



## Rapid susceptibility of Carbapenem resistant *Pseudomonas aeruginosa* and its resistance gene to non-thermal plasma treatment in a batch reactor

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### ARTICLE INFO

Editor: Sadao Araki

#### Keywords:

Carbapenems  
Cold atmospheric plasma  
Disinfection  
*Pseudomonas aeruginosa*  
Wastewater treatment plants

### ABSTRACT

The critically ranked carbapenem-resistant *Pseudomonas aeruginosa* has been observed to infect immunocompromised patients that consume polluted waters, leading to critical infections and more hospital costs. To save lives and unburden the public health sectors of preventable costs, non-thermal plasma (NTP) technology was investigated as an alternative disinfection step that could be applied in wastewater treatment plants (WWTPs) to inactivate this bacterium and its prominent carbapenem resistance gene (*bla<sub>NDM-1</sub>*). Culture and molecular-based techniques were employed to confirm carbapenem resistance in *P. aeruginosa* (27853). Culture suspensions of carbapenem-resistant ATCC *P. aeruginosa* (16 h culture) were prepared from confirmed isolates and subjected to plasma treatment at varying time intervals (3 min, 6 min, 9 min, 12 min and 15 min) in triplicates. The plasma treated samples were evaluated for re-growth and the presence of *bla<sub>NDM-1</sub>*. The treatment resulted in a 0.68 log reduction after 3 min and the highest log reduction of  $\geq 8$  after 12 min, suggesting that plasma disinfection has a great potential to be an efficient tertiary treatment step for WWTPs. Moreover, the gel image showed that band intensity of *bla<sub>NDM-1</sub>* reduced with treatment time, thereby suggesting a probable reduction of amplified genes. Notwithstanding, longer treatment time, a grounded electrode with a larger surface ( $\geq 40$  mm diameter) and/or oxygen-containing feeding gas is warranted to completely inactivate its antibiotic resistance gene (ARG), which might be bound by biofilms as they seem to protect *P. aeruginosa* from the action of non-thermal plasma (NTP) disinfection.

### 1. Introduction

Wastewater treatment plants (WWTPs) are regarded as important hotspots for spreading antibiotic resistance in different bacteria, when compared to other water environments [1]. This is because conventional WWTPs were not designed to remove the antibiotic resistant bacteria (ARB) and genes (ARGs) [2] that are often reticulated from the sewage systems of households, healthcare services, antibiotic manufacturing facilities, agricultural activities and animal feedlots [3–5]. Moreover, conventional disinfection processes, such as chlorination, UV irradiation and ozone oxidation, which are applied in WWTPs, only kill a great fraction of ARB, whereas others enter a state of dormancy due to stress and are resuscitated when the stressors are released. Sometimes, they decrease the abundance of the genes (gene copies per mL of sample)

while the prevalence of the gene (gene copies per total bacteria) increases [6,7]. Whilst some of the disinfection processes kill the bacteria, ARGs may still persist for a long time, after discharge from WWTPs in cell debris and in the environment, transferring and adapting into microbiota of natural water bodies (drinking water sources), eventually leading to the development of antibiotic resistance [6,8–10]. As a result, WWTPs keep recycling drinking water polluted with high concentrations of ARB and ARGs, back to municipal water supplies and health care facilities [11]. *Pseudomonas aeruginosa*, a carbapenem-resistant bacteria that has been categorised as critical priority 1 class by the World Health Organisation (WHO) [12,13], is an example of an ARB found in the effluent of WWTPs, in lakes, rivers and swimming pools [1]. Moreover the compromise of the clinical efficacy of carbapenems, such as meropenem and imipenem is due to the spread of the carbapenemase New

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<https://doi.org/10.1016/j.jwpe.2024.105915>

Received 14 April 2024; Received in revised form 24 July 2024; Accepted 1 August 2024

Available online 3 August 2024

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Delhi metallo- $\beta$ -lactamase-1 plasmids ( $bla_{NDM-1}$ ), which also induce *P. aeruginosa* multidrug resistance to other antibiotics [14].

*P. aeruginosa*, is a Gram-negative, aerobic bacteria that is also considered as a facultative anaerobe [15–17]. It survives the most diverse environments, spreading quickly in a new habitat, as it has minimal nutritional needs and is able to get energy from a variety of carbon sources [15,17]. In municipal drinking water distribution systems, it proliferates forming biofilm [1]. Apart from the presence of  $bla_{NDM-1}$ , *P. aeruginosa* carries a variety of virulence factors, such as the type III secretion system, which enhances disease severity by injecting effector proteins into host cells [15,17]. In hospital settings, it causes infections in patients in the intensive care units (ICU) or in patients with serious underlying disorders, such as cystic fibrosis (CF) or suppressed immune system, infection of the urinary tract, ear, sinuses, wounds, skin, connective tissues and pulmonary disease [15,18]. Even with the adoption of control measures [18], infections continue to become more complex and difficult to treat and are associated with significant hospital and societal burdens. Moreover, infections eventually lead to high morbidity and mortality rate [15,17,18] as compared to infections caused by members of Enterobacteriaceae or other non-lactose fermenting Gram-negative bacilli [15,17].

Considering that (i) *P. aeruginosa* is intrinsically resistant to many commonly used antibiotics and additionally acquires resistance by horizontal gene transfers or mutation [1]; (ii) *P. aeruginosa* is resistant to the last resort antibiotics (carbapenems) [19,20]; (iii) second-line treatment options for *P. aeruginosa* are often accompanied by toxicity and are less defined in their efficacy [21] and; (iv) *P. aeruginosa* in healthcare settings, environment and water sources prevents control of its acquisition among the most vulnerable patients [15], elimination of an environmental reservoir could be the answer to controlling some of the outbreaks [22]. The use of conventional treatment methods and even other advanced oxidation processes have been evaluated in a previous study, where the merits were observed to outweigh any drawbacks [23]. Hence, for the use of an Advanced Oxidation Process (AOP), recent studies have demonstrated the ability of ozone-electrolyzers, photo-electro and electro-disinfection to achieve logarithmic removal of *Klebsiella pneumoniae* and considerable attenuation of its ARGs from hospital urine and wastewater, respectively [24,25]. In consequence of the above, this study adopted a Non-Thermal Plasma (NTP) treatment technique, which we hypothesized that it would facilitate rapid elimination of ARB and attenuation of its ARG. It was also presumed that it could serve as an alternative disinfection step in WWTPs, as it is able to break down organic matter while causing irreversible damage to cells

and inactivating ARB (in this case *P. aeruginosa*, carbapenem-resistant) and its ARG [26,27].

## 2. Materials and method

### 2.1. Non-thermal plasma reactor

The reactor adopted in this study involved a configuration comprising a designed air-tight machined Polytetrafluoroethylene (PTFE), copper electrodes and high voltage cable, assembled and fitted to a 250 mL capacity Duran Schott glass bottle. The reactor was powered with a high voltage direct current (HVDC) developed by Jeeneel Technology Services Pty Ltd. in South Africa, having a maximum capacity of 40 kV and 15 mA (Fig. 1). A flat copper disk functioned as the ground electrode, with a 50 mm gap between it and the hollow copper electrode. The gap between the hollow copper electrode and the surface of the solution was 15 mm. The voltage on the HVDC was set at 23 kV to ignite the plasma but was regulated to 10 kV with a constant current of 0.7 mA, sustaining a 7 W discharge power throughout the treatment.

### 2.2. Organism and chemicals

The bacterial strain used in this study (ATCC *P. aeruginosa* (27853)) was obtained from Laboratory Specialties PTY LTD Trading as Thermo Fisher Scientific in Randburg, South Africa. Imipenem antibiotic, Luria Bertani (LB) broth and LB agar plates were procured from Sigma Merck, South Africa. Nucleomag DNA/RNA Water kit was purchased from Separations in Randburg, South Africa and primers were delivered by Inqaba in Pretoria, South Africa.

### 2.3. Antibiotic screening and non-thermal plasma treatment

#### 2.3.1. Antibiotic screening

The preliminary screening of *P. aeruginosa* for antibiotic resistance was performed using culture-based methods [28]. The bacteria were inoculated into Luria Bertani (LB) broth and orbitally incubated at 37 °C, 160 rpm for 24 h. Thereafter, the cultures were successively screened with increasing gradients of imipenem (2  $\mu$ g/mL, 4  $\mu$ g/mL, 6  $\mu$ g/mL, 8  $\mu$ g/mL, 12  $\mu$ g/mL and 24  $\mu$ g/mL), which were supplemented in LB agar and incubated at 37 °C for 24 h [29,30]. The growth of *P. aeruginosa* on the plates supplemented with imipenem confirmed their resistance to carbapenems. The isolate that grew on the plates supplemented with 24  $\mu$ g/mL imipenem antibiotic was used as the standard carbapenem

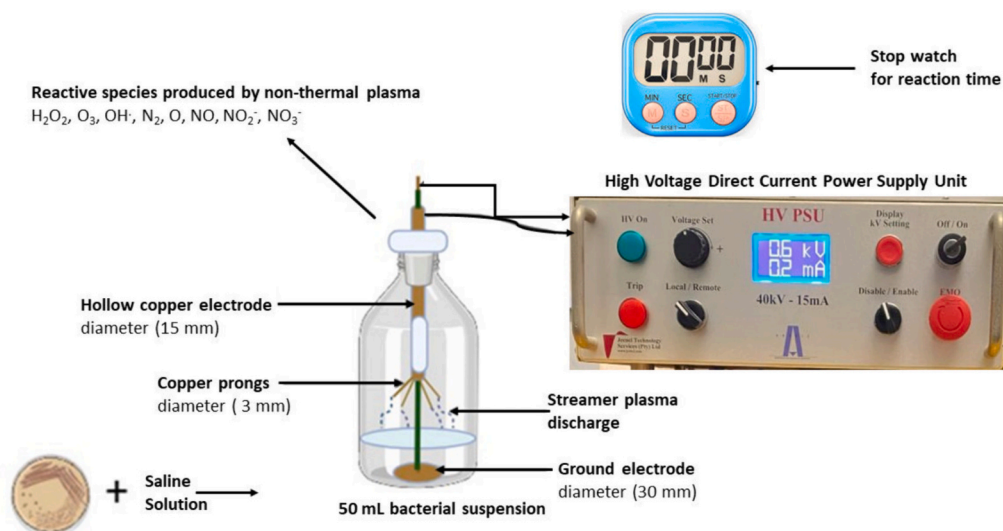


Fig. 1. Schematic diagram of the non-thermal plasma experimental setup, showing all variables.

resistance bacteria strain (CRBS) for the inactivation experiment in this study.

### 2.3.2. Non-thermal plasma treatment

The standard *P. aeruginosa* strain underwent a series of procedures. It was initially introduced into nutrient broth and incubated at 37 °C, 160 rpm for 16 h to minimize biofilms. After incubation, bacterial pellets were collected, washed, and re-suspended in saline. The resulting solution was plated on LB agar with 24 µg/mL imipenem, yielding an average plate count of  $3.0 \times 10^9$  CFU/mL before plasma treatment. Plasma exposure involved treating a 50 mL bacterial suspension at different time intervals (3, 6, 9, 12, and 15 min) in triplicates. Post-plasma treatment samples were plated on LB agar with 24 µg/mL imipenem and incubated at 37 °C over 48 h. Colony-forming units per mL were then estimated for log reduction calculations. Additionally, to explore copper's potential antimicrobial properties, bacterial suspensions were treated with the copper electrode alone for 15 min, referencing studies by [31,32].

### 2.4. Physical chemical and structural characterization

The discharge characteristics and chemical species formed during treatment were assessed using the Black Comet C-25 Spectrometer (StellarNet, Inc.). The  $H_2O_2$  and nitrite/nitrate ions recorded unmeasurable absorbances at <350 nm, hence the Lovibond® SpectroDirect water testing instrument (Tintometer Group, Germany) was used to accurately determine the concentration of ( $H_2O_2$ ), ( $NO_2^-$ ) and ( $NO_3^-$ ) after treating the bacterial suspensions with NTP for 3, 9 and 15 min [33]. The resultant bacterial suspensions (10 mL) were assayed for  $NO_2^-$  and  $H_2O_2$  concentrations measurement using Titanium Tetrachloride reagent and N-(1-Naphthyl)-ethylenediamine, respectively. For  $NO_3^-$  concentrations measurement, 0.5 mL of the resultant bacterial suspensions was assayed using 2,6-Dimethylphenole reagent. The reactions were monitored spectrophotometrically after respective time intervals specified by the user manual [34]. The pH of the bacterial suspension before the plasma treatment and after 3, 6, 9, 12 and 15 min of treatment were measured using PL-700AL pH meter. The conductivity of the bacterial suspension before the plasma treatment and after 3, 6, 9, 12 and 15 min of treatment were determined using a WTW – Portable conductivity meter ProfilLine Cond 3310 (Einzelgerät, Zubehör).

#### 2.4.1. Scanning electron microscopic analysis

Bacterial suspensions were treated with NTP, and both pre- and post-treatment, underwent centrifugation to discard the supernatant and recover pelleted cells. The cells were washed with a phosphate buffer to remove serum and media, followed by fixation using a 2.5 % Glutaraldehyde/Formaldehyde solution for 1 to 24 h. Subsequently, the fixed pellets were subjected to multiple washes with phosphate buffer and post-fixed with a 1 % Osmium Tetroxide (OsO4) solution for 1 h. Dehydration was achieved through a graded ethanol series (30 %, 50 %, 70 %, 90 % and  $3 \times 100$  %), and a mixture of Hexamethyldisilazane (HMDS) and ethanol was applied before drying. The samples were then mounted on aluminium stubs, coated with carbon, and examined using a scanning electron microscopy (SEM) (Zeiss Gemini Ultra Plus FEG-SEM (Field Emission Gun – Scanning Electron Microscope) with BS, energy dispersive spectroscopy (EDS) and Electron Backscatter Diffraction (EBSD) detectors).

**Table 1**  
Primer used in this study.

Name	Forward primer (5' – 3')	Reverse primer (5' – 3')	Size (bp)	Reference
<i>bla</i> <sub>NDM-1</sub>	GGTGCATGCCCGGTGAAATC	ATGCTGGCCTTGGGGAACGS	660	[35]

### 2.5. Molecular analysis

The extraction of DNA was conducted both before and after the plasma treatment of the bacterial suspension, using the NucleoMag® DNA/RNA Water extraction kit, according to the manufacturer's instructions. The extracted DNA served as the template for the PCR assay, aiming to validate the presence of the *bla*<sub>NDM-1</sub> gene in *P. aeruginosa* (Table 1).

The amplification of the *bla*<sub>NDM-1</sub> gene was carried out using a 10 µL PCR mixture, consisting of Taq 2× Master Mix (New England Biolabs), forward and reverse primers (0.5 µM), template DNA (2.5 µL), and Milli-Q water (1.5 µL) in a BIO RAD T100 Thermal cycler. The reaction conditions included initial denaturation (95 °C for 3 min) and 30 cycles for denaturation (95 °C for 1 min), annealing (55 °C for 1 min), and extension (72 °C for 1 min 30 s), and a terminal extension (72 °C for 10 min). The ethidium bromide-stained PCR products were visualized through electrophoresis in a 1.5 % agarose gel [36,37], employing the BIO RAD PowerPac basic with Mini Sub Cell GT. Gel electrophoresis facilitates the efficient separation of DNA fragment and ethidium bromide enables DNA fluorescence and detection under UV light [38]. This method allowed for the observation of gene presence or absence on gel images, providing insights into the effectiveness of plasma treatment in inactivating resistance genes [35].

## 3. Results and discussion

### 3.1. Physical chemical and structural characterization

Reactive nitrogen species (RNS) and reactive oxygen species (ROS) play an important role in the inactivation of bacteria [39,40]. During NTP treatment, the electric field discharges ROS, such hydroxyl (OH) and hydrogen peroxide ( $H_2O_2$ ) [41,42] and RNS, such as nitrites ( $NO_2^-$ ) and nitrate ( $NO_3^-$ ) [41–43] were produced in both the adjoining liquid and gaseous media [42]. The oxygen ( $O_2$ ) and nitrogen ( $N_2$ ) lines (316 nm, 337 nm, 404 nm), independently underwent electron impact ionization reactions, resulting in electrons and positive ions ( $O_2^+$ ,  $N_2^+$ ), that split and eventually increased and spread the streamer (Fig. 2). Then the NO lines (227 nm, 239.5 nm),  $O^+$  lines (327 nm, 435.5 nm, 464.5 nm),  $N^+$  lines (344 nm, 395.5 nm) and O lines (777 nm) were generated in the gas phase [44]. In wet air or in liquids, the •OH lines (307.5 nm, 309 nm),  $H_2O_2$  and ozone ( $O_3$ ) were produced [43,44]. The nitrogen oxides (NO) dissolved in water forming  $NO_2^-$  and  $NO_3^-$  [42]. The •OH caused impairment to DNA, as it has strong oxidation potential (2.8 V) that is greater than the conventional disinfectants, chlorine (1.36 V) and ozone (2.07 V) [45–49]. Among ROS, •OH have the greatest oxidation potential, are the most reactive, and are considered to play a crucial role in NTP bacterial treatment [43,50,51]. Correspondingly, the •OH radical causes irreversible damage to cells and inactivates ARBs and ARGs, as it has diverse impact on normal protein structure which is one of the primary targets in bacteria during disinfection, including oxidation of amino acids, modification of sulphur groups [26,42]. The ( $H_2O_2$ ), ( $NO_2^-$ ) and ( $NO_3^-$ ) had a relatively long lifetime and could react to secondary products post-discharge. The post-discharge reactions between the by-products occurring in plasma activated water (PAW) might result in the generation of peroxynitrous ( $HNO_3$ )/peroxynitrite ( $ONOO^-$ ) acid, which significantly participates in the antibacterial activity of PAW. These long-lived species [52,53] have been observed to inactivate cells even after the discharge had been switched off [54]. The long-term, post-plasma effect of the PAW was mainly caused by the reaction

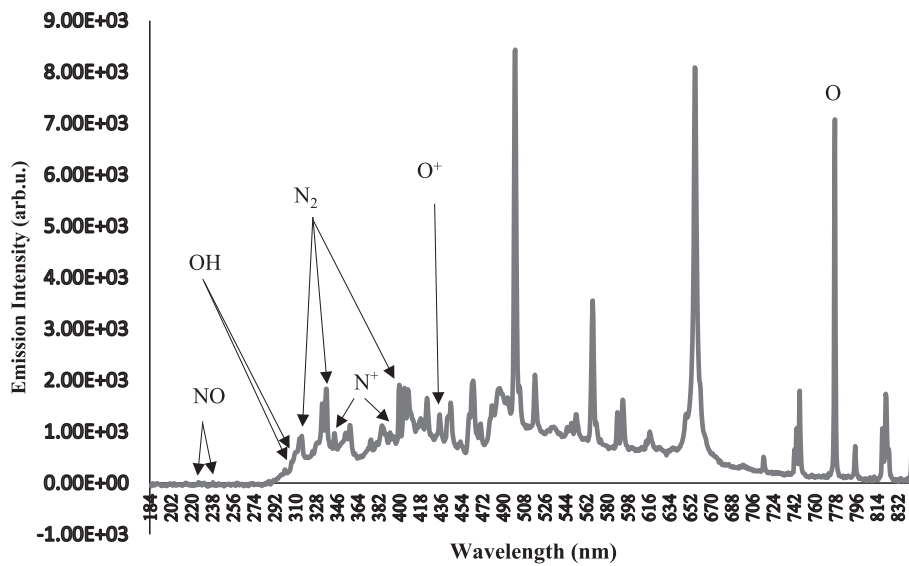


Fig. 2. Optical emission spectra from hydroxyl radical OH species in NTP during treatment of *P. aeruginosa*.

between ozone and  $H_2O_2$  during the peroxone process that forms  $\bullet OH$  [51]. The concentration of these species increased with time during NTP treatment of *P. aeruginosa*, as can be seen in Fig. 2b-d. The highest concentrations of the species were observed when no more re-growth of *P. aeruginosa*.

Nitrates ( $NO_3^-$ ) and nitrites ( $NO_2^-$ ) were incrementally discharged into the solution with time and the highest concentration of  $6.5 \pm 0.325$  mg/L and  $1.9 \pm 0.095$  mg/L of  $NO_3^-$  (Fig. 3a) and  $NO_2^-$  (Fig. 3b), respectively, were observed at 15 min. Another study confirmed an increase in the concentration of  $NO_3^-$  with treatment time, although a much higher concentration of 41.41 mg/L of  $NO_3^-$  and 5.27 mg/L of  $NO_2^-$  was attained after the same treatment time. However, the result was from treating deionised water only with a dielectric barrier discharge configuration [55]. The oxidation of nitrite to nitrate [56] may have

contributed to there being more  $NO_3^-$  than  $NO_2^-$  in this study.

There was no measurable concentration of hydrogen peroxide ( $H_2O_2$ ) observed at 3 min; however,  $0.15 \pm 0.0075$  mg/L and  $0.36 \pm 0.018$  mg/L concentrations were observed at 9 and 15 min (Fig. 3c). A study resulted in the formation of  $H_2O_2$  which was immediately observed after plasma treatment but the concentration decreased with time of incubation [42], another study resulted in nil production of  $H_2O_2$  after plasma treatment [55]. Plasma treatment usually results in quadratic or linear increase of  $H_2O_2$  with treatment time but the cells in the medium uptake  $H_2O_2$  with incubation time [42,55].

The pH was reduced to  $3.85 \pm 0.1925$  from  $7.12 \pm 0.356$  after 15 min of plasma treatment (Fig. 3d). This outcome is similar to other studies, which also treated water for 15 min with plasma, both achieving pH of 3.85 and 3.78, respectively [55]. The reduction in pH was caused

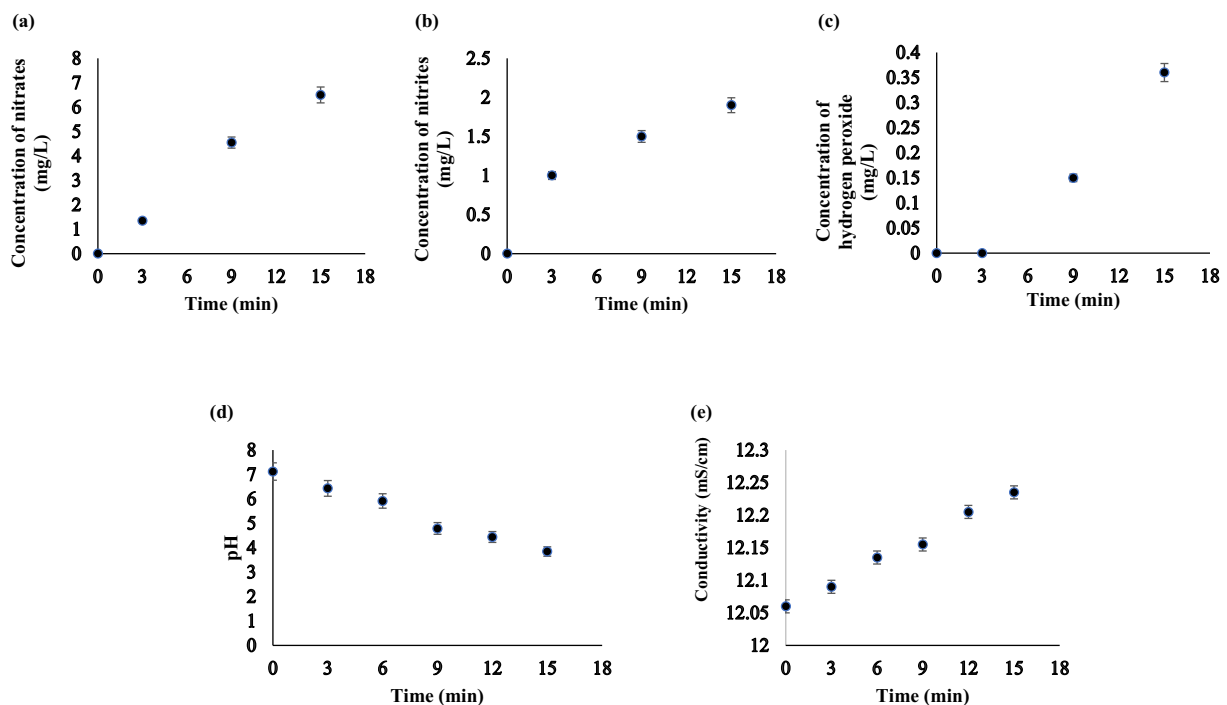


Fig. 3. Concentration of (a) nitrates, (b) nitrites and (c) hydrogen peroxide during non-thermal plasma treatment of *P. aeruginosa*. (d) pH and (e) Conductivity readings of *P. aeruginosa* bacterial suspension during non-thermal plasma treatment. NB: data are mean of duplicate readings.



by the presence of nitrates and nitrites in the suspension, which led to the generation of  $\text{HNO}_3$ . The low pH keeps the oxidizing potential of ozone at 2.08 V, which can decrease to 1.4 V under alkaline conditions [56,57]. The production of hydrogen radicals also increased under acidic conditions, which then reacted with  $\text{H}_2\text{O}_2$  and  $\text{H}_2\text{O}$  to produce more  $\bullet\text{OH}$ . An acid pH range of 3–4 is said to be conducive for production of  $\bullet\text{OH}$  [51] and the reactive molecules readily penetrate cell walls under the low pH conditions because the cell membrane permeability increases [43].

The conductivity was  $12.06 \pm 0.603$  mS/cm before treatment; it increased to  $12.24 \pm 0.612$  mS/cm after 15 min of plasma treatment (Fig. 3e). The conductivity of water tends to vary as a result of the ROS and RNS produced during NTP treatment. This is evident as one study resulted in an increase of conductivity from  $1 \mu\text{S}/\text{cm}$  to  $123 \mu\text{S}/\text{cm}$  over 15 min treatment time [55] while in another study the conductivity fluctuated between 2.57 mS/cm to 3.31 mS/cm over 30 min NTP treatment [58]. The conductivity in this study increased with treatment time, but the increment was low, perhaps because a low conductivity is the one that favours the production of  $\text{H}_2\text{O}_2$  and  $\text{O}_3$ , which contribute to the destruction of the pollutants [57,59]. The increase in conductivity is an indication of a loss in cell membrane integrity of bacteria [60].

### 3.2. Inactivation of ARB and ARG

#### 3.2.1. Re-growth assessment

The observed trend indicated an escalating log reduction of *Pseudomonas aeruginosa*, implying a logarithmic inactivation of ARB with extended treatment time ( $9.48 \pm 0.474$  after 12 and 15 min) (Fig. 4). The close to 100 % reduction attained in this investigation indicates that NTP can be used as an alternative disinfection step (Table 2).

Isolation of copper on its own yielded the smallest log reduction (0.15) of *P. aeruginosa*, potentially due to the solid state of the copper employed in this study. Soluble copper is acknowledged for its high toxicity and reactivity, owing to oxidative power, catalyzed ROS formation (i.e.,  $\text{H}_2\text{O}_2$ ), leading to lipid peroxidation and DNA/RNA damage, ultimately resulting in bacterial inactivation [61,62]. A comparative study of copper compounds (copper oxide, copper acetate, copper nitrate, and copper sulfate), all possessing bactericidal properties, revealed that copper in salt form exhibited the highest antimicrobial effectiveness [31]. Moreover, considering the low capacity of  $\text{Cu}^{2+}$  to stimulate bacterial membrane and nucleic acid destruction through copper-catalyzed Fenton-like reactions (driven by the reactions of  $\text{Cu}^{2+}/\text{H}_2\text{O}_2$ ) [63], it was therefore reasonable to deduce that the copper electrode had negligible impact on the reduction of *P. aeruginosa*.

Consequently, the observed log reductions were attributed solely to the plasma discharge. The challenging characteristics of *P. aeruginosa* as a Gram-negative bacterium that thrives in biofilms likely contributed to the limited efficacy of solid copper in its inactivation [61].

#### 3.2.2. Evidence of cellular disintegration

The rod-like shape disseminated by SEM imaging confirmed the structural integrity of *P. aeruginosa* cells before NTP treatment (Fig. 5) [64].

Following NTP treatment, SEM images indicated a loss of structural integrity on the surface of *P. aeruginosa* cells (Fig. 6). The cellular remnants of *P. aeruginosa* assumed an amorphous shape, deviating from their usual rod-like structure observed before NTP treatment. This observation highlighted NTP's capability to overcome the tensile strength of the cell membrane, leading to rupture, leakage of intracellular components, and cell death. These effects effectively impeded the growth of *P. aeruginosa* and negated its protection of ARGs [43]. The rupturing phenomenon was attributed to the accumulation of ROS and RNS free radicals on the cell membrane, surpassing the membrane's tensile strength and causing it to rupture [43,65]. Notably, NTP appears to have a greater inhibitory effect on gram-negative bacteria than gram-positive bacteria, as demonstrated by its ability to disrupt the membrane of *P. aeruginosa*, a gram-negative bacterium [43,66]. A study compared SEM images of Gram-negative and Gram-positive cell after plasma treatment, the destruction was more evident on Gram-negative as there was cell breakage effects on the Gram-negative and only irregular shape and cell shrinkage of the Gram-positive cells [67]. This susceptibility was attributed to the thinner peptidoglycan layer and the presence of an outer membrane with components such as proteins and lipopolysaccharide (LPS), which are sensitive to ROS in Gram-negative bacteria [43]. A hypothesized dual mechanism for the destruction of cellular membranes involves lipid peroxidation and electroporation. In lipid peroxidation, reactive radicals generated by plasma, particularly OH groups, detach polar head moieties and fatty acid tails from phosphatidylcholine residues in the lipid bilayer of cell plasma membranes. This leads to crosslinks between adjacent fatty acid tails, allowing the unrestricted influx of radicals and water molecules, resulting in membrane lesions and pore formation. Alternatively, electric fields generated by plasma could enhance transmembrane potential, initiating the breakdown of the lipid bilayer and the formation of membrane pores. The recoverability of these pores may vary, and with increased electric field strength and duration, they may become irrecoverable, ultimately causing necrosis and cell rupture [42].

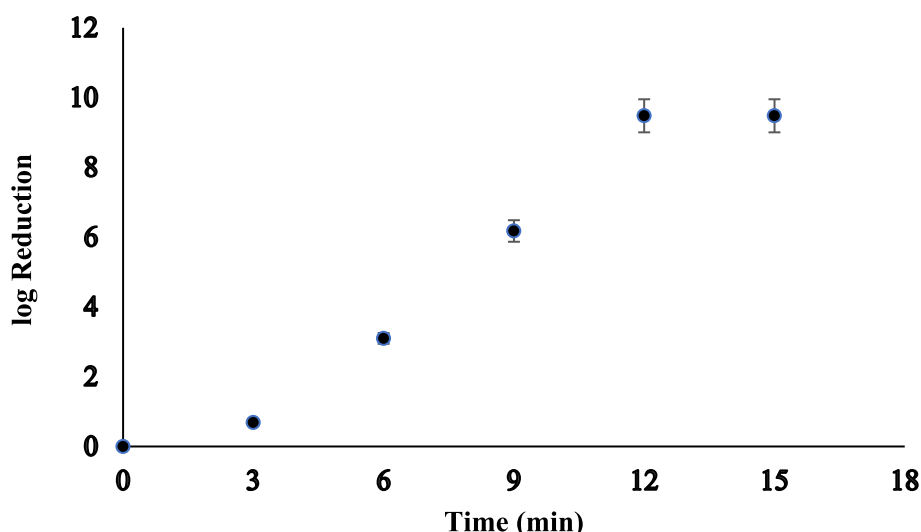
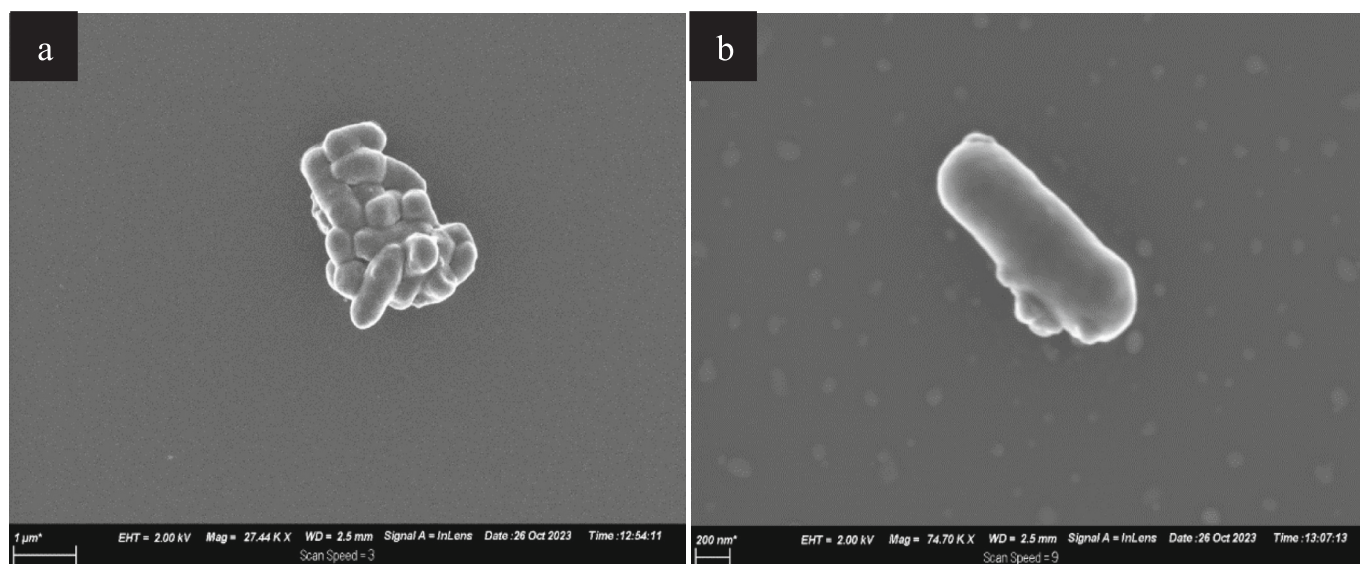


Fig. 4. Log reduction of *P. aeruginosa* (27853) after non-thermal plasma treatment. Where log reduction =  $\log_{10}$  (initial CFU/final CFU).

**Table 2**  
Percentage reductions of *P. aeruginosa* (BAA 1605) after non-thermal plasma treatment.

	Control (0 min)	Copper (15 min)	3 min	6 min	9 min	12 min	15 min
CFU/mL	$3.0 \times 10^9$	$2.1 \times 10^9$	$6.2 \times 10^8$	$2.4 \times 10^6$	$3.65 \times 10^5$	NG	NG
log reduction	0	0.15	0.68	3.1	6.18	$\infty$	$\infty$
% reduction	0	30	79.3	99.92	$\leq 100$	100	100

NB:  $\leq 100$  implies that theoretical reduction was 100 % however, the presence few of culturable colonies practically negates occurrence of 100 % reduction. NG implies absence of *P. aeruginosa* growth, whereas  $\infty$  signifies the infinite log reduction of culturable planktonic cells.



**Fig. 5.** SEM image of *P. aeruginosa* (a) cell cluster and (b) single cell before non-thermal plasma treatment.

### 3.2.3. Detection of resistance gene ( $bla_{NDM-1}$ ) attenuation

In this study, the diminishing intensity of *P. aeruginosa*  $bla_{NDM-1}$  (240 bp) bands over time, as depicted in Fig. 7, suggested the capacity of NTP to inactivate ARGs. Fluorescence of the bands is presumed to indicate the DNA concentration [68]. However, at the final treatment duration (15 min), the ARGs were not entirely inactivated, possibly due to the protective biofilm surrounding *P. aeruginosa*. Biofilm-inhabiting bacteria are known to exchange more genes than those in a planktonic/free-living state [69]. The NDM-1 gene has been reported to exhibit increased expression in optimal biofilm formation conditions [70]. Biofilms, with their high cell densities, facilitate intercellular gene transfer due to the short physical distance between microorganisms [69]. The protective nature of the biofilm also hinders NTP reactive species from penetrating it [71]. Despite this, a study utilizing plasma-generated Fenton-oriented reactions ( $Cu^{2+}/H_2O_2$  and  $Fe^{2+}/H_2O_2$ ) recorded measurable gene copies of  $bla_{TEM-1}$  after 10 min, despite enhanced ARG inactivation by  $Cu^{2+}$  and  $Fe^{2+}$  [63]. This underscores the efficiency of our treatment process, which does not rely on exogenous Fenton-oriented inorganic ions ( $Cu^{2+}$  and  $Fe^{2+}$ ). Comparative studies show UV achieving a maximum reduction of 99 % [72], chlorination achieving 100 % reduction [73], and ozone achieving a maximum of 98.1 % [74] of ARGs. However, the concentrations of these disinfectants were considerably higher and impractical compared to those used in WWTPs, raising environmental concerns due to the formation of harmful by-products like halo-organics [75,76] and bromate [74,77] from chlorine and ozone, respectively. Furthermore, we are of the opinion that extending NTP treatment time might lead to complete inactivation of ARGs, offering the advantage of eliminating the need for additional use of potentially harmful chemicals. Additional parameters, such as increasing the grounded electrode surface or diameter at a fixed discharge gap, could be optimized to enhance the area of plasma discharge. The choice of feeding gas also plays a role, with oxygen-

containing gas being demonstrated to lead to the fastest contaminant degradation, followed by air and argon, while nitrogen results in the least removal. Oxygen-containing gas induces the production of  $O^-$  based active species and  $O_3$ , enhancing the degradation rate of pollutants [51,74].

## 4. Conclusions

An experiential extermination of *P. aeruginosa* was achieved over 12 min treatment, which correspondingly suggested the diminishing presence of the resistance gene at terminal treatment time (15 min). The observed log reductions further validated the efficacy of NTP in completely eliminating ARBs over time. This effectiveness was attributed to the increasing presence of  $\bullet OH$  and the long-lived species ( $H_2O_2$ ,  $NO_2^-$  and  $NO_3^-$ ) which can generate more  $\bullet OH$  and reacting to secondary products post-discharge, significantly contributing to the antibacterial activity of PAW. These species also played a role in lowering the pH to levels conducive to bacterial destruction, while the conductivity remained consistently low.

The SEM results provided additional confirmation that NTP disrupts the tensile strength of the cell membrane, leading to rupture and ultimately cell death. This mechanism, in turn, prevents bacterial growth, which might in some cases include the disintegration of DNA at the ARG loci. However, it was observed that the electrophoretic gel visualization of  $bla_{NDM-1}$  bands might not be sufficient for accurate monitoring of log reduction in gene copies per treatment time. In this regard, future studies will be devoted to evaluating gene copies as well as abundance of nucleotide bases per treatment time. This will not only substantiate the  $bla_{NDM-1}$  disintegration claims in this study, but also provide insight on the interaction of plasma treatment with the chemical bonds of the DNA. The study portrayed the inconsequential antimicrobial contribution of copper, further emphasizing that the observed log reductions were

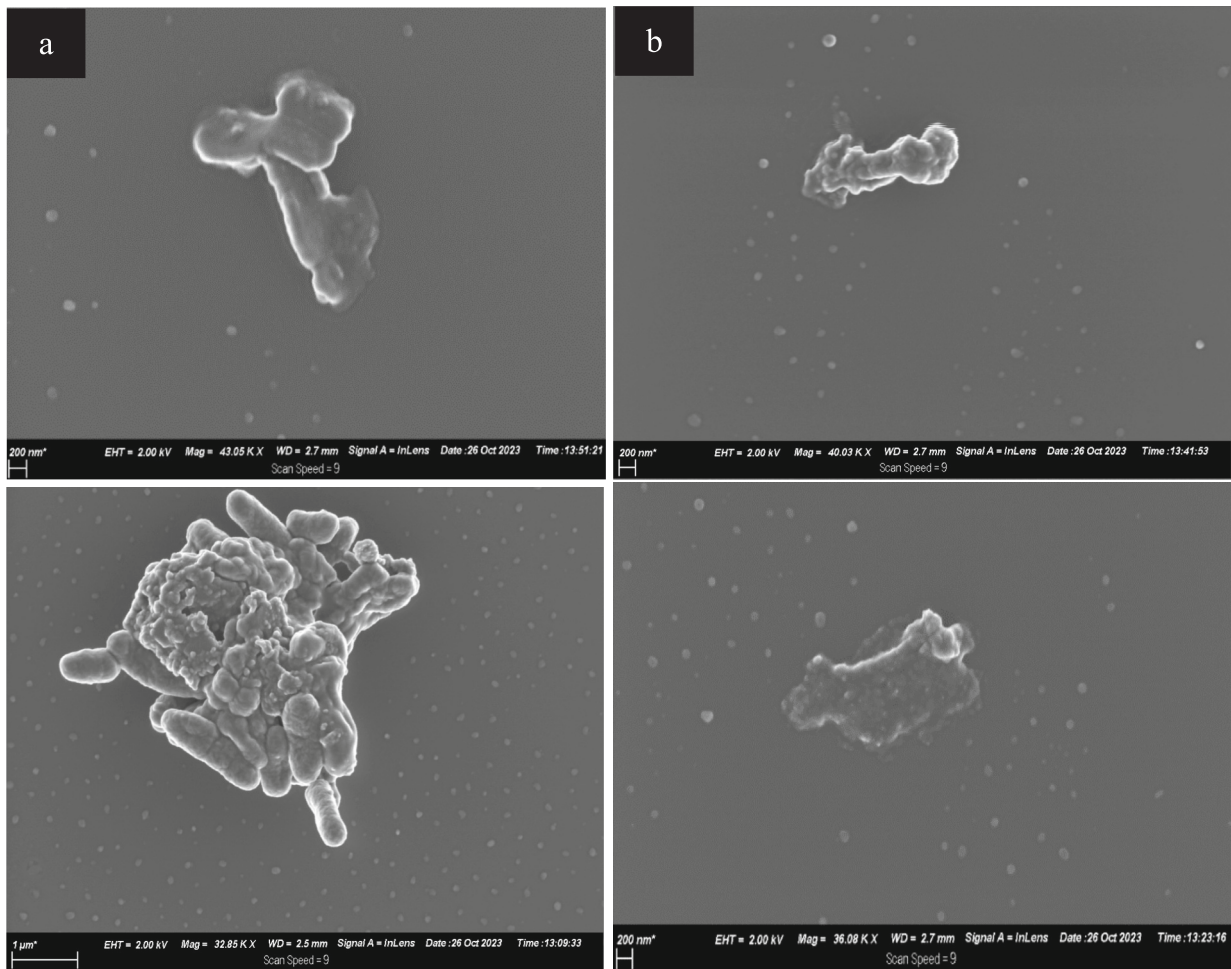


Fig. 6. SEM image of *P. aeruginosa* after 15 min of non-thermal plasma treatment. (a) single cell rupture and cytosol leakage (b – c) rupture of cell cluster (d) total disintegration of single cell.

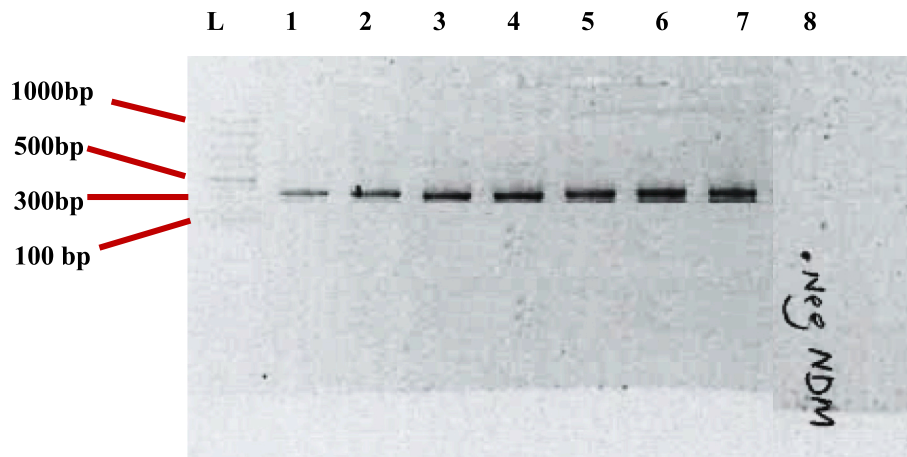


Fig. 7. Effect of periodic non-thermal plasma treatment on the carbapenem resistant gene of *P. aeruginosa*. L: DNA ladder, 1: 15 min, 2: 12 min, 3: 9 min, 4: 6 min, 5: 3 min, 6: copper control, 7: +ve control, 8: -ve control.

solely attributed to NTP.

Overall, these findings underscore the potential of plasma treatment as an effective disinfection step for wastewater. In correspondence, the potential of NTP at the tertiary stage of wastewater treatment has been evaluated for environmental and economic footprint, especially when retrofitted to other technologies at the tertiary treatment stage in a

WWTP [78]. Notwithstanding, Mosaka et al. [23] highlighted the challenges that still needs to be overcome in its scale up or adoption at real-life, large-scale scenarios. In view of the aforementioned challenges, and the observations reported in this study, it is recommended that addressing biofilms near *P. aeruginosa*, extending the reaction time, using oxygen-containing feed gas and enlarging the diameter of the



grounded electrode may further enhance the complete inactivation of ARGs.

### CRediT authorship contribution statement

**Thabang B.M. Mosaka:** Writing – review & editing, Writing – original draft, Validation, Methodology, Investigation, Formal analysis, Data curation. **John O. Unuofin:** Writing – review & editing, Writing – original draft, Visualization, Validation, Supervision, Methodology, Investigation, Formal analysis, Data curation, Conceptualization. **Michael O. Daramola:** Writing – review & editing, Validation, Supervision, Resources, Project administration, Investigation, Formal analysis. **Chedly Tizaoui:** Writing – review & editing, Validation, Supervision, Project administration, Investigation, Formal analysis. **Samuel A. Iwarere:** Writing – review & editing, Validation, Supervision, Resources, Project administration, Methodology, Investigation, Funding acquisition, Formal analysis, Conceptualization.

### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

### Data availability

Data will be made available on request.

### Acknowledgements

SAI acknowledges the support by the Government of the United Kingdom through The Royal Society FLAIR award [FLR\R1\201683]. JOU acknowledges the financial support of the National Research Foundation (Grant No: 138445).

### Ethical approval

None applicable.

### References

- [1] A. Osińska, E. Korzeniewska, M. Harnisz, E. Felis, S. Bajkacz, P. Jachimowicz, S. Niestępski, I. Konopka, Small-scale wastewater treatment plants as a source of the dissemination of antibiotic resistance genes in the aquatic environment, *J. Hazard. Mater.* 381 (2020) 121221.
- [2] F.T. Fadare, A.I. Okoh, The abundance of genes encoding ESBL, pAmpC and Non-β-Lactam resistance in multidrug-resistant Enterobacteriaceae recovered from wastewater effluents, *Front. Environ. Sci.* 2021, p. 9.
- [3] Y. Ben, C. Fu, M. Hu, L. Liu, M.H. Wong, C. Zheng, Human health risk assessment of antibiotic resistance associated with antibiotic residues in the environment: A review, *Environ. Res.* 169 (2019) 483–493.
- [4] J.O. Unuofin, Garbage in garbage out: the contribution of our industrial advancement to wastewater degeneration, *Environ. Sci. Pollut. Res.* 27 (18) (2020) 22319–22335.
- [5] D. Rodríguez-Molina, P. Mang, H. Schmitt, M.C. Chifiriuc, K. Radon, L. Wengenroth, Do wastewater treatment plants increase antibiotic resistant bacteria or genes in the environment? Protocol for a systematic review, *Systemat. Rev.* 8 (2019) 304.
- [6] L. Chen, Z. Zhou, C. Shen, Y. Xu, Inactivation of antibiotic-resistant bacteria and antibiotic resistance genes by electrochemical oxidation/electro-Fenton process, *Water Sci. Technol.* 81 (2020) 2221–2231.
- [7] M. Jin, L. Liu, D.N. Wang, D. Yang, W.L. Liu, J. Yin, Z.W. Yang, H.R. Wang, Z. G. Qiu, Z.Q. Shen, D.Y. Shi, Chlorine disinfection promotes the exchange of antibiotic resistance genes across bacterial genera by natural transformation, *ISME J.* 14 (7) (2020) 1847–1856.
- [8] M. Jin, L. Liu, D.N. Wang, D. Yang, W.L. Liu, J. Yin, Z.W. Yang, H.R. Wang, Z. G. Qiu, Z.Q. Shen, D.Y. Shi, H.B. Li, J.H. Guo, J.W. Li, Chlorine disinfection promotes the exchange of antibiotic resistance genes across bacterial genera by natural transformation, *ISME J.* 14 (2020) 1847–1856.
- [9] C. Sarangapani, D. Zuzina, P. Behan, D. Boehm, B. Gilmore, P.J. Cullen, P. Bourke, Degradation kinetics of cold plasma-treated antibiotics and their antimicrobial activity, *Sci. Rep.* 9 (2019) 3955.
- [10] Y.H. Wu, Y.H. Wang, S. Xue, Z. Chen, L.W. Luo, Y. Bai, X. Tong, H.Y. Hu, Increased risks of antibiotic resistant genes (ARGs) induced by chlorine disinfection in the reverse osmosis system for potable reuse of reclaimed water, *Sci. Total Environ.* 815 (2022) 152860.
- [11] N. Hassoun-Kheir, Y. Stabholz, J.-U. Kreft, R. De La Cruz, J.L. Romalde, J. Nesme, S.J. Sorensen, B.F. Smets, D. Graham, M. Paul, Comparison of antibiotic-resistant bacteria and antibiotic resistance genes abundance in hospital and community wastewater: A systematic review, *Sci. Total Environ.* 743 (2020) 140804.
- [12] K. Soni, K. Jyoti, H. Chandra, R. Chandra, Bacterial antibiotic resistance in municipal wastewater treatment plant; mechanism and its impacts on human health and economy, *Bioresour. Technol. Rep.* 19 (2022) 101080.
- [13] World Health Organisation, WHO Bacterial Priority Pathogens List, 2024: Bacterial Pathogens of Public Health Importance to Guide Research, Development and Strategies to Prevent and Control Antimicrobial Resistance, World Health Organisation, Geneva, 2024, p. 2024. <https://iris.who.int/bitstream/handle/10665/376776/9789240093461-eng.pdf?sequence=1>. (Accessed July 2024).
- [14] Y. Ding, J.W. Teo, D.I. Drautz-Moses, S.C. Schuster, M. Givskov, L. Yang, Acquisition of resistance to carbapenem and macrolide-mediated quorum sensing inhibition by *Pseudomonas aeruginosa* via ICETn 4371 6385, *Nat. Comm. Biol.* 1 (1) (2018) 57.
- [15] W. Huang, X. Wei, G. Xu, X. Zhang, X. Wang, Carbapenem-resistant *Pseudomonas aeruginosa* infections in critically ill children: prevalence, risk factors, and impact on outcome in a large tertiary pediatric hospital of China, *Front. Public Health* 11 (2023) 1088262.
- [16] Mcauley, D. 2018. *Pseudomonas aeruginosa* [Online]. Global RPH. Available: <https://globalrph.com/bacteria/pseudomonas-aeruginosa/> [Accessed 27 October 2023].
- [17] M.S. Walters, J.E. Grass, S.N. Bulens, E.B. Hancock, E.C. Phipps, D. Muleta, J. Mounsey, M.A. Kainer, C. Concannon, G. Dumyati, C. Bower, J. Jacob, P. M. Cassidy, Z. Beldavs, K. Culbreath, W.E. Phillips Jr., D.J. Hardy, R.L. Vargas, M. Oethinger, U. Ansari, R. Stanton, V. Albrecht, A.L. Halpin, M. Karlsson, J. K. Rasheed, A. Kallen, Carbapenem-resistant *Pseudomonas aeruginosa* at US emerging infections program sites, 2015, *Emerg. Infect. Dis.* 25 (2019) 1281–1288.
- [18] F. Yuan, M. Li, X. Wang, Y. Fu, Risk factors and mortality of carbapenem-resistant *Pseudomonas aeruginosa* bloodstream infection in haematology department: A 10-year retrospective study, *J. Global Antimicrob. Resist.* 37 (2024) 150–156.
- [19] C.M. Gill, D. Santini, D.P. Nicolau, In vitro activity of cefiderocol against a global collection of carbapenem-resistant *Pseudomonas aeruginosa* with a high level of carbapenemase diversity, *J. Antimicrob. Chemother.* 79 (2) (2024) 412–416.
- [20] S.N. Shahab, A. van Veen, A.C. Büchler, Y.R. Saharman, A. Karuniawati, M.C. Vos, Voor in't holt, A.F. and Severin, J.A., In search of the best method to detect carriage of carbapenem-resistant *Pseudomonas aeruginosa* in humans: a systematic review, *Ann. Clin. Microbiol. Antimicrob.* 23 (1) (2024) 50.
- [21] Y.J. Kim, H.J. Huh, H. Sung, Challenges of carbapenem-resistant *Pseudomonas aeruginosa* in infection control and antibiotic management, *Annals Lab. Medicine* 44 (1) (2024) 1–2.
- [22] R. Valencia-Martín, V. Gonzalez-Galan, R. Alvarez-Marín, A.M. Cazalla-Foncueva, T. Aldabó, M.V. Gil-Navarro, I. Alonso-Araujo, C. Martín, R. Gordon, E.J. García-Núñez, R. Perez, G. Peñalva, J. Aznar, M. Conde, J.M. Cisneros, et al., A multimodal intervention program to control a long-term *Acinetobacter baumannii* endemic in a tertiary care hospital, *Antimicrob. Resist. Infect. Control* 8 (2019) 199.
- [23] T.B.M. Mosaka, J.O. Unuofin, M.O. Daramola, C. Tizaoui, S.A. Iwarere, Inactivation of antibiotic-resistant bacteria and antibiotic-resistance genes in wastewater streams: current challenges and future perspectives, *Front. Microbiol.* 13 (2022) 1100102.
- [24] M. Herraiz-Carboné, S. Cotillas, E. Lacasa, P. Canizares, M.A. Rodrigo, C. Saez, Depletion of ARGs in antibiotic-resistance *Klebsiella*, *Pseudomonas* and *Staphylococcus* in hospital urines by electro and photo-electro disinfection, *J. Water Process Eng.* 49 (2022) 103035.
- [25] S.E. Correia, V. Pertegal, M. Herraiz-Carboné, E. Lacasa, P. Cañizares, M. A. Rodrigo, C. Sáez, Inactivation of waterborne *Klebsiella pneumoniae* with ozone to diminish the risk of hospital effluents using an absorption-based process, *J. Water Process Eng.* 57 (2024) 104732.
- [26] Chen, Y., Duan, X., Zhou, X., Rupeng, W., Wang, S., Ren, N.-Q., Ho, S.-H. 2020c. Advanced oxidation processes for water disinfection: features, mechanisms and prospects. *Chem. Eng. J.* 409, 128207.
- [27] M. Umar, From conventional disinfection to antibiotic resistance control-status of the use of chlorine and UV irradiation during wastewater treatment, *Int. J. Environ. Res. Public Health* 19 (2022).
- [28] Z. Rashmei, H. Bornasi, M. Ghoranneviss, Evaluation of treatment and disinfection of water using cold atmospheric plasma, *J. Water Health* 14 (2016) 609–616.
- [29] K.E. Ebomah, A.I. Okoh, Detection of Carbapenem-resistance genes in *Klebsiella* species recovered from selected environmental niches in the eastern Cape Province, South Africa. *Antibiotics* 9 (2020) 425.
- [30] R.A. Reinke, J. Quach-Cu, N. Allison, B. Lynch, C. Crisostomo, M. Padilla, A method to quantify viable carbapenem resistant gram-negative bacteria in treated and untreated wastewater, *J. Microbiol. Methods* 179 (2020) 106070.
- [31] L. Benhalima, S. Amri, M. Bensouilah, R. Ouzrout, Antibacterial effect of copper sulfate against multi-drug resistant nosocomial pathogens isolated from clinical samples, *Pak. J. Med. Sci.* 35 (2019) 1322–1328.
- [32] C. Ortega-Nieto, N. Losada-García, B. Pessela, P. Domingo-Calap, J.M. Palomo, Design and synthesis of copper Nanobiomaterials with antimicrobial properties, *ACS Bio & Med Chem Au* 3 (2023).
- [33] T. Zhang, R. Zhou, P. Wang, A. Mai-Prochnow, R. Mcconchie, W. Li, R. Zhou, E. W. Thompson, K. Ostrikov, P.J. Cullen, Degradation of cefixime antibiotic in water



- by atmospheric plasma bubbles: performance, degradation pathways and toxicity evaluation, *Chem. Eng. J.* 421 (2021) 127730.
- [34] T. GmbH, *SpectroDirect / PC Spectro II 8c 02/2021* [Online], Tintometer GmbH, Germany, 2021. Available: [https://www.lovibond.com/ix\\_pim\\_assets/Wasseranal\\_ytik/Instruction\\_Manuals/Photometer/SpectroDirect/ins\\_spectrodirect\\_gb\\_lovi.pdf](https://www.lovibond.com/ix_pim_assets/Wasseranal_ytik/Instruction_Manuals/Photometer/SpectroDirect/ins_spectrodirect_gb_lovi.pdf) (Accessed 15 January 2024).
- [35] R. Anand, K. Ellappan, H. Narasimha, Prevalence and characterization of NDM-1 and OXA-48 carbapenemase gene harboring Enterobacteriaceae in a tertiary care hospital, South India. *African J. Bacteriol. Res.* 7 (2015) 60–63.
- [36] T. Awoke, B. Tekla, A. Aseffa, S. Sebre, A. Seman, B. Yeshitela, T. Abebe, A. Mihret, Detection of Bla KPC and Bla NDM carbapenemase genes among *Klebsiella pneumoniae* isolates in Addis Ababa, Ethiopia: dominance of Bla NDM, *PLoS One* 17 (4) (2022) e0267657.
- [37] A. Thapa, K.M. Upreti, N.K. Bimali, B. Shrestha, A.K. Sah, K. Nepal, B. Dhungel, S. Adhikari, N. Adhikari, B. Lekhak, K.R. Rijal, Detection of NDM variants (*bla*<sub>NDM-1</sub>, *bla*<sub>NDM-2</sub>, *bla*<sub>NDM-3</sub>) from carbapenem-resistant *Escherichia coli* and *Klebsiella pneumoniae*: first report from Nepal. *Infect. Drug Resistance* (2022) 4419–4434.
- [38] P. Wittmeier, S. Hummel, Agarose gel electrophoresis to assess PCR product yield: comparison with spectrophotometry, fluorometry and qPCR, *Biotechniques* 72 (4) (2022) 155–158.
- [39] S. Das, V.P. Gajula, S. Mohapatra, G. Singh, S. Kar, Role of cold atmospheric plasma in microbial inactivation and the factors affecting its efficacy, *Health Sci. Rev.* 4 (2022) 100037.
- [40] M. Domonkos, P. Ticha, J. Trejbal, P. Demo, Applications of cold atmospheric pressure plasma Technology in Medicine, Agriculture and Food Industry. *Appl. Sci.* 11 (2021) 4809.
- [41] R.C. Sanito, S.-J. You, Y.-F. Wang, Degradation of contaminants in plasma technology: an overview, *J. Hazard. Mater.* 424 (2022) 127390.
- [42] P.R. Sreedevi, K. Suresh, Cold atmospheric plasma mediated cell membrane permeation and gene delivery-empirical interventions and pertinence, *Adv. Colloid Interf. Sci.* 320 (2023) 102989.
- [43] H. Zhang, C. Zhang, Q. Han, Mechanisms of bacterial inhibition and tolerance around cold atmospheric plasma, *Appl. Microbiol. Biotechnol.* 107 (2023) 5301–5316.
- [44] S. Nijdam, J. Teunissen, U. Ebert, The physics of streamer discharge phenomena, *Plasma Sources Sci. Technol.* 29 (10) (2020) 103001.
- [45] T. Azuma, M. Usui, T. Hayashi, Inactivation of antibiotic-resistant Bacteria in wastewater by ozone-based advanced water treatment processes, *Antibiotics* 11 (2022) 210.
- [46] J. Foster, Plasma-based water purification: challenges and prospects for the future, *Phys. Plasmas* 24 (2017) 055501.
- [47] A. Mai-Prochnow, R. Zhou, T. Zhang, K. Ostrikov, S. Mugunthan, S.A. Rice, P. J. Cullen, Interactions of plasma-activated water with biofilms: inactivation, dispersal effects and mechanisms of action, *npj Biofilms Microbiomes* 7 (2021) 11.
- [48] C.V. Rekhate, J.K. Srivastava, Recent advances in ozone-based advanced oxidation processes for treatment of wastewater- A review, *Chem. Eng. J. Adv.* 3 (2020) 100031.
- [49] V.K. Sharma, X. Yu, T.J. McDonald, C. Jinadatha, D.D. Dionysiou, M. Feng, Elimination of antibiotic resistance genes and control of horizontal transfer risk by UV-based treatment of drinking water: A mini review, *Front. Environ. Sci. Eng.* 13 (2019) 1–9.
- [50] J. Beber De Souza, F. Queiroz Valdez, R.F. Jeranoski, C.M.D.S. Vidal, G.S. Cavallini, Water and Wastewater Disinfection with Peracetic Acid and UV Radiation and Using Advanced Oxidative Process PAA/UV, *Int. J. Photoenerg.* 2015, p. 860845.
- [51] M. Magureanu, F. Bilea, C. Bradu, D. Hong, A review on non-thermal plasma treatment of water contaminated with antibiotics, *J. Hazard. Mater.* 417 (2021) 125481.
- [52] F. Rezaei, P. Vanraes, A. Nikiforov, R. Morent, N. De Geyter, Applications of plasma-liquid systems: A review, *Materials* 12 (2019) 2751.
- [53] E. Tsoukou, P. Bourke, D. Boehm, Temperature stability and effectiveness of plasma-activated liquids over an 18 months period, *Water* 12 (2020).
- [54] M. Naïtali, G. Kamgang-Youbi, J.M. Herry, M.N. Bellon-Fontaine, J.L. Brisset, Combined effects of long-living chemical species during microbial inactivation using atmospheric plasma-treated water, *Appl. Environ. Microbiol.* 76 (2010) 7662–7664.
- [55] Pandey, S., Jangra, R., Ahlawat, K., Mishra, R., Mishra, A., Jangra, S., Prakash, R. 2023. Selective generation of nitrate and nitrite in plasma activated water and its physicochemical parameters analysis. *Phy. Letts. A*, 474, 128832.
- [56] R. Picetti, M. Deeney, S. Pastorino, M.R. Miller, A. Shah, D.A. Leon, A.D. Dangour, R. Green, Nitrate and nitrite contamination in drinking water and cancer risk: A systematic review with meta-analysis, *Environ. Res.* 210 (2022) 112988.
- [57] H. Zeghioud, P. Nguyen-Tri, L. Khezami, A. Amrane, A.A. Assadi, Review on discharge plasma for water treatment: mechanism, reactor geometries, active species and combined processes, *J. Water Process Eng.* 38 (2020) 101664.
- [58] K.J. Liew, X. Zhang, X. Cai, D. Ren, J. Chen, Z. Chang, K. Chong, Tan Chun Yun, M., Chong, C.S., The biological responses of *Staphylococcus aureus* to cold plasma treatment, *Processes* 11 (4) (2023) 1188.
- [59] B. Jiang, J. Zheng, S. Qiu, M. Wu, Q. Zhang, Z. Yan, Q. Xue, Review on electrical discharge plasma technology for wastewater remediation, *Chem. Eng. J.* 236 (2014) 348–368, <https://doi.org/10.1016/j.cej.2013.09.090>.
- [60] Y. Wang, Y. Han, L. Li, J. Liu, X. Yan, Distribution, sources, and potential risks of antibiotic resistance genes in wastewater treatment plant: A review, *Environ. Pollut.* 310 (2022) 119870.
- [61] I. Salah, I. Parkin, E. Allan, Copper as an antimicrobial agent: recent advances, *RSC Adv.* 11 (2021) 18179–18186.
- [62] M. Virieux-Petit, F. Hammer-Dedet, F. Aujoulat, E. Jumas-Bilak, S. Romano-Bertrand, From copper tolerance to resistance in *Pseudomonas aeruginosa* towards Patho-adaptation and hospital success, *Genes* 13 (2022) 301.
- [63] H. Li, R. Song, Y. Wang, R. Zhong, Y. Zhang, J. Zhou, T. Wang, H. Jia, L. Zhu, Inhibited conjugative transfer of antibiotic resistance genes in antibiotic resistant bacteria by surface plasma, *Water Res.* 204 (2021) 117630.
- [64] S.J. Wood, T.M. Kuzel, S.H. Shafikhani, *Pseudomonas aeruginosa*: infections, animal modeling, and therapeutics, *Cells* 12 (2023).
- [65] A. Mazandarani, S. Goudarzi, M. Jafarabadi, E. Azimi Nekoo, Effects of cold plasma on *Staphylococcus aureus*, *J. Family Reprod Health* 16 (2022) 212–216.
- [66] D. Yan, A. Malyavko, Q. Wang, K.K. Ostrikov, J.H. Sherman, M. Keidar, Multimodal biological destruction by cold atmospheric plasma, *Capability and Mechanism. Biomedicines* 9 (2021).
- [67] L. Han, S. Patil, D. Boehm, V. Milosavljević, P.J. Cullen, P. Bourke, Mechanisms of inactivation by high-voltage atmospheric cold plasma differ for *Escherichia coli* and *Staphylococcus aureus*, *Appl. Environ. Microbiol.* 82 (2016) 450–458.
- [68] W.D. Bradford, L. Cahoon, S.R. Freel, L.L.M. Hoopes, T.T. Eckdahl, An inexpensive gel electrophoresis-based polymerase chain reaction method for quantifying mRNA levels, *Cell Biol. Education* 4 (2) (2005) 157–168.
- [69] W. Song, B. Wemheuer, P.D. Steinberg, E.M. Marzinielli, T. Thomas, Contribution of horizontal gene transfer to the functionality of microbial biofilm on a macroalgae, *ISME J.* 15 (2021) 807–817.
- [70] M. Al-Bayati, S. Samarasinghe, Biofilm and gene expression characteristics of the Carbapenem-resistant *Enterobacteriales*, *Escherichia coli* IMP, and *Klebsiella pneumoniae* NDM-1 associated with common bacterial infections, *Int. J. Environ. Res. Public Health* 19 (2022).
- [71] B.W.N. Grehs, M.A.O. Linton, B. Clasen, A. De Oliveira Silveira, E. Carissimi, Antibiotic resistance in wastewater treatment plants: understanding the problem and future perspectives, *Arch. Microbiol.* 203 (2021) 1009–1020.
- [72] L. Chen, Y. Xu, X. Dong, C. Shen, Removal of intracellular and extracellular antibiotic resistance genes in municipal wastewater effluent by electrocoagulation, *Environ. Eng. Sci.* 37 (2020) 783–789.
- [73] D. Mao, S. Yu, M. Rysz, Y. Luo, F. Yang, F. Li, J. Hou, Q. Mu, P.J. Alvarez, Prevalence and proliferation of antibiotic resistance genes in two municipal wastewater treatment plants, *Water Res.* 85 (2015) 458–466.
- [74] T. Jäger, N. Hembach, C. Elpers, A. Wieland, J. Alexander, C. Hiller, G. Krauter, T. Schwartz, Reduction of antibiotic resistant Bacteria during conventional and advanced wastewater treatment, and the disseminated loads released to the environment, *Front. Microbiol.* 9 (2018) 2599.
- [75] E.T. Anthony, M.O. Ojemaye, O.O. Okoh, A.I. Okoh, A critical review on the occurrence of resistomes in the environment and their removal from wastewater using apposite treatment technologies: limitations, successes and future improvement, *Environ. Pollut.* 263 (2020) 113791.
- [76] T. Luukkonen, J. Teeriniemi, H. Prokkola, J. Rämö, U. Lassi, Chemical aspects of peracetic acid based wastewater disinfection, *Water SA* 40 (2014) 73–80.
- [77] J. Gomes, A. Matos, M. Gmurek, R.M. Quinta-Ferreira, R.C. Martins, Ozone and photocatalytic processes for pathogens removal from water: A review, *Catalysts* 9 (2019).
- [78] K.-I. Naicker, P. Kaweesa, M.O. Daramola, S.A. Iwarere, Non-thermal plasma review: assessment and improvement of feasibility as a retrofitted Technology in Tertiary Wastewater Purification, *Appl. Sci.* 13 (2023) 6243.