.23
Planch.						
<i>Scutia myrtina</i>	Rhamnaceae	0.95	12.50	0.69	28.09	89.32
Burm.F.						
<i>Chromolaena odorata</i>	Asteraceae	0.57	4.75	0.46	10.68	100.00
L.						
61-80 m.a.s.l.						
Average similarity: 30.00						
<i>Scutia myrtina</i>	Rhamnaceae	0.87	30.00	0.76	100.00	100.00
Burm.F.						

480

Figures

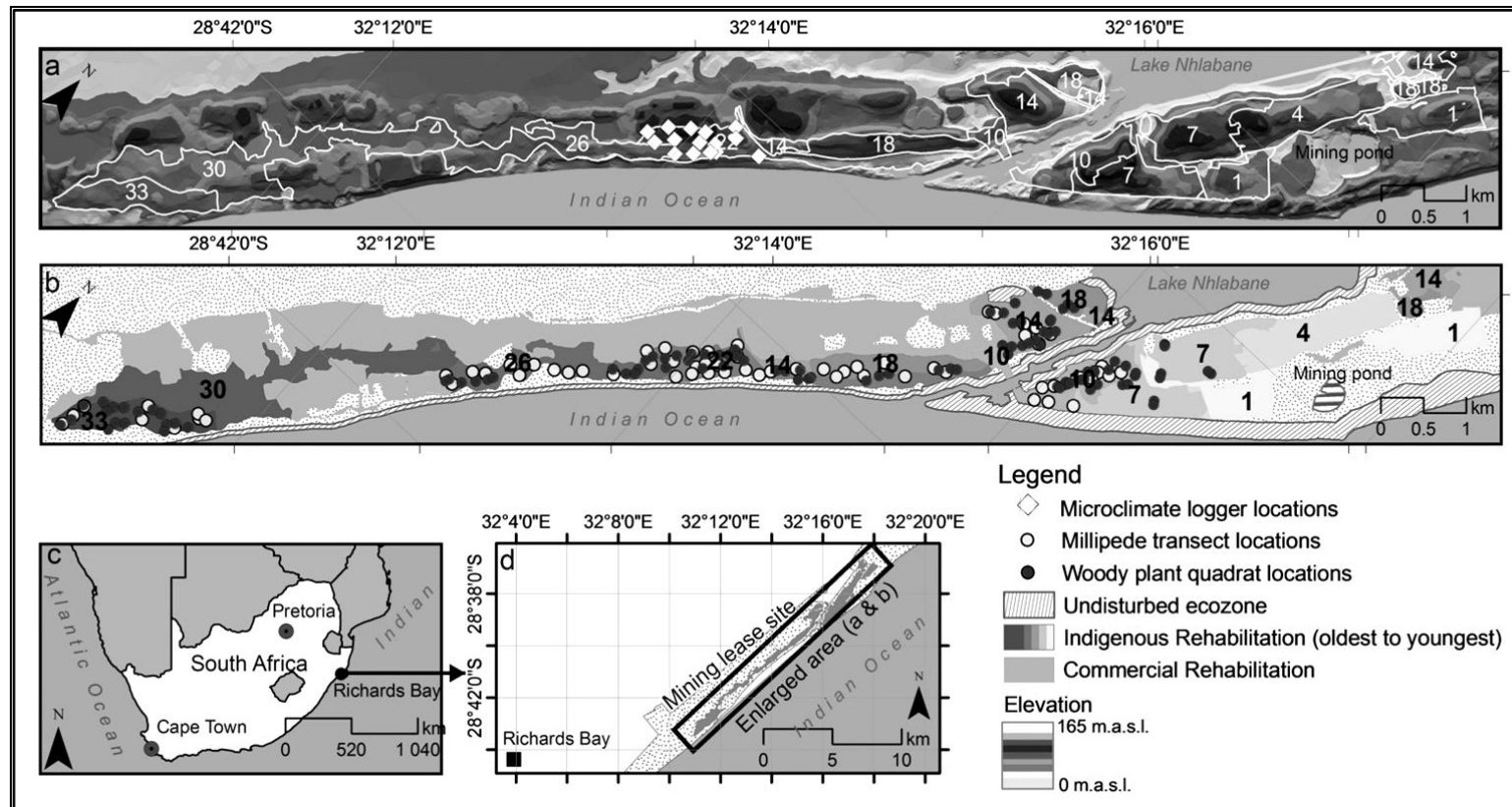


Fig. 1 Digital elevation model of the study area also showing the delineation of rehabilitating stands according to age, and the sites where data loggers were deployed (a). The locations of woody plant quadrats and millipede transect surveys were conducted are shown in relation to stand age (b). Inset maps provide geographical context (c & d).

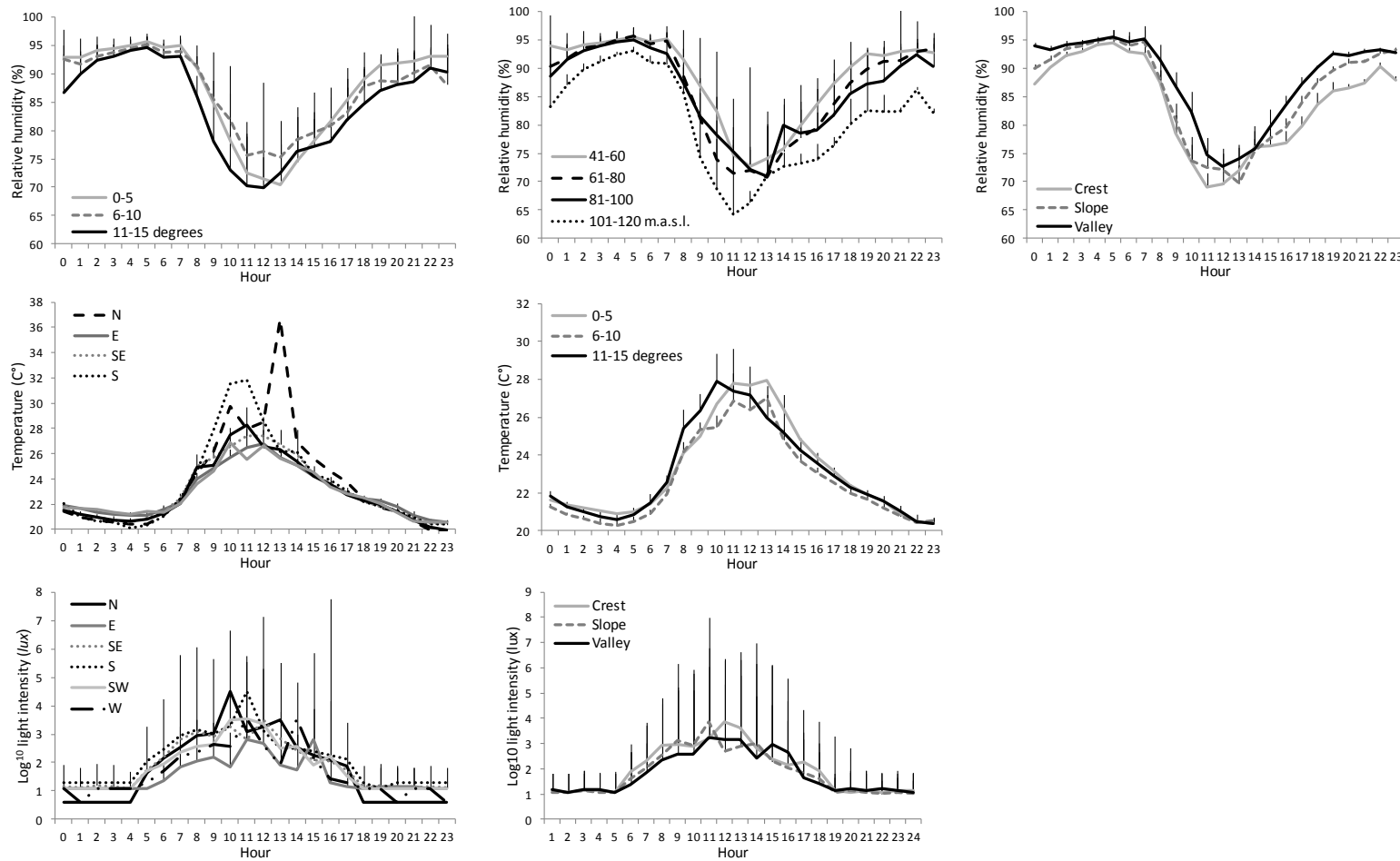
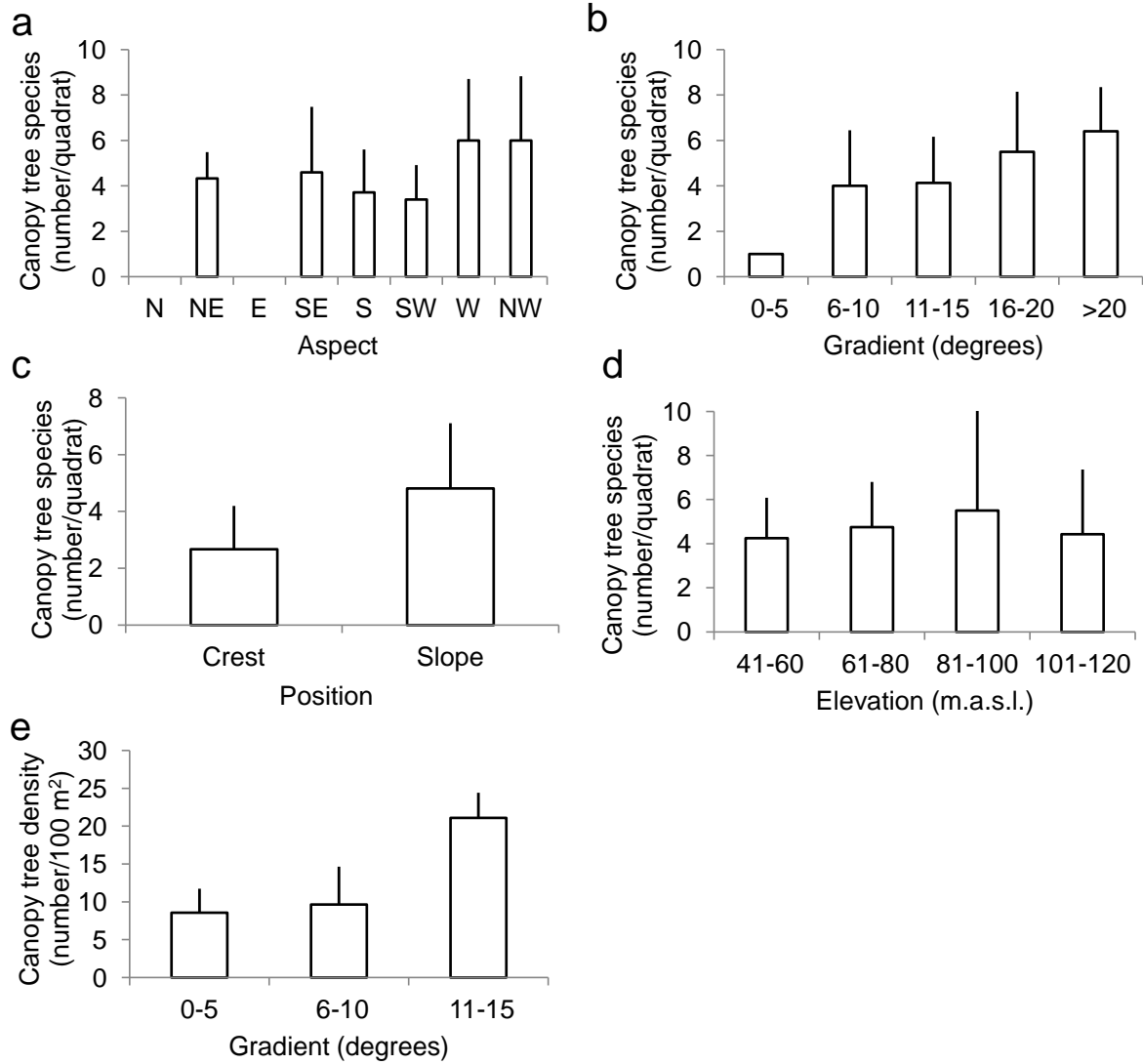


Fig. 2 Mean \pm one standard deviation of the mean of three microclimatic variables (relative humidity, temperature, and light intensity, from top to bottom, respectively) that showed significant responses to variables of dune morphology according to the repeated measures ANOVA, as recorded between 01h00 and 24h00 on the 29th of January 2011.

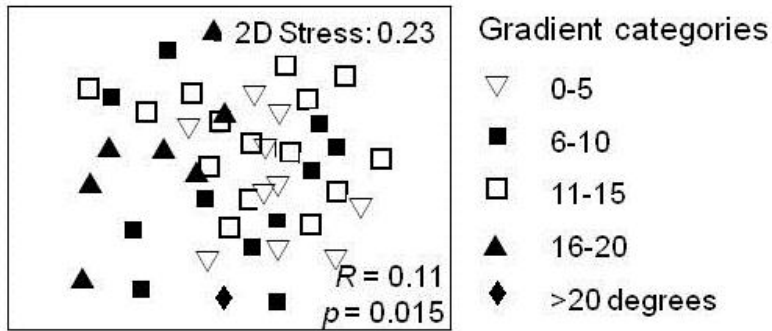


1

2 **Fig. 3** Mean \pm one standard deviation of the mean of woody plant response variables
3 presented as a function of those variables that best-explained their variability significantly
4 despite stand age (see Table 2).

Seral stage 2

Species-specific canopy abundances



Seral stage 3

Species-specific understory abundances

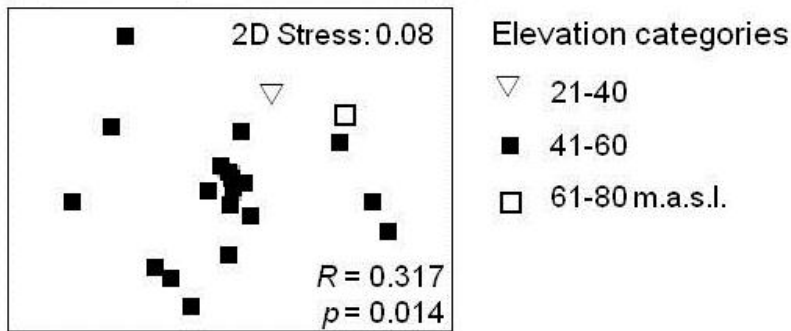
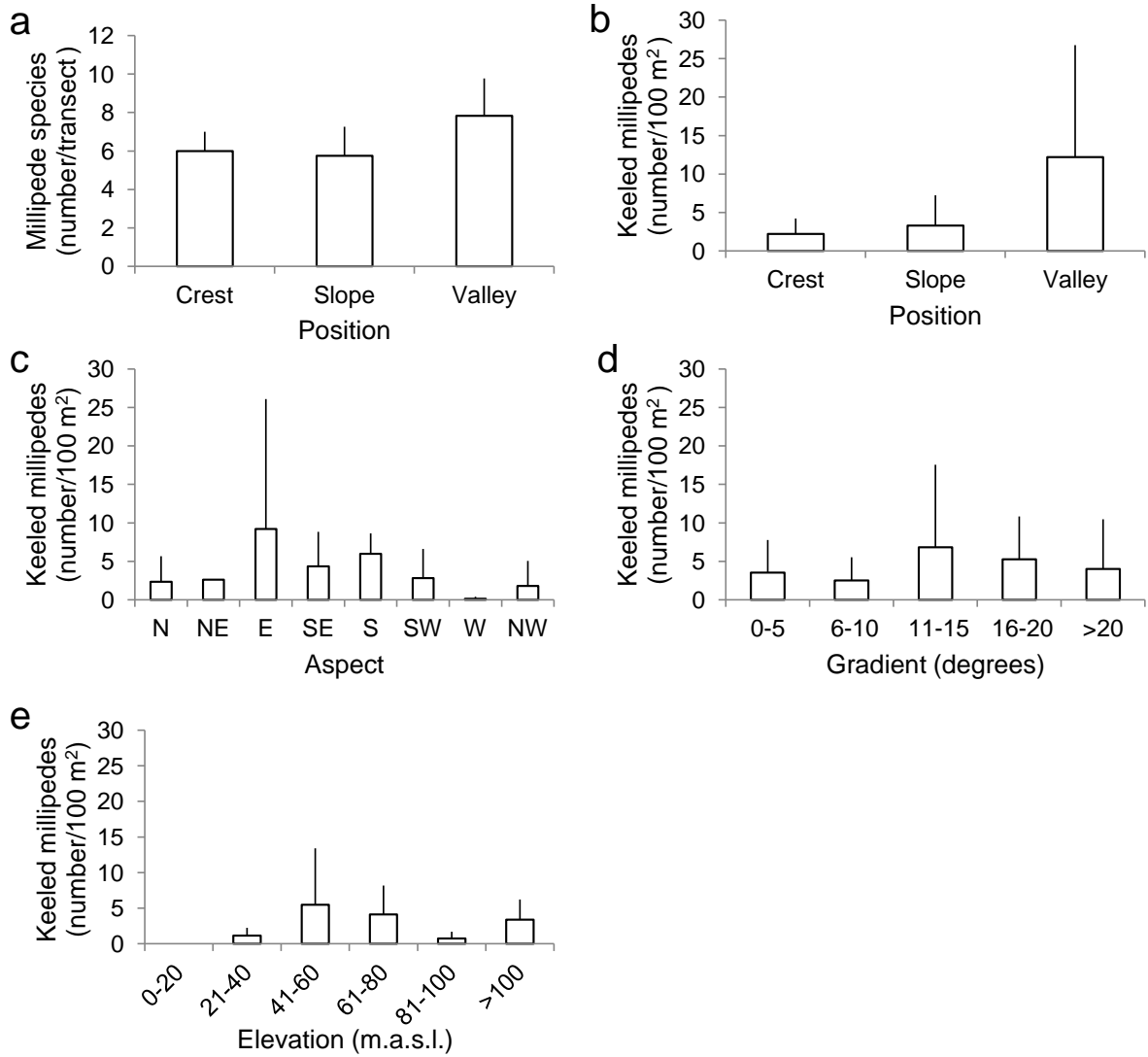


Fig. 4 Non-metric multi-dimensional scaling (NMDS) plots of woody plant abundances in the canopy (top) and understory (bottom) where analysis of similarity revealed significant ($p < 0.05$) community separation attributable to dune morphological characteristics (elevation, gradient, position) according to seral stages two (11-25) and three (>25 years), respectively.



6

7 **Fig. 5** Mean \pm one standard deviation of the mean of millipede response variables presented
8 as a function of those variables that best-explained their variability significantly despite
9 stand age (see Table 2).