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 Therefore, precautions must be taken to eliminate azo dyes from waste waters and to avoid human exposure to the most dangerous ones.

The problem is also related to the worldwide growing demand for water with high enough quality, necessity to protect water resources and to purify the wastewaters [3]. The overall increase of the industrial production capacity (textile, mining, papers, plastics, etc.) lead to extensive usage of water and obtaining of large volumes wastewaters. Dying and textile industry are among the largest consumers of water and around 15% of the dyes are released in the environment without proper treatment [4]. These practices pose serious environmental risks and hence proper purification methods and technologies should be investigated and introduced.

Acid orange 7 is an azo dye with an appealing red-orange color. It is mainly used for silk and wool dyeing, polyamide fiber fabric of direct printing, leather and paper coloration, hair dyeing products, and is also utilized as an indicator and biological stain [5]. This azo dye is produced by coupling of *p*-aminobenzenesulfonic acid (after diazotation) and 2-naphtol. Reductive cleavage by bacteria or degradation by chemical or physical factors could result in the production of corresponding aromatic amino compounds. Brüschweiler and Merlot [6] reported that the *o*- and *p*-aminobenzenesulfonic acid could be mutagenic and are found in some wool, cotton and other cellulose fibers. *P*-aminobenzenesulfonic acid was amongst the most frequently determined aromatic amines, found in 37 textile samples in concentrations in the $100-442$ mg/kg range [5].

 Various methods for decontamination and removal of compounds like azo dyes were proposed and investigated. Some of these methods are based on physicochemical interactions (such as flocculation, adsorption, ozonation, electrolytic and photooxidation, etc.), phytoremediation [7] and others rely on microbial or fungal degradation [8-12]. Taheri et al.

 [13] performed a study for the removal of two azo dyes: Acid orange 7 and Acid red 18 by

several methods to compare the costs and effectiveness of the methods for decontamination. The authors found that the utilization of activated carbon in combination with membrane filtration in a sequencing batch reactor showed the most promising results. However, the initial investments are usually higher than performing a simple adsorption or microbial/fungal degradation. Numerous adsorbents have been tested for the removal of azo dyes [13-16] with the most significant advantage being the utilization of renewable and cheap resources as adsorbents. Such abundant and easily available materials are very often by-products / wastes from the food and agricultural industries which generate large amounts of biodegradable phytomass annually.

 lignocellulosic materials and polyphenols [20-25]. 109 111 114 116 118 The essential oil industry, because of the usually low amount of the essential oil in the raw plant material, produces tonnes of solid by-products which are commonly discarded [17]. Lavender and roses are amongst the most processed crops in the essential oil industry and the by-products obtained from their distillation could be used as a cheap, abundant, and biodegradable adsorbent for pollutant removal. The solid lavender and rose post-distillation biomass have been successfully employed in numerous works for the removal of heavy metals in model studies and in waste waters [17] but few studies have specifically focused on the application for the adsorption of organic pollutants [18, 19]. Additionally, the residues from the adsorption of azo dyes could be subjected to further degradation by higher fungi, which are known for their abilities to produce enzymatic complexes able to degrade

Hence the aim of the present work was to investigate the applicability of rose and lavender industrial post-distillation by-products as natural adsorbents for removal of the azo dye Acid orange 7 and to study the factors, kinetics, thermodynamic parameters, and further fungal biodegradation of the post-adsorption residues. The study proposes an innovative method for lavender and rose post-distillation by-products valorization. It is known that due to their high

polyphenolic compounds content their direct disposal in landfills are likely environmentally harmful and therefore the method would contribute to sustainability, a circular economy, and zero waste approaches in general.

2. Materials and methods

2.1. Samples and reagents

 from local distributors (Table 1). 134 136 141 142 143 Lavender (bio-certified *Lavandula angustifolia* Mill., "Sevtopolis" cultivar) – L and rose (bio-certified *Rosa Damascena* Mill.) – RD, by-products from industrial steam distillation of fresh plant stems and flowers were provided by the ECOMAAT distillery (Mirkovo, region of Sofia, Bulgaria; 2021 harvest). The L and RD by-products were collected when the distillation was finished: for RD residues the mass was pumped out from the still and after passing through a screw conveyor the solid residue was separated from the liquid phase; for L residue the mass was removed from the baskets in the still and collected. The solid residues were inspected for impurities and dried. The dried rose and lavender were washed with distilled water (100 g with 1600 mL water; added in portions of 400 mL) on a Buchner funnel. The residual mass was dried, milled, and sieved. For adsorption of the azo dye, the biomass fraction with particle size 100-250 μm were used. Acid orange 7 (4-(2-Hydroxy-1-naphthylazo) benzenesulfonic acid sodium salt) was obtained

2 3			Molecular	Mw	Density,	Solubility,		
	Name	Structure	formula	(g/mol)	g/mL	g/L (20 \textdegree C)		
7 8 9 10 11 12 13 14 15 16 17 18 19 20 21 22	Acid orange 7	SO_3 Na ⁺ ,OH	$C_{16}H_{11}N_2NaO_4S$	350.3	1.6	27		
23 151 24								
25 152 26 27	2.2. Methods							
28 153 $\frac{29}{30}$								
154 31 32	2.2.1. Adsorption procedure							
155 33 34	The adsorption of the azo dye was carried out as follow: 20 mL water solution of Acid orange							
35 156 36	7 with specified concentrations (20, 40, 60, 80 and 100 mg/L) were added to 1 g adsorbent in							
37 157 38 39	a 50 mL centrifuge tube and the tubes were placed on a laboratory shaker MLW THYS 2							
40 158 41		(VEB MLW Labortechnik Ilmenau, Germany), placed in a thermally controlled laboratory						
42 159 43	oven (the experiments were performed at three temperatures: 293 K, 303 K and 313 K). The							
44 45 160 46		shaker was started (100 rpm) and at a specified time $(2.5, 5, 10, 20, 40, 60, 90, 120)$						
47 161 48	minutes) a centrifuge tube was removed from the shaker and the mass was filtered first							
49 162 50	through a paper filter and second through a syringe filter $CA\,0.45 \,\mu m$ (Isolab, Germany).							
51 52 163 53		Adsorption of the filtrate was measured at 500 nm using LLG-uniSPEC 2 UV-Vis						
54 164 55		spectrophotometer (LLG Labware, Germany). The concentration of non-adsorbed Acid						
56 57 165 58		orange 7 was calculated using a calibration curve, prepared with water solutions of the dye						
59 166 60		with known concentrations. The adsorption of Acid orange 7 by L and RD residues at						
61 62			$\overline{7}$					
63 64 65								

 Table 1. Characteristics of the azo dye Acid orange 7

 where K_F is Freundlich constant associated with adsorption capacity, L/g ; *n* is an empirical constant connected to adsorption intensity. The linear form of Eq. 4 is expressed as $[27]$: 240 further assumes that the intermolecular attractive forces diminish steadily with distance increases. The original non-linear form of the Langmuir equation is expressed as: $q_{e=\frac{qmK_{L}c_{e}}{1+K_{L}c_{e}}}$ Eq. 2 where q_m is the maximum adsorption capacity, mg/g; K_L is Langmuir constant associated with the affinity of the binding sites and energy of adsorption, L/g. The linear form of Langmuir Eq. 2 is expressed as [27]: c_e $\frac{c_e}{q_e} = \frac{c_e}{q_m} +$ $rac{1}{q_m K_L}$ Eq. 3 2) Freundlich isotherm model The Freundlich model postulates a multilayer adsorption process, with non-uniform distribution of adsorption enthalpy and affinities onto the heterogeneous adsorbent surface without lateral interaction. The energetically favored binding sites are theorized to be occupied first and the binding strength decreases sequentially with increased coverage of the binding sites. The non-linear Freundlich equation is expressed as [28]: $q_e = K_F c_e^{1/n}$ Eq. 4 248

 The coefficients *k*d were calculated by the method suggested by Khan and Singh [30] from the intercept of the plots of ln (*q*e/*c*e) vs. *qe*. The coefficient *k*d has the dimension L/g. It needs 43 ²⁸⁰ 264 ൌ Eq. 5 265 266 267 268 269 ²⁷⁰ 271 3) Temkin isotherm model The Temkin isotherm model considers the effect of indirect adsorbate-adsorbing species interactions on adsorption. It postulates that the heat of adsorption as a function of temperature of all the molecules in the layer reduces linearly rather than logarithmically with surface coverage due to adsorbent–adsorbate interactions [29]. The non-linear Temkin isotherm equation is given by: 272 ²⁷³ ൌ ሺ ሻ ൌ ሺ ሻ Eq. 6 274 275 276 277 278 279 281 46 282 283 where *T* is the absolute temperature, K; *R* = 8.314 is the universal gas constant, J/(mol.K); *AT* is the equilibrium binding constant, L/g; *b* is the variation in adsorption energy, J/mol; *B* is the Temkin constant. The linear form of Temkin equation is expressed as [27]: ൌ Eq. 7 4) Thermodynamic parameters determination Thermodynamic parameters were determined using the thermodynamic equilibrium distribution constants *k*d at the studied temperatures: 284 ൌ Eq. 8

 3. Results and discussion 3.1. Kinetics of the adsorption of Acid orange 7 on L and RD by-products One of the most important parameters influencing the adsorption process is the contact time of the solution with the adsorbent. Fig. 1 shows the kinetics of the adsorption and the 288 to be converted into dimensionless form K_e by multiplying with the density of water (1000) g/L) to be used in the thermodynamic equation [31]: $K_e = 1000 k_d$ Eq. 9 Then, the change in Gibbs free energy ΔG^0 , J/mol, enthalpy ΔH^0 , J/mol, and entropy ΔS^0 , were calculated from the Van't-Hoff equation: $-\Delta H^0$ 1 $ln K_e = \left(\frac{-\Delta H^0}{R}\right) \frac{1}{T} + \frac{\Delta S^0}{R}$ Eq. 10 296 From the plot of ln*Ke* as a function of $1/T$ (van't Hoff plot) according to Eq. 10 the values of Δ*H*0 and Δ*S*0 J/(mol.K), were calculated using the slope and intercept of the straight lines respectively. The values of ΔG^0 were calculated using Eq. 11. $\Delta G^0 = \Delta H^0 - T \Delta S^0$ Eq. 11 2.3. Statistical analysis The experiments were performed in triplicate and data values are expressed as mean \pm SD (standard deviation). Statistical analysis was carried out by one way ANOVA (Tukey's *post hoc* test; $p < 0.05$).

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	c_e , mg/L							
$C0$,		L			RD			
mg/L	293K	303K	313K	293K	303K	313K		
20	5.85 ± 0.15^a	$5.37 \pm 0.44^{a,b}$	4.19 ± 0.42^b	3.27 ± 0.20 ^c	2.92 ± 0.12 ^{c,d}	2.51 ± 0.29 ^d		
40	8.45 ± 0.08^b	9.61 ± 0.49^a	6.74 ± 0.32 ^d	7.66 ± 0.37 °	5.87 ± 0.24 ^e	5.06 ± 0.29 ^f		
60	$12.21 \pm 0.21^{a,b}$	$12.76 \pm 0.62^{\text{a}}$	11.60 ± 0.51^b	8.21 ± 0.20 ^c	7.90 ± 0.19 ^c	6.42 ± 0.20 ^d		
80	14.47 ± 0.14^b	$16.55 \pm 0.29^{\mathrm{a}}$	15.18 ± 0.26^a	10.68 ± 0.27 c	9.19 ± 0.35 ^d	8.45 ± 0.22 ^d		
100	21.17 ± 0.24 ^a	$19.65 \pm 0.45^{\rm b}$	18.54 ± 0.24 ^c	14.68 ± 0.18 ^d	13.60 ± 0.21 ^e	12.98 ± 0.79 ^f		

313 K for L and RD. Amount of RD and L adsorbents: 1 g; volume of dye solution: 20 mL; pH=7.05; shaking speed 100rpm

Table 2. Equilibrium concentration (*^c*e, mg/L) at 20, 40, 60, 80 and 100 mg/L initial concentrations (*^c*0, mg/L) and temperatures 293, 303 and

365 Table 3. Influence of the pH medium on the adsorption of Acid orange 7 on L and RD.

²366 Adsorbed amount: mg/g (in brackets is given the efficiency of the adsorption in percentage).

367 Adsorbent: 1g; initial azo dye: 20 mL 100 mg/L (2 mg dye in the volume); temperature 313

370 371 372 32 373 34 374 Table 4 presents the results for adsorption of Acid orange 7 by L and RD compared with different adsorbents found in the literature. Although, the L and RD residues showed lower adsorption capacities compared with other adsorbents, it should be noted that the results were achieved at neutral pH values. At acidic pH medium the effectiveness of the azo dye removal reached 98.75% for both adsorbents.

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376 Table 4. Comparison of adsorption capacities of different adsorbents of Acid orange 7

$40\,$							
41 42	Adsorption capacity						
43 44	Adsorbent	Conditions	(mg/g)/removal	References			
45							
46			efficiency, %				
47							
48							
49	Salvinia natans		46.53/				
50		293 K, pH 1		$[36]$			
51	biomass		46.53%				
52							
53			24.23/				
54	Canola waste						
55		288 K		$[37]$			
56			$>99\%$				
57							
58							
59							
60							
61							
62			18				
63							

64 65

 $\frac{1}{2}$ 404 405 6
7 407 $\begin{array}{c} 11 \\ 12 \end{array}$ 408 $\frac{14}{15}$ 409 16
17 412 $^{23}_{24}$ Furthermore, the morphology of the adsorbents and the post-adsorption residues of RD and L were investigated by SEM and the observed image is presented in Fig. 3. The RD by-product image revealed the presence of a fiber-type structure probably due to polysaccharide chains while the L by-product could be characterized with a fluffier and more porous surface. The post-adsorption residues generally retain their structure. Although, some studies [16, 42] suggested that there is a significant difference in the adsorbent structure before and after adsorption, in our case no observable changes in the RD and L post-adsorption residues morphology, compared to the RD and L by-products, could be observed. This could be explained with the physical type of the adsorption process of the Acid orange 7 onto the surface of the RD and L by-products.

415 **Fig. 3** Morphological analysis by SEM of the RD (a. before adsorption; b. after adsorption)

 $\frac{45}{16}$ 416 and L (c. before adsorption; d. after adsorption) residues

3.3. Adsorption isotherms parameters determination

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50 421 the liquid and solid phase [43]. To evaluate the applicability of different isotherm models The adsorption isotherm characterizes the equilibrium state of molecules distribution between

60 422 (Langmuir, Freundlich, Temkin) to describe the process of adsorption their linearized form

was used (Eq. 3, 5 and 7, section 2.2.5). The Langmuir isotherm model assumes that adsorption occurs on a relatively homogenous surface with limited interactions in the surface plane. The Freundlich model empirically describes adsorption on non-homogenous, porous surfaces. The Temkin isotherm model describes interactions of heterogeneous solid surfaces with molecules from a solution and this assumption is made on the basis that temperature increases lead to decreases of surface covering. The correlation of the models to the experimental results is expressed by fitted plots and presented in Fig. 4.

residues for 293, 303 and 313 K **Fig. 4** Langmuir, Freundlich and Temkin isotherms of Acid orange 7 adsorption on L and RD

Table 5. Parameters of the Langmuir, Freundlich and Temkin models of adsorption of Acid orange 7 on L and RD at 293, 303 and 313 K

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 $\frac{48}{49}$ 504 51 505 $\frac{53}{1}$ 506 adsorption residues as feeding substrates were conducted. The twenty-days cultivations, comprising of several stages, resulted in homogenous mycelium-based bio-composites with a desired morphology (Fig. 6). Both obtained materials demonstrated acceptable conformational stability and had velvety tactile surfaces that was the result of well-developed aerial mycelium and dense hypha networks. These characteristics are of important $\frac{1}{2}$ 485 486 6
7 488 $\begin{array}{c} 11 \\ 12 \end{array}$ 489 $\frac{14}{15}$ 490 16
17 $18 \\ 19$ 493 $^{23}_{24}$ 495 28
29 498 35
36 500 $^{40}_{41}$ 502 503 often proposed solution but again the problem with safely handling the washings arises. The metabolic activity of different higher fungi and their immense capacity to grow and degrade different natural and synthetic substrates could support the conversion of these postadsorption residues into value added products with low ecological footprints. Angelova et al. [20] demonstrated recently that lignocellulose compounds, together with the non-cellulosic polysaccharides and uronic acids polymers of steam distilled lavender straw and hexane extracted rose flowers were essential prerequisite for vegetative growth of Basidiomycetes fungi. The above chemical compounds stimulated mycelium net formation and supported the full lignocellulose substrates colonization that at the end yielded innovative mycelium-based materials with desirable properties. *G. resinaceum* was chosen not only because of the strain's ability to grow well on different lignocellulose, but also it's potential to degrade synthetic dyes. *G. resinaceum* belongs to white-rot fungi (*Basidiomycota* phylum*,* order *Polyporales*). Due to their ability to produce extracellular lignin-degrading enzymes (lignin peroxidases, manganese peroxidases and laccases) in submerged or solid-state cultivation conditions, white-rot fungi have their advantages in mycoremediation processes for degradation of insoluble chemicals and hazardous environmental pollutants [45, 46]. In many studies focused on decolorization of synthetic dyes the "*key players*" are white-rot fungi such as *Trametes versicolor, Pleurotus ostreatus, Ganoderma spp.,* etc*.* [22, 24, 25, 47]. In the present study solid-state cultivations of *G. resinaceum* GA1M using L and RD post-

 significance for the physical and mechanical properties of these mycelium-based biocomposites which in turn determine the application of the developed biomaterials [20]. The obtained mycelium-based bio-composites were analyzed for residual Acid orange 7. The initial concentration of Acid orange 7 in spent adsorbents was in the 1.2-1.3 mg/g adsorbent range. After the fungal growth, transformation and degradation of the lignocellulosic matrix, metabolization of Acid orange 7 also occurred and the dye was not detected in either mycelium-based materials.

 The azo bond in the azo dyes is the main reason for their persistency since it is nonbiodegradable under aerobic conditions. As a member of the Basidiomycetes, *G. resinaceum* possess a ligninolytic enzyme system with very broad substrate specificity which can break the recalcitrant azo bond. It is well known that degradation is not always equal to detoxification, but the Basidiomycetes representatives are able to decolorize and detoxify synthetic dyes and their metabolites. The enzymes that are most extensively used (dependently and independently) in the synthetic dyes' degradation are laccases, lignin peroxidases and manganese peroxidases [21].

- **Fig. 6** Mycelium-based biocomposites produced from the L and RD residues resulted from
- the adsorption of Acid orange 7

 The Basidiomycetes *G. resinaceum* proved to be effective in removing the dye Acid orange 7 during solid-state cultivation on post-adsorption L and RD residues and simultaneously leading to innovative mycelium-based bio-composites. This approach is consistent with *zero waste* trends which will benefit overall environmental sustainability. Further research is needed to analyze physico-mechanical properties of the produced biomaterials and identify potential applications.

4. Conclusions

due to the aromatic ring stretching vibrations from the aromatic moieties of Acid orange 7 suggesting successful adsorption of the dye onto the surface of the RD and L by-products. The adsorption of the azo dye was evaluated by Langmuir, Freundlich, and Temkin isotherm models and it was found that the Freundlich equation best described the adsorption of Acid orange 7 using RD and L by-products. The coefficient n in the Freundlich equation connected The present study focused on decontamination and removal of azo dye Acid orange 7 from aqueous solutions by combined treatment using industrial rose and lavender solid by-products as adsorbents and further degradation of the post-adsorption plant biomass by higher fungi resulting in production of mycelium-based biocomposites. The results suggested that more than 87% of the colorant was removed by the RD adsorbent and around 82% by L residue for one hour at 293 K. The highest adsorption capacity was demonstrated by both adsorbents with the highest initial concentration of 100 mg/L of the azo dye. The effectiveness of the adsorption increased decreasing the pH medium: for L at pH 1.5 the removal was 1.975±0.08 mg/g (98.75%) and for RD at pH 3.5 was 1.975 ± 0.08 (98.75%). The infrared spectroscopy analyses revealed that around $1500-1560$ cm⁻¹ a new peak at 1540 cm⁻¹ (for RD postadsorption residue) and 1541 cm^{-1} (for L post-adsorption residue) appeared. These bands are

 advanced wastewater treatment technologies. Bioresour Technol 344:126249.

